



Empirical food webs of 12 tropical reservoirs in Singapore

Clare Wilkinson[‡], Rayson B H Lim[‡], Jia Huan Liew[§], Jeffrey T B Kwik[‡], Claudia L Y Tan[‡], Tan Heok Hui[‡], Darren C J Yeo^{‡,¶}

[‡] National University of Singapore, Singapore, Singapore

[§] Lingnan University, Hong Kong, China

[¶] Lee Kong Chian Natural History Museum, National University of Singapore, Kent Ridge, Singapore

Corresponding author: Clare Wilkinson (cwilkinson024@gmail.com), Rayson B H Lim (dbslbr@nus.edu.sg), Jia Huan Liew (jhliew@ln.edu.hk), Darren C J Yeo (dbsyeod@nus.edu.sg)

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Abstract

Background

Food webs summarise trophic interactions of the biotic components within an ecosystem, which can influence nutrient dynamics and energy flows, ultimately affecting ecosystem functions and services. Food webs represent the hypothesised trophic links between predators and prey and can be presented as empirical food webs, in which the relative strength/importance of the respective links are quantified. Some common methods used in food web research include gut content analysis (GCA) and stable isotope analysis (SIA). We combine both methods to construct empirical food web models as a basis for monitoring and studying ecosystem-level outcomes of natural (e.g. species turnover in fish assemblage) and intentional environmental change (e.g. biomanipulation).

New information

We present 12 food webs from tropical reservoir communities in Singapore and summarise the topology of each with widely-used network indices (e.g. connectance, link density). Each reservoir was surveyed over 4–6 sampling occasions, during which, representative animal groups (i.e. fish species and taxonomic/functional groups of zooplankton and benthic macroinvertebrates) and all likely sources of primary production (i.e. macrophytes, periphyton, phytoplankton and riparian terrestrial plants) were collected. We analysed gut content in fishes and bulk isotope ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) profiles of all animals (i.e. fishes and invertebrates) and plants collected. Both sets of information were used to estimate the relative strength of trophic relationships using Bayesian mixing models. We document our protocol here, alongside a script in the R programming language for executing data management/analyses/visualisation procedures used in our study. These data can be used to glean insights into trends in inter- and intra-specific or guild interactions in analogous freshwater lake habitats.

Keywords

gut content, stable isotope analysis, freshwater communities, reservoirs, trophic interactions

Introduction

Food webs depict feeding interactions in an ecosystem; they are indicative of energy flow between species and/or communities through ecosystems (Garvey and Whiles 2016, Barnes et al. 2018) and are shaped by evolutionary, ecological and neutral processes (Akin and Winemiller 2006). Food webs can, therefore, be used to study a range of topics, ranging from: predator-prey and competitive relationships, to the web network including food chain length, connective and potential regulation via top-down or bottom-up control (Winemiller and Polis 1996).

Gut content analysis (GCA), theoretical models based on data from literature and allometric scaling were common methods used in earlier food web studies (e.g. Cohen 1978, Pimm 1982), but technological advances have enabled the construction of more precise, empirical food webs. More contemporary or newer methods include bulk stable isotope analysis (SIA) and compound specific isotope analysis (CSIA) of specific molecules (e.g. amino acids or fatty acids; Arrington et al. 2006, Grey 2006, Chikaraishi et al. 2014, Liew et al. 2019), of which, bulk SIA is one of the most widely used, given its accessibility and relative affordability. These approaches provide a quantitative measure of resource assimilation (as opposed to ingestion measured by gut contents analysis) and allow ecologists to assess the relative strength of trophic pathways in a food web (Michener and Lajtha 2007). Food web construction informed by SIA is based on the differential ratios of naturally-occurring carbon-13 and nitrogen-15 isotopes, expressed as $\delta^{13}\text{C}$ (‰) and $\delta^{15}\text{N}$ (‰), respectively. The former is indicative of contributions from different

carbon (i.e. food) sources, while the latter correlates with trophic position (Hopkins and Ferguson 2012, Parnell 2020), thereby yielding insights into resource-consumer links and roles within the community.

As part of an in-depth study of 12 reservoirs in the tropical island nation of Singapore, we combined two complementary methods: GCA and SIA, to elucidate the empirical food web structure of the freshwater reservoir communities. These systems constitute novel ecosystems (Hobbs et al. 2009) which are typically dominated by a mix of disturbance-tolerant native and non-native species (Yeo and Lim 2011, Liew et al. 2016, Tan et al. 2020) and this represents the first study to characterise and quantify the trophic relationships of the reservoir communities in Singapore.

General description

Purpose: The aim of this paper is to provide the most comprehensive dataset to date, describing trophic interactions between key plant and animal groups in 12 of Singapore's 17 reservoirs. The data were constructed using a standardised procedure utilising gut content and stable isotope data, which we detail in the following sections. This facilitates unbiased comparisons of food web indices (e.g. fraction of basal producers, intermediate, top predators; mean Pianka's overlap; niche breadth), allowing researchers to address questions pertaining to natural (e.g. environmental) and artificial (e.g. urbanisation) drivers of food web trends across spatial or temporal gradients.

