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## **Map of livestock density in Central Apennines: a standardised protocol**

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## 20 Abstract

21 Effective ecosystem management requires a deep understanding of how human  
22 activities, such as livestock farming, impact ecological dynamics. Livestock farming  
23 influences vegetation structure, nutrient cycling, and wildlife behaviour, yet there are  
24 limited standardised methods for estimating livestock grazing pressure on a local  
25 scale. Here we developed a standardised protocol for mapping livestock density at  
26 cadastral sheet resolution, and we tested it in a mid-mountain area of Central  
27 Apennines, Italy. The protocol combines municipal grazing data related to seasonal  
28 high-altitude pasture with interviews and geospatial mapping to create fine-scale  
29 livestock distribution maps. We focused on different livestock species and we  
30 produced a separate map for each: cattle, sheep, goats, and horses. Our protocol  
31 addressed a critical gap in conservation research by providing a robust framework  
32 for quantifying grazing pressure. These data are crucial for understanding livestock-  
33 wildlife interactions and informing ecosystem management strategies on local  
34 territory.

35

36

## 37 Introduction

38 The sustainable management of ecosystems requires a comprehensive  
39 understanding of the different elements and processes that interact within a territory,  
40 particularly the relationship between human activities and ecological dynamics.  
41 Livestock farming, one of the main anthropogenic activities impacting terrestrial  
42 ecosystems, plays a significant role in altering nutrient cycles, leading often to  
43 biodiversity loss, habitat degradation, and soil erosion (Gordon 2018). Livestock  
44 farming significantly influences vegetation structure, primary productivity, and overall  
45 ecosystem services (Li et al. 2021). Livestock presence can have negative influence  
46 on wildlife behaviour, with cascading effects on biodiversity (Briske et al. 2011). For  
47 instance, the conversion of land for grazing reduces the availability of native  
48 vegetation, which impacts wild herbivores and small mammals by limiting their food  
49 resources and protective cover (Foley et al. 2005, Schieltz and Rubenstein 2016).  
50 Moreover, the increased overlap between livestock and wildlife due to habitat  
51 fragmentation enhances the risk of predation, competition, and pathogen  
52 transmission (Ekernas et al. 2017, Jori et al. 2021). Grazing is characterised by a  
53 variety of factors such as timing, frequency, duration, season, and intensity (Briske et  
54 al. 2011, Schieltz and Rubenstein 2016). The intensity of grazing, defined as the  
55 amount of grazing per unit of primary productivity (Bouwman et al. 2005, Haberl et  
56 al. 2007, Petz et al. 2014), is the most crucial factor influencing changes in  
57 ecosystems (Schieltz & Rubenstein, 2016). While grazing generally reduces  
58 vegetation quantity, in some cases it can improve plant quality by stimulating  
59 regrowth, benefiting certain herbivore species through a facilitation process (Fraser  
60 et al. 2014, Schieltz and Rubenstein 2016).

61 Intensively managed grasslands and arable lands for livestock feed generally  
62 support low biodiversity (Newbold et al. 2015), while extensive grazing helps  
63 maintaining landscapes diversity by preventing shrub encroachment and  
64 reforestation (Rook and Tallwin 2003). Together with traditional agriculture,  
65 extensive grazing is essential for preserving many Europe's semi-natural habitats,  
66 which have been shaped over millennia and host many threatened species (Halada  
67 et al. 2011, Malek et al. 2024b). Due to their low carrying capacities in terms of  
68 climatic, soil and terrain conditions, semi-natural habitats often suffer from  
69 overgrazing, which leads to an alteration of vegetation states (Kosmas et al. 2016,  
70 Pulido et al. 2018, Sartorello et al. 2020). At the same time, many European semi-  
71 natural habitats have declined over the past 50 years due to land abandonment,  
72 making them some of the most threatened ecosystems (Falcucci et al. 2007, IPBES  
73 2018, EEA 2020, Quaranta et al. 2020). While in fact land abandonment initially has  
74 beneficial effects on biodiversity for up to 30 years, then this positive effect decline  
75 over the years, due to forest encroachment and a consequent reduction in species  
76 richness, particularly in mountainous areas (MacDonald et al. 2000, Plieninger et al.  
77 2014, Sartorello et al. 2020).