Project description

Title: Project 1 (Dec 2014–May 2016): Biodiversity and biological interactions in six Singapore reservoirs — a pilot study of food web and trophic structure and implications for environmental and water quality management.

Project 2 (June 2016–Oct 2018): Biodiversity and biological interactions in Singapore's reservoirs and waterways — a study of food web and trophic structure and implications for environmental and water quality management.

Personnel: Darren C J Yeo, Heok Hui Tan, Timothy Jardine, Jeffrey T B Kwik, Rudolf Meier, Jia Huan Liew, Clare Wilkinson, Rayson B H Lim, Claudia L Y Tan, Ming Li Chen, Wen Qing Ng, Yvonne Y W Kwang, Abel C Y Saw and Shan Shan Liu

Study area description: The data presented in this paper were collected from 12 man-made reservoirs in the tropical island of Singapore (1°21.0' N, 103°49.11' E; Fig. 1, Table 1), six being sampled in Project 1 and six sampled in Project 2. The reservoirs are located within catchments with land-use regimes ranging from protected nature reserves (i.e. riparian vegetation consisting of secondary and primary forests) to urban centres (i.e. riparian zone consisting of rocky rip-rap and impervious surfaces with minimal herbaceous plant cover). The reservoirs vary in age (i.e. time since impoundment), ranging from eight to more than 100 years (Yeo and Lim 2011). The reservoirs are typically characterised by

warm (mean water surface temperature of 28.9–31.0°C), slightly turbid and basic (mean pH from 7.1–8.4) waters that are dominated by non-native fish assemblages (Kwik et al. 2020, Tan et al. 2020).

Table 1.

Additional information on the 12 reservoirs in Singapore, including the number of transects sampled, sampling period, the type of reservoir (estuarine – reservoir has a tidal gate, but water is not saline; forest – predominantly surrounded by forest in the riparian zone; urban - predominantly developed riparian zone), mean pH and salinity and the year the reservoir construction was completed.

	Latitude / Longitude	# of transects	Sampling period	Type	Mean pH	Mean salinity (ppt)	Year constructed
Res 1	1°20.5'N, 103°55.5'E	6	Sep–Oct 2015	Urban	8.24	0.101	1984
Res 2	1°17.2'N, 103°52.0'E	6	Apr–May 2016	Estuarine	7.7	0.169	2008
Res 3	1°18.9'N, 103°44.6'E	6	Jun–Jul 2015	Urban	8.27	0.141	1974
Res 4	1°24.2'N, 103°53.2'E	6	Dec 2014–Feb 2015	Estuarine	-	-	2006
Res 5	1°23.4'N, 103°55.0'E	6	Mar–Apr 2015	Estuarine	8.4	0.168	2011
Res 6	1°22.1'N, 103°48.3'E	8	Dec 2015–Feb 2016	Forest	7.34	0.13	1974
Res 7	1°20.5'N, 103°43.7'E	6	Jan–Mar 2017	Urban	7.4	0.141	1971
Res 8	1°22.2'N, 103°49.4'E	6	Apr–May 2018	Forest	7.11	0.071	1910
Res 9	1°20.7'N, 103°49.3'E	6	Jul–Aug 2017	Forest	6.56	0.031	1907
Res 10	1°24.2'N, 103°48.0'E	8	Jan–Mar 2018	Forest	7.48	0.069	1969
Res 11	1°24.3'N, 103°50.6'E	6	May–Jun 2018	Estuarine	8.19	0.093	1984
Res 12	1°25.5'N, 103°44.4'E	8	Aug–Oct 2018	Estuarine	8.37	0.095	1975

Funding: This project was funded by PUB, Singapore's Water Agency [National University of Singapore grant number R-154-000-619-490 and R-154-000-A20-490].

Taxa	Res 1	Res 2	Res 3	Res 4	Res 5	Res 6	Res 7	Res 8	Res 9	Res 10	Res 11	Res 12
<i>Parambassis siamensis</i>			1				1		1			1
<i>Potamotrygon motoro</i>								1		1	1	
<i>Pterygoplichthys disjunctivus</i>		1			1							
<i>Pterygoplichthys pardalis</i>		1										1
<i>Pterygoplichthys</i> spp.							1					
<i>Rasbora borapertensis</i>											1	
<i>Rhinogobius similis</i>	1		1	1	1							
<i>Scleropages formosus</i>						1		1	1	1		
<i>Vieja melanura</i>		1	1		1		1					

Sampling description: Fishes were collected using a combination of cast netting (net dimensions: 4 m radius, 2 cm mesh size), trapping (trap dimensions: 50 cm × 60 cm × 40 cm, 2 cm mesh size) and boat electrofishing (pulsed DC electrofishing, model ETS-MBS-1D-COL) in the littoral zone of each reservoir to optimise sampling coverage across various depths and fish sizes. We performed 10 casts, deployed three traps (for 48 hours) and conducted four 5-minute bursts of electrofishing, per transect. Live specimens were identified to species level (Baker and Lim 2008, Ng and Tan 2010), measured for standard length (cm), weighed for total wet weight (g) using PESOLA weighing scales and up to three individuals per species were euthanised for GCA and SIA. Invertebrate samples were collected using multiple methods. Specifically, benthic invertebrates were collected using two submerged colonisers filled with exogenous substrate (i.e. coconut palm fronds) which were deployed at a fixed depth of 1–1.5 m in the littoral zone for a total of 8–12 weeks at each transect (Loke et al. 2010). Additionally, a hand-held dip net (500 µm mesh size) was used to collect invertebrates from the surface and plankton nets (100 µm mesh size) used to collect zooplankton. All fixed (in 70% ethanol) invertebrate specimens were identified to the highest taxonomic resolution possible, based on identification keys to the freshwater macroinvertebrates of the Malaysian region and of Singapore (Yule and Yong 2004, Blakely et al. 2010). Phytoplankton were also collected by towing a 80 µm and 50 µm mesh size plankton net at the surface for ≈ 100 m per transect. Zooplankton and phytoplankton samples were then separated manually using pipettes or by successive filtration using cell strainers (100 µm, 70 µm and 45 µm mesh size) and the purity and identity of the samples evaluated under a dissecting microscope. Other basal resources were collected: periphyton and epiphytic algal samples were sampled from hard surfaces in the littoral zones; and macrophytes and riparian plants and/or grass samples comprised leaf clippings.

Quality control: To facilitate representative sampling, each reservoir was systematically subdivided into six or eight transects (dependent on reservoir size; ≈ 200 m per transect) spanning multiple littoral habitat types (i.e. rocky bund, forested, macrophyte-dominated). Sampling effort and protocols for the various taxonomic groups were standardised to

facilitate comparability between reservoirs as described above. In addition, fish species identity was validated by taxon experts from the Lee Kong Chian Natural History Museum.

Step description: We followed general protocols from Liew et al. (2018) to combine GCA and SIA data to construct empirical food webs for all 12 reservoirs. We used GCA to provide information for the selection of potential prey, before estimating the relative strength of trophic relationships using Bayesian Mixing Models.

I. Gut content analysis

The gut content of 47 fish species with at least four individuals having full gut were analysed to complement the stable isotope analysis (Kwik et al. 2020). In total, 1269 suitable guts were included for the dietary analysis. Dietary items for each fish species were identified under a dissecting microscope in a Petri dish and grouped into ten broad categories: 1) substrate (i.e. inorganic sediment), 2) unidentifiable animal matter (i.e. highly-digested matter of non-fish vertebrates), 3) plant matter (e.g. whole or parts of leaves, fruit and flowers), 4) periphyton (i.e. benthic and filamentous algae mats), 5) phytoplankton (i.e. pelagic algae, e.g. *Microcystis*, *Planktolyngbya* and *Staurastrum*), 6) zooplankton (e.g. Rotifera, Copepoda), 7) insect larvae (i.e. whole or parts of aquatic insect larvae, e.g. Chironomidae, Caenidae), 8) gastropods (i.e. aquatic snails and bivalves, e.g. Viviparidae, Ampullariidae), 9) decapod crustaceans (i.e. crabs and shrimp, e.g. Palaemonidae) and 10) fishes (i.e. whole body, scales or bones). The relative contributions of various dietary items for each fish species were determined, based on their frequency of occurrence ($FO\%$) and volumetric contribution ($VO\%$) to calculate the feeding index (Sá-Oliveira et al. 2014):

$$FI_i = \frac{(FO_i \times VO_i)}{\sum(FO_i \times VO_i)}$$

where FI_i = feeding index of species i , FO_i = frequency of occurrence of diet item i and VO_i = volume of diet item i .

II. Stable isotope analysis

We estimated the relative strength of trophic interactions of consumers and resources in the food webs using bulk $^{13}\text{C}/^{12}\text{C}$ (i.e. $\delta^{13}\text{C}$) and $^{15}\text{N}/^{14}\text{N}$ (i.e. $\delta^{15}\text{N}$) isotope profiles. Tissue samples for primary producers (e.g. riparian plants and macrophytes, phytoplankton, periphyton) consisted of leaf clippings from plants, filtered phytoplankton samples and substrate-free periphyton samples. Invertebrate samples comprised whole organisms for smaller specimens (e.g. dipterans) and muscle tissue for the larger taxa (e.g. gastropods). For fish samples, we extracted muscle tissue from the dorsal region of each individual. We collected a minimum of three samples per taxon and excluded rare species (with less than three individuals collected) from subsequent analyses. However, we made an exception for taxa that were abundant in the study site, but were difficult to isolate for SIA (e.g. Copepoda).

All samples were oven-dried for 48–72 h at 68.5–70.0°C, homogenised, ground to a fine powder and weighed (to the nearest 0.0001 g): 1 mg for consumers (e.g. fish, invertebrates) or 4–5 mg for primary producers (e.g. phytoplankton and plants) following protocols described in Jardine et al. (2003). In total, 5339 tissue samples were prepared and analysed. These consisted of 1042 primary producers, 1476 invertebrates and 2821 fish samples, respectively. These were packed in standard tin capsules and sent to the Stable Isotope Facility at the University of California, Davis for determination of stable isotopic profiles ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$).