78 In Italy, less productive and mountain areas have undergone extensive land  
79 abandonment, especially in the Alps and Apennines (Mazzoleni et al. 2004,  
80 Chauchard et al. 2007, Primi et al. 2024). In central Italian Apennines, grazing  
81 pastures and marginal meadows contribute to the preservation of open habitats.  
82 Without grazing, these areas are encroached upon by shrubs and woodlands,  
83 reducing landscape heterogeneity (Falcucci et al. 2007, Ponzetta et al. 2010). Here,  
84 the shift from open habitats to woodlands has significant ecological impacts,  
85 including a decrease in biodiversity, especially for species that rely on ecotonal and

86 transitional zones (Argenti et al. 2000, Silver et al. 2000). Among these, there are  
87 species such as roe deer, or passerines which occur on traditional farming and  
88 pastoral systems (e.g., rock sparrow (*Petronia petronia*), ortolan bunting (*Emberiza*  
89 *hortulana*), red-backed shrike (*Lanius collurio*)) (Caballero et al. 2009). Conversely,  
90 the increase in forest cover is driving the expansion of large carnivores in many  
91 areas of Europe, including the Apennines (Pereira and Navarro 2015, Cimatti et al.  
92 2021). Central Apennines is home to the relict and critically endangered Marsican  
93 bear population (*Ursus arctos marsicanus*) (Ciucci and Boitani 2008). Following land  
94 abandonment and forest recover, habitat availability has increased also for this  
95 charismatic carnivore, but the population is still under threat largely due to limited  
96 environmental connectivity leading to dispersal-related mortality and especially direct  
97 human persecution (Ciucci and Boitani 2008, Falcucci et al. 2008). In this context,  
98 understanding the distribution and intensity of anthropogenic activities in Central  
99 Apennines, including grazing, is crucial for preserving the habitat of the bear and for  
100 other species inhabit this area as well as predicting and preventing human-wildlife  
101 conflict.

102 The interaction between agricultural activities and wildlife has a long and often  
103 conflictual history in central Apennines. For instance, bear-related damages have  
104 been reported on livestock (51%), domestic poultry (18%), beehives (16%), and  
105 crops and fruit trees (15%) (Ciucci and Boitani 2008). Despite the protected area's  
106 long-standing compensation program, illegal shootings of wolves and bears have  
107 continued, suggesting that the underlying societal conflict remains unresolved  
108 (Posillico et al. 2004).

109 Despite grazing's importance for certain aspects of biodiversity conservation,  
110 uncertainties remain about its optimal management. Understanding grazing patterns  
111 is essential for mitigating biodiversity loss and guiding conservation efforts.

112 As grazing intensity is a function of livestock density (Abdalla et al. 2018), the  
113 quantification of the latter is often used to calculate grazing pressure. There are  
114 several works providing livestock density estimates at global (Gilbert et al. 2018)  
115 regional (Malek et al. 2024b, 2024a) and national levels (Kolluru et al. 2023, Liu et al.  
116 2024). However, the resolution of these datasets is not sufficient to represent grazing  
117 intensity at a scale which is useful for local management, i.e., below the level of  
118 municipality. This generates a substantial gap in the scientific literature regarding  
119 standardised methods for estimating grazing intensity in terms of livestock density at  
120 a local scale. The difficulty in collecting fine-scale livestock density data results in  
121 most studies focussing on livestock grazing presence, ignoring its intensity (e.g.,  
122 Kothmann et al. 2009, Andriuzzi and Wall 2017, Filazzola et al. 2020).

123 Here we present density maps for different categories of livestock (i.e., cattle, sheep,  
124 goats and horses) in an area of Central Apennines, derived using a standardised  
125 protocol of data collection and mapping. The protocol was applied in central Italy in a  
126 mid-mountain area adjacent to national parks. These maps may provide key  
127 information for the management of a territory (Hadjigeorgiou et al. 2005) constituting  
128 a valuable tool for the in-depth study of the relationships between livestock, habitats  
129 and wildlife.