III. Construction of empirical food webs

We summarised the trophic information derived from GCA and SIA into predation matrices ($n = 12$). The GCA information was used to inform and/or restrict the pool of potential prey included in the models for each fish species, while information from published literature were used to identify potential resources for the invertebrate taxa. We used Bayesian stable isotope mixing models to estimate the proportional source contribution to diets of consumers by fitting probability models to the isotopic data (e.g. isotopic ratios, elemental concentrations, sample variations and trophic fractionation).

Before running the mixing models, we corrected the $\delta^{13}\text{C}$ isotopic profiles of samples that comprised whole individuals (e.g. small invertebrates) for lipid-enrichment using procedures described in Logan et al. (2008). We also adjusted the stable nitrogen and carbon isotope data to account for trophic discrimination, using taxon-specific values for $\delta^{15}\text{N}$ (following Bunn et al. 2013) and generic values for $\delta^{13}\text{C}$ ($0.4 \pm 1.3\text{‰}$; Post 2002). We note that fractionation values reported in Bunn et al. (2013) were derived from lotic ecosystems, while our study systems are lentic. Nevertheless, the range of fractionation values fall within commonly reported fractionation ranges of 2–4‰, so we did not expect our findings to deviate significantly if we used lentic-specific (but not taxon-specific) values instead. Moreover, strong evidence of inter-taxa differences in $\delta^{15}\text{N}$ fractionation (Sweeting et al. 2007, Bunn et al. 2013) suggest the importance of accounting for taxon-specific variations in isotopic fractionation, especially given the broad taxonomic coverage of our work. We ran 30,000 iterations of each mixing model (i.e. each consumer) with a burn-in of 10,000 draws using Markov Chain Monte Carlo (MCMC) processes on the *simmr* statistical package, version 0.3 (Parnell 2020). Model convergence was assessed using Gelman-Rubin diagnostic parameter (≈ 1).

We assessed the feasibility of all our mixing models prior to extracting finalised source contribution values using two criteria. First, we ensured that consumer isotopic profiles were bounded within mixing polygons (evaluated by an isopleth produced by the model; Phillips et al. 2014). The mixing polygons were also used to assess isotopic overlap between taxa in a food web. If there was significant isotopic overlap (ANOVA/t-test; p -value < 0.05) between phenotypically similar taxa, taxa were pooled to a higher taxonomic level, for example, genus for fish and to the order or class level for invertebrates, to create isotopically distinct taxonomic units. Second, we ensured that sources (i.e. prey) warranted inclusion in mixing models by removing items which fall below a minimum threshold. We used the following equation to calculate consumer specific threshold values: $(1/N) \times 0.25$,

where N is number of sources for a consumer. If multiple food sources fall below the threshold value, we deemed them as 'unimportant' sources and removed them sequentially (starting from the lowest contributor), while ensuring that the removal of each source does not result in consumer profiles falling outside mixing polygons. We also checked for missing consumer sources by assessing respective mixing polygons, finding that a primary source was missing in three reservoirs where phytoplankton were too scarce to meet the required weight for SIA. For these three reservoirs, we created an 'unknown producer' source node using the common practice of adapting isotopic signatures of zooplankton, a likely predator of this missing resource (Grey et al. 2000, Post 2002, Matthews and Mazumder 2003). Finally, we summarised source-contribution for each consumer using the median value of posterior distributions into a predation matrix (column = predator; row = prey).

We provide the R script used for producing the empirical food webs, as well as the raw data from one reservoir (res 6) to facilitate replication of our procedure (Suppl. materials 1, 2) and produce a food web (Fig. 2).

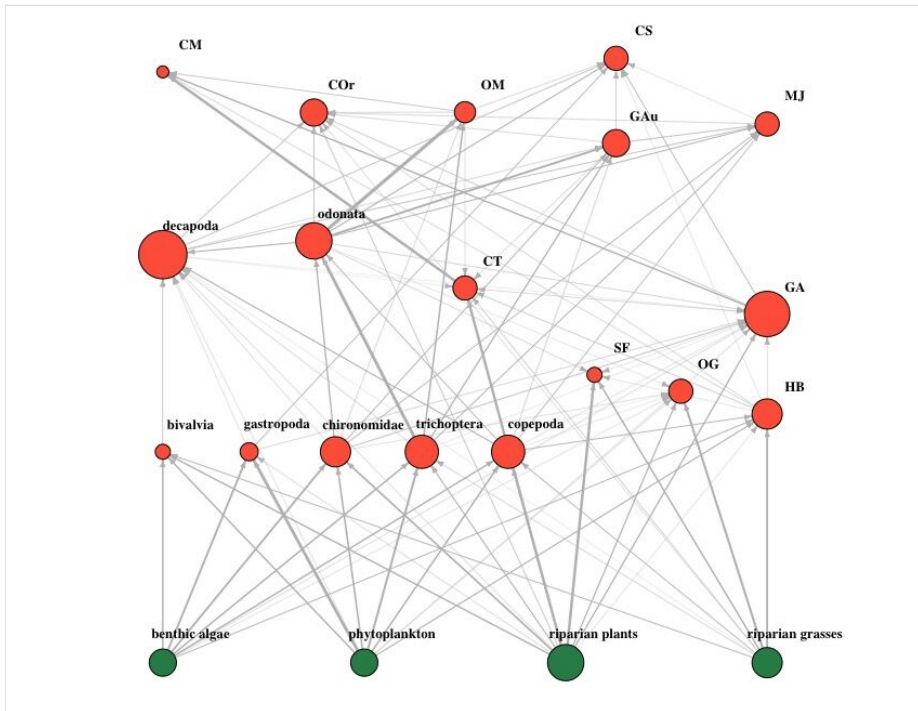


Figure 2. [doi](#)

Food web diagram for Reservoir 6. Created from stable isotope data, using the code in Suppl. material 1 and data in Suppl. material 2. The size of each node representing individual taxa is indicative of the number of links formed with other taxa (larger nodes indicate more links), while the thickness of arrows representing predator-prey relationship indicates the relative interaction strengths (thicker arrows indicate greater interaction).

Geographic coverage

Description: N.A. (already included in “Study area description”)

Taxonomic coverage

Description: Table 2

Traits coverage

N.A.

Temporal coverage

Notes: December 2014 to October 2018

Usage licence

Usage licence: Creative Commons Public Domain Waiver (CC-Zero)

Data resources

Data package title: Diet composition of the fish species and food web matrix for 12 shallow tropical reservoirs in Singapore.

Resource link: <https://doi.org/10.5061/dryad.jsxksn088>

Number of data sets: 4

Data set name: diet composition_fish.xlsx

Data format: Excel spreadsheet

Description: This dataset contains information on the diet information fish species with more than four replicates recorded from 12 reservoirs in Singapore.

Column label	Column description
Reservoir	Identifier for the reservoir.
Latitude	The value, in degrees minutes, of the water body's position north of the equator as determined from Google Earth.
Longitude	The angular distance, in degrees minutes, of the water body's position east of the meridian at Greenwich, England, as determined from Google Earth.

Sampling period	Period during which the fishes were collected from our surveys using electrofishing, traps and cast nets.
Fish species	Scientific name of various fish species collected for gut content analysis.
Replicates	Number of full guts examined.
substrate	The mean proportion of the total gut content volume accounted for by small rocks and sand.
unidentified.animal.matter	The mean proportion of the total gut content volume accounted for by unidentifiable prey items.
plant	The mean proportion of the total gut content volume accounted for by plant materials including leaf fragments, seeds, fruits and woody debris.
periphyton	The mean proportion of the total gut content volume accounted for by benthic and filamentous algae mats.
phytoplankton	The mean proportion of the total gut content volume accounted for by pelagic algae.
zooplankton	The mean proportion of the total gut content volume accounted for by zooplankton including Rotifers as well as from orders Cladocera, Cyclopoida, Harpacticoida and Calanoida.
insect	The mean proportion of the total gut content volume accounted for by terrestrial or aquatic insects.
decapod	The mean proportion of the total gut content volume accounted for by freshwater shrimps, crabs and crayfish.
mollusc	The mean proportion of the total gut content volume accounted for by gastropods and bivalves including shells and operculum.
fish	The mean proportion of the total gut content volume accounted for by fishes, including bones and scales.
detritus	The mean proportion of the total gut content volume accounted for by fine and coarse particulate organic matter.
hirudinea	The mean proportion of the total gut content volume accounted for by leeches.

Data set name: Bottom-up matrices_FW_v2.xlsx

Data format: Excel spreadsheet

Description: This dataset contains information on the food web topology (feeding links) and interaction strength (proportional contribution of diets to consumers) summarised in a predator (consumer)-prey (resources) matrix for 12 reservoirs in Singapore (Wilkinson et al. 2021). Each row in the matrix represents a resource/prey, while each column represents a consumer/predator. For each matrix element that is > 0 , it expresses the relative contribution of each resource to the diet of its consumer.

Column label	Column description
benthic algae	Primary producer: refers to filamentous algae.
emergent macrophytes	Primary producer: for example, <i>Ludwigia ascendens</i> .
floating macrophytes	Primary producer: includes vegetation within floating wetlands, for example, cattail (genus: <i>Typha</i>) plants.
macrophytes	Primary producer: includes plants in the genera: <i>Hydrilla</i> , <i>Mayaca</i> and <i>Valisineria</i> .
periphyton	Primary producer: refers to encrusted algae/periphyton, scraped from rocks.
phytoplankton	Primary producer.
riparian grasses	Primary producer: includes all C4 plants that are typically grasses in the riparian zone.
riparian plants	Primary producer: includes all C3 plants in the riparian zone.
unknown producer	Primary producer: We were unable to collect phytoplankton (or other pelagic/planktonic producers from three reservoirs, so a node was created to simulate this. The values of this node were subsequently based on the zooplankton nodes in each of the three reservoirs (Grey et al. 2000, Post 2002, Matthews and Mazumder 2003).
Ampullariidae	Invertebrate family, predominantly <i>Pomacea</i> .
Bivalvia	Invertebrate class.
Conchostraca	Invertebrate suborder.
Decapoda	Invertebrate order.
Chironomidae	Invertebrate family.
Ephemeroptera	Invertebrate order.
Gastropoda	Invertebrate class.
Hemiptera	Invertebrate order.
Hirudinea	Invertebrate subclass.
Nassariidae	Invertebrate family, assassin snails.
Odonata	Invertebrate order.
Oligochaeta	Invertebrate subclass.
Ostracoda	Invertebrate class.
Trichoptera	Invertebrate order.
Copepoda	Invertebrate subclass.
AA	Fish species: <i>Aplocheilichthys armatus</i> .
AC	Fish species: <i>Amphiphilichthys citrinellus</i> .
AH	Fish species: <i>Acarichthys heckelii</i> .