## 130 Methods

131 Our study was conducted on a 218.75 km<sup>2</sup> area which correspond to two ecological  
132 corridors identified to enhance movements of the Marsican brown bear (*Ursus arctos*  
133 *marsicanus*) in Central Apennines, Italy (Ciucci et al. 2016, Ministero dell’Ambiente e  
134 della Sicurezza Energetica & ISPRA 2016) (Fig. 1).



135

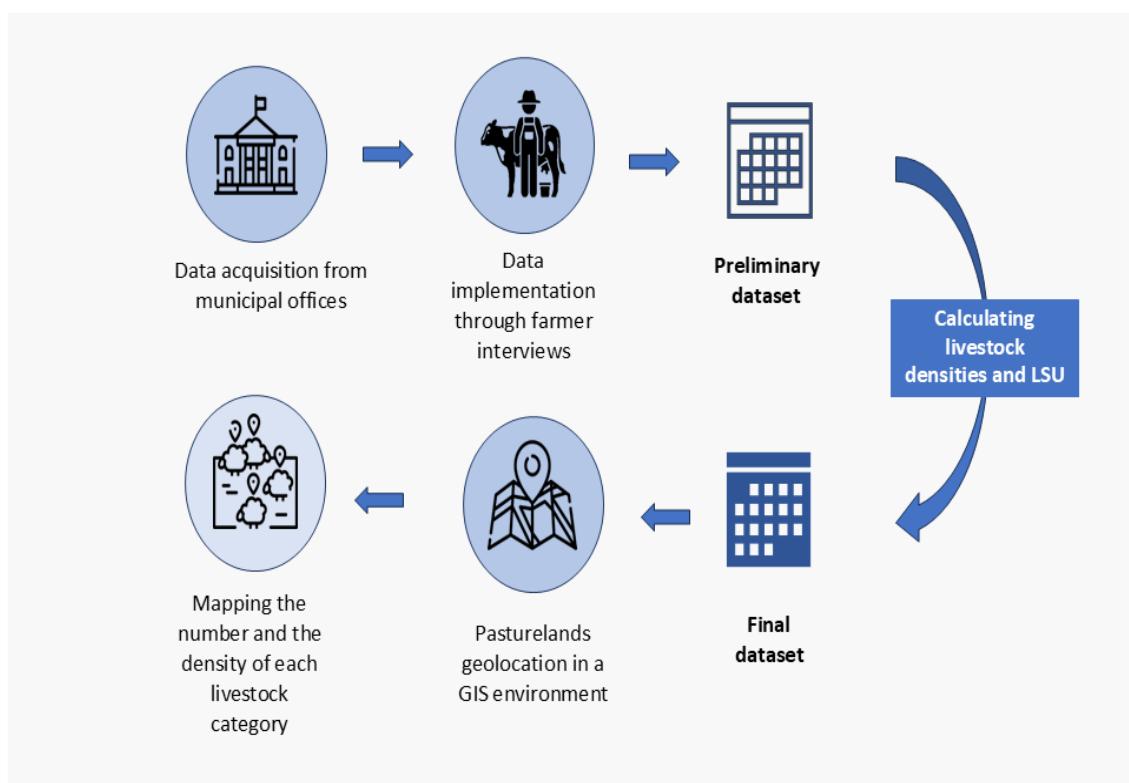
136 *Figure 1: Map of the study area, corresponding to two ecological corridors for the*  
137 *Marsican brown bear.*

138

139 Corridor 1 spans between the Sirente Velino Regional Natural Park and the Abruzzo,  
140 Lazio, and Molise National Park (ALMNP), while Corridor 2 connects ALMNP with  
141 the Majella National Park. These corridors facilitate movement for Marsican brown  
142 bears, but are also an important habitat for other mammal species, such as roe deer,  
143 red deer, wild boar, and porcupine (Dragonetti et al. 2024). Extensive livestock



144 grazing is common in these areas, where cattle and horses roam freely, while sheep  
 145 and goats are guarded by shepherds and dogs and are sheltered at night.  
 146 With this protocol, we collected and mapped livestock densities on a fine scale  
 147 based on data collected from individual municipalities. We requested from the  
 148 municipal offices of our study area the data on the number of livestock heads for  
 149 each municipal pastureland in 2023. We implemented this data with farmer  
 150 interviews, and we calculated livestock load and densities. Finally, we geolocated the  
 151 pasturelands in a GIS environment and integrated them with livestock load data to  
 152 create livestock distribution maps (Fig. 2).