AS	Fish species: <i>Atractosteus spatula</i> .
BS	Fish species: <i>Barbonymus schwanefeldii</i> .
CA	Fish species: <i>Cyclocheilichthys apogon</i> .
CC	Fish species: <i>Cyprinus carpio</i> .
CG	Fish species: <i>Clarias gariepinus</i> .
Cichla	Fish genus: <i>Cichla</i> spp.
CL	Fish species: <i>Channa lucius</i> .
CM	Fish species: <i>Channa micropeltes</i> .
CO	Fish species: <i>Cichla orinocensis</i> .
COr	Fish species: <i>Chitala ornata</i> .
CS	Fish species: <i>Channa striata</i> .
CT	Fish species: <i>Cichla temensis</i> .
MU	Fish species: <i>Mayaheros urophthalmus</i> .
DC	Fish species: <i>Dermogenys collettei</i> .
DM	Fish species: <i>Datnioides microlepis</i> .
ES	Fish species: <i>Etroplus suratensis</i> .
GA	Fish species: <i>Geophagus altifrons</i> .
Gam	Fish species: <i>Gambusia affinis</i> .
GAu	Fish species: <i>Glossogobius aureus</i> .
Goby	Fish family: Gobiidae.
HB	Fish species: <i>Heterotilapia buttkoferi</i> .
HQ	Fish species: <i>Hyporhamphus quoyi</i> .
HR	Fish species: <i>Hemigrammus rodwayi</i> .
LR	Fish species: <i>Leptobarbus rubripinna</i> .
MC	Fish species: <i>Megalops cyprinoides</i> .
MJ	Fish species: <i>Monopterus javanensis</i> .
MZ	Fish species: <i>Macrognathus zebrinus</i> .
NN	Fish species: <i>Notopterus notopterus</i> .
OG	Fish species: <i>Osphronemus goramy</i> .
OH	Fish hybrid genus: <i>Oreochromis</i> spp. (hybrid).
OJ	Fish species: <i>Oryzias javanicus</i> .

OM	Fish species: <i>Oxyeleotris marmorata</i> .
OMo	Fish species: <i>Oreochromis mossambicus</i> .
ON	Fish species: <i>Oreochromis niloticus</i> .
OV	Fish species: <i>Osteochilus vittatus</i> .
PD	Fish species: <i>Pterygoplichthys disjunctivus</i> .
PMa	Fish species: <i>Parachromis managuense</i> .
PMot	Fish species: <i>Potamotrygon motoro</i> .
PP	Fish species: <i>Pterygoplichthys pardalis</i> .
PS	Fish species: <i>Parambassis siamensis</i> .
Pter	Fish genus: <i>Pterygoplichthys</i> spp.
RB	Fish species: <i>Rasbora boraptensis</i> .
RS	Fish species: <i>Rhinogobius similis</i> .
SF	Fish species: <i>Scleropages formosus</i> .
VM	Fish species: <i>Vieja melanura</i> .

Data set name: R_script.R

Data format: R script

Description: This file contains the R code to create the predation matrix and food web for a sample reservoir (Upper Peirce; Res 6). We used R version 3.5.2 to develop the code. Within R, the following packages were used and are necessary to create the predation matrices and food webs: simmr version 0.4.2 (Parnell 2020); NetIndices; ggplot2; and igraph.

Column label	Column description
R code	R code.

Data set name: MixingModelInputs_Res6.zip

Data format: Zip file of excel spreadsheets

Description: This folder contains both the source data (feeding links) and stable isotope ratios for both carbon and nitrogen isotopes for each taxa within the food web for a sample reservoir (Upper Peirce; Res 6). For each taxa, there will be two csv files, the first TaxaX.csv (listed by the taxa identifiers described above) with the SIA data for that taxa and the second TaxaX.Sources.csv with all the source data for that taxa.

Column label	Column description
Sheet 1: TaxaX.csv D13C	The $d^{13}\text{C}$ ratio for each individual of the specified taxa.
Sheet 1: TaxaX.csv D15N	The $d^{15}\text{N}$ ratio for each individual of the specified taxa.
Sheet 2: TaxaX.Sources.csv Sources	Taxa identifier of the prey source.
Sheet 2: TaxaX.Sources.csv Meand13C	Mean $d^{13}\text{C}$ ratio for the prey source.
Sheet 2: TaxaX.Sources.csv SDd13C	Standard deviation $d^{13}\text{C}$ ratio for the prey source.
Sheet 2: TaxaX.Sources.csv Meand15N	Mean $d^{15}\text{N}$ ratio for the prey source.
Sheet 2: TaxaX.Sources.csv SDd15N	Standard deviation $d^{15}\text{N}$ ratio for the prey source.
Sheet 2: TaxaX.Sources.csv tefd13C	^{13}C trophic enrichment factor for the prey source.
Sheet 2: TaxaX.Sources.csv tefSDd13C	Standard deviation of ^{13}C trophic enrichment factor for the prey source.
Sheet 2: TaxaX.Sources.csv tefd15N	^{15}N trophic enrichment factor for the prey source.
Sheet 2: TaxaX.Sources.csv tefSDd15N	Standard deviation of ^{15}N trophic enrichment factor for the prey source.
Sheet 2: TaxaX.Sources.csv conc13C	The weight of carbon recorded in the sample.
Sheet 2: TaxaX.Sources.csv conc15N	The weight of nitrogen recorded in the sample.

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Author contributions

All authors were involved in the collection of data. RBHL, JHL and CLYT sorted, processed and analysed the data. CLW and RBHL wrote the manuscript. All authors reviewed and edited the manuscript.

References

- Akin S, Winemiller KO (2006) Seasonal variation in food web composition and structure in a temperate tidal estuary. *Estuaries and Coasts* 29 (4): 552-567. <https://doi.org/10.1007/bf02784282>
- Arrington DA, Davidson BK, Winemiller KO, Layman CA (2006) Influence of life history and seasonal hydrology on lipid storage in three neotropical fish species. *Journal of Fish Biology* 68: 1347-1361. <https://doi.org/10.1111/j.0022-1112.2006.00996.x>
- Baker N, Lim KKP (2008) Wild animals of Singapore: a photographic guide to mammals, reptiles, amphibians and freshwater fishes. Draco Publishing, 180 pp. [ISBN 978-9810594596]
- Barnes AD, Jochum M, Lefcheck JS, Eisenhauer N, Scherber C, O'Connor MI, de Ruiter P, Brose U (2018) Energy flux: The link between multitrophic biodiversity and ecosystem functioning. *Trends in Ecology & Evolution* 33 (3): 186-197. <https://doi.org/10.1016/j.tree.2017.12.007>
- Blakely TJ, Harding JS, Clews E, Winterbourn MJ (2010) An illustrated guide to the freshwater macroinvertebrates of Singapore. School of Biological Sciences, University of Canterbury, Christchurch, 74 pp. [ISBN 978-0-473-16730-1]
- Bunn SE, Leigh C, Jardine TD (2013) Diet-tissue fractionation of $\delta^{15}\text{N}$ by consumers from streams and rivers. *Limnology and Oceanography* 58 (3): 765-773. <https://doi.org/10.4319/lo.2013.58.3.0765>
- Chikaraishi Y, Steffan SA, Ogawa NO, Ishikawa NF, Sasaki Y, Tsuchiya M, Ohkouchi N (2014) High-resolution food webs based on nitrogen isotopic composition of amino acids. *Ecology and Evolution* 4 (2423): 2423-2449. <https://doi.org/10.1002/ece3.1103>
- Cohen JE (1978) Food Webs and Niche Space. 11. Princeton University Press <https://doi.org/10.2307/j.ctvx5wc04>
- Garvey JE, Whiles MR (2016) Trophic ecology. CRC Press, Boca Raton, Florida, USA. <https://doi.org/10.1201/9781315367804>
- Grey J, Jones RI, Sleep D (2000) Stable isotope analysis of the origins of zooplankton carbon in lakes of differing trophic state. *Oecologia* 123 (2): 232-240. <https://doi.org/10.1007/s004420051010>
- Grey J (2006) The use of stable isotope analyses in freshwater ecology: Current awareness. *Polish Journal of Ecology* 54 (563): 563-584.
- Hobbs RJ, Higgs E, Harris JA (2009) Novel ecosystems: implications for conservation and restoration. *Trends in Ecology & Evolution* 24 (11): 599-605. <https://doi.org/10.1016/j.tree.2009.05.012>
- Hopkins JB, Ferguson JM (2012) Correction: Estimating the diets of animals using stable isotopes and a comprehensive Bayesian mixing model. *PLOS One* 7 (1). <https://doi.org/10.1371/annotation/d222580b-4f36-4403-bb1f-cfd449a5ed74>
- Jardine TD, McGeachy SA, Paton CM, Savoie M, Cunjak RA (2003) Stable isotopes in aquatic systems: Sample preparation, analysis, and interpretation. Canadian Manuscript Report of Fisheries and Aquatic Science. No. 2656: 39 p..
- Kwik JTB, Lim RBH, Liew JH, Yeo DCJ (2020) Novel cichlid-dominated fish assemblages in tropical urban reservoirs. *Aquatic Ecosystem Health & Management* 23 (3): 249-266. <https://doi.org/10.1080/14634988.2020.1778308>