153

154 *Figure 2: Graphic framework of methods adopted to collect and map livestock*  
 155 *densities.*

156

## 157 Data collection

158 As required by the legislation in force in Italy, municipal lands are entrusted to  
159 farmers in annual or seasonal concession under the "*fida pascolo*" system, regulated  
160 by *Legge 16 giugno 1927, n. 1766* and *Regio Decreto 6 febbraio 1928, n. 332*  
161 (Ministero della Giustizia 1927, 1928). This system regulates the allocation of  
162 municipal pasturelands to both resident and, in some cases, non-resident farmer  
163 applicants, who pay a fee for the exercise of the common grazing rights. Both the fee  
164 and the amount of land allocated vary based on the number of livestock heads  
165 owned by the applicant. Land boundaries are defined on the Italian cadastral map,  
166 which is divided into cadastral sheets (i.e., cadastral map sections that depict a  
167 specific area of a municipality) and particles (i.e., individual, numbered land parcels  
168 with the same type of crop within a cadastral sheet) as established by the  
169 Masedaglia Law, *Legge 1° marzo 1886, n. 3682* (Ministero della Giustizia 1886,  
170 Zonetti 2017).

171 Thus, we investigated livestock densities and geographic distribution by focusing on  
172 municipal grazing lands, excluding livestock held in farms with private pasturelands.  
173 We identified 15 municipalities that fall entirely or partially within our study area:  
174 Villalago, Secinaro, Scanno, Rocca Pia, Pettorano sul Gizio, Pescina, Ortona dei  
175 Marsi, Introdacqua, Goriano Sicoli, Gagliano Aterno, Cocullo, Castelvechio  
176 Subequo, Castel di Ieri, Bugnara, Anversa degli Abruzzi. We then requested  
177 municipal offices to provide public documents pertaining to the civic use of grazing  
178 on municipal properties. Data obtained from the municipal records included both the  
179 cadastral sheets and particles assigned to each livestock farm and the number of  
180 livestock heads for each farm in the municipality. Data quality varied across

181 municipalities, and some datasets lacked complete animal category breakdowns or  
182 livestock distribution in cadastral sheets or particles. To address data gaps and  
183 inconsistencies, we conducted additional interviews with farmers. Interviews and on-  
184 site visits to livestock farms also provided data on the exact location of pastures, in  
185 terms of cadastral sheet, as well as exact animal numbers by species (i.e., cattle,  
186 sheep, goats, horses) and age class, essential for calculating the total grazing  
187 pressure. We conducted these interviews anonymously, and we only provide  
188 aggregate data to protect the identity and location of each farm (Appendix S1).

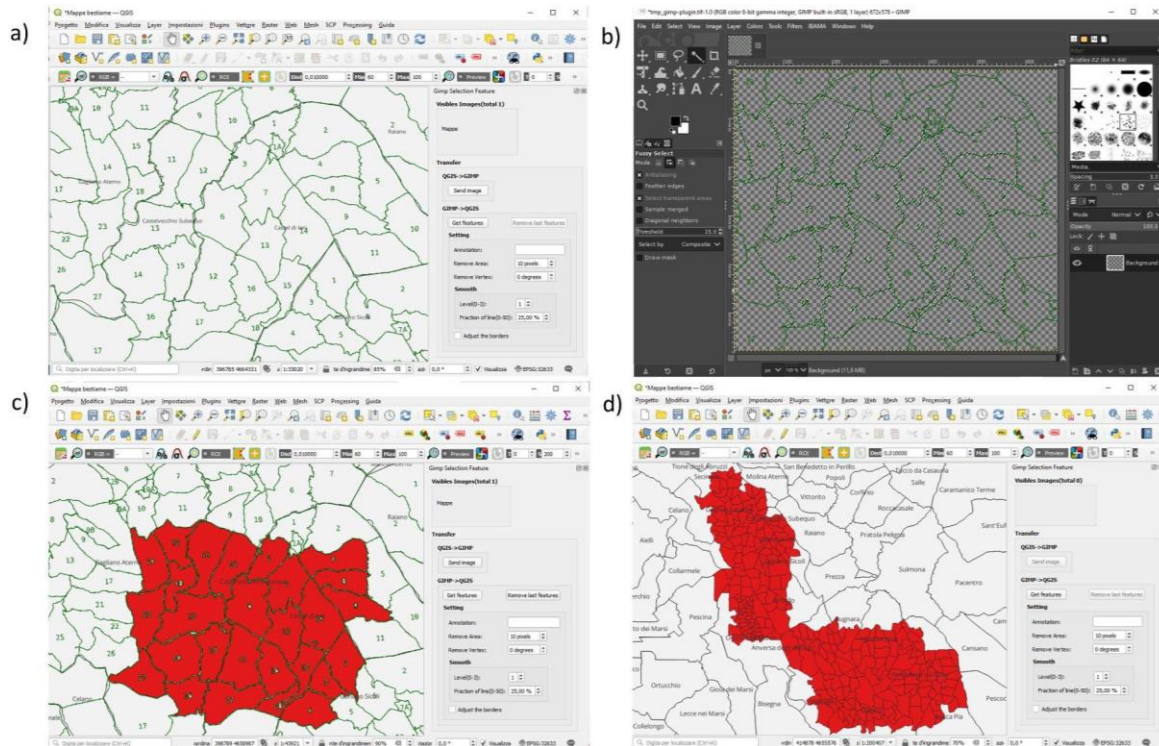
189 At the end of the data collection, we obtained a comprehensive database with the  
190 following details: municipality, farm's identification code (anonymised), cadastral  
191 sheet, cadastral sheet area assigned to the farm (hectares), number of farm-raised  
192 cattle >24 months, number of farm-raised cattle 6-24 months, number of farm-raised  
193 cattle <6 months, number of farm-raised sheep/goats >12 months, number of farm-  
194 raised horses >6 months (Table S1). Our preliminary dataset details, for each  
195 cadastral sheet in a municipality, the land area allocated to a specific farm, as well  
196 as the corresponding livestock number of each farm in that municipality, categorised  
197 by species and age.

198

## 199 Mapping livestock density

200 Using the QGIS software (QGIS.org 2023), through the GIMP plugin (Motta 2020)  
201 we vectorized in the map of the Cadastral Cartography available as Web Map  
202 Service (WMS) from the Italian national territory on the *Agenzia delle Entrate* website  
203 (Agenzia delle Entrate 2023). We manually selected the cadastral sheets falling  
204 entirely or in part in the study area, and exported each cadastral sheet selected in

205 GIMP in shapefile format. Finally, we merged all the polygons into a single layer,  
206 creating a vector map of the cadastral sheets of the study area (Fig. 3).



207

208 *Figure 3 - a) Detailed view of the area of interest. The image is then sent with the*  
 209 *command “send image” to GIMP b) On GIMP the cadastral sheets were selected*  
 210 *with the “magic wand” tool c) With the command “Get features” the selected items on*  
 211 *GIMP are loaded in QGIS, vectorialised and a new layer is created d) Resulting map*  
 212 *of all the cadastral sheets of the entire study area.*

213

214 We selected the sheets intended for "fida pascolo" to calculate the total surface of  
 215 the grazing areas used by each one of the farms. We had four types of information to  
 216 combine: the total number of livestock heads, for each livestock category, associated  
 217 with each farm  $i$  in a municipality  $m$  ( $L_{im}$ ), the areal coverage of each farm's pasture  
 218 in a municipality ( $Area_{im}$ ), the areal coverage of each farm's pasture in a cadastral  
 219 sheet  $s$  ( $Area_{is}$ ), the areal size of each sheet