- Liew JH, Tan HH, Yeo DCJ (2016) Dammed rivers: impoundments facilitate fish invasions. *Freshwater Biology* 61 (9): 1421-1429. <https://doi.org/10.1111/fwb.12781>
- Liew JH, Jardine TD, Lim RBH, Kwik JTB, Tan HH, Kho ZY, Yeo DCJ (2018) Bottom-up influences on tropical freshwater food web structure support the “environmental filtering” hypothesis. *Limnology and Oceanography* 63 (5): 1877-1890. <https://doi.org/10.1002/lno.10813>
- Liew JH, Chua KWJ, Arsenault ER, Thorp JH, Suvarnaraksha A, Amirrudin A, Yeo DCJ (2019) Quantifying terrestrial carbon in freshwater food webs using amino acid isotope analysis: Case study with an endemic cavefish. *Methods in Ecology and Evolution* 10 (9): 1594-1605. <https://doi.org/10.1111/2041-210x.13230>
- Logan J, Jardine T, Miller T, Bunn S, Cunjak R, Lutcavage M (2008) Lipid corrections in carbon and nitrogen stable isotope analyses: comparison of chemical extraction and modelling methods. *Journal of Animal Ecology* 77 (4): 838-846. <https://doi.org/10.1111/j.1365-2656.2008.01394.x>
- Loke LHL, Clews E, Low E, Belle CC, Todd PA, Eikaas HS, Ng PKL (2010) Methods for sampling benthic macroinvertebrates in tropical lentic systems. *Aquatic Biology* 10 (2): 119-130. <https://doi.org/10.3354/ab00274>
- Matthews B, Mazumder A (2003) Compositional and interlake variability of zooplankton affect baseline stable isotope signatures. *Limnology and Oceanography* 48 (5): 1977-1987. <https://doi.org/10.4319/lo.2003.48.5.1977>
- Michener R, Lajtha K (Eds) (2007) *Stable Isotopes in Ecology and Environmental Science*. 2nd edition. Wiley-Blackwell, 594 pp. [ISBN 9781405126809] <https://doi.org/10.1002/9780470691854>
- Ng HH, Tan HH (2010) An annotated checklist of the non-native freshwater fish species in the reservoirs of Singapore. *Cosmos* 6: 95-116. <https://doi.org/10.1142/S0219607710000504>
- Parnell A (2020) *simmr*: A stable isotope mixing model. R package. 0.4.2. URL: <https://CRAN.R-project.org/package=simmr>
- Phillips DL, Inger R, Bearhop S, Jackson AL, Moore JW, Parnell AC, Semmens BX, Ward EJ (2014) Best practices for use of stable isotope mixing models in food-web studies. *Canadian Journal of Zoology* 92 (10): 823-835. <https://doi.org/10.1139/cjz-2014-0127>
- Pimm SL (1982) *Food Webs*. Chapman & Hall, New York <https://doi.org/10.1007/978-94-009-5925-5>
- Post DM (2002) Using stable isotopes to estimate trophic position: Models, methods, and assumptions. *Ecology* 83 (3): 703-718. [https://doi.org/10.1890/0012-9658\(2002\)083\[0703:usitet\]2.0.co;2](https://doi.org/10.1890/0012-9658(2002)083[0703:usitet]2.0.co;2)
- Sá-Oliveira JC, Angelini R, Isaac-Nahum VJ (2014) Diet and niche breadth and overlap in fish communities within the area affected by an Amazonian reservoir (Amapá, Brazil). *Anais da Academia Brasileira de Ciências* 86 (1): 383-406. <https://doi.org/10.1590/0001-3765201420130053>
- Sweeting CJ, Barry J, Barnes C, Polunin NV, Jennings S (2007) Effects of body size and environment on diet-tissue $\delta^{15}\text{N}$ fractionation in fishes. *Journal of Experimental Marine Biology and Ecology* 340 (1): 1-10. <https://doi.org/10.1016/j.jembe.2006.07.023>
- Tan HH, Lim KKP, Liew JH, Low BW, Lim RBH, Kwik JTB, Yeo DCJ (2020) The non-native freshwater fishes of Singapore: An annotated compilation. *Raffles Bulletin of Zoology* 68: 150-195.

- Wilkinson C, Lim RBH, Liew JH, Kwik JTB, Yeo DCJ, Tan CLY, Tan HH (2021) Empirical food webs of 12 tropical reservoirs in Singapore. Dataset. Dryad. URL: <https://doi.org/10.5061/dryad.jsxksn088>
- Winemiller KO, Polis GA (1996) Food webs: What can they tell us about the World? Food Webs1-22. https://doi.org/10.1007/978-1-4615-7007-3_1
- Yeo DCJ, Lim KKP (2011) Freshwater ecosystems. In: Ng PKL, Corlett RT, Tan HTW (Eds) Singapore biodiversity: An encyclopedia of the natural environment and sustainable development. Editions Didier Millet in association with the Raffles Museum of Biodiversity Research, Singapore, 552 pp. [ISBN 978-981-4260-08-4].
- Yule CM, Yong HS (2004) Freshwater invertebrates of the Malaysian region. Akademi Sains Malaysia [ISBN 983-41936-0-2]

Supplementary materials

Suppl. material 1: R Script for creating a food web of reservoir 6 [doi](#)

Authors: Clare L Wilkinson, Rayson B H Lim, Jia Huan Liew, Jeffrey J T Kwik, Claudia L Y Tan, Heok Hui Tan, Darren C J Yeo

Data type: R script for reproducing work presented in the paper.

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Suppl. material 2: Food web data for reservoir 6 [doi](#)

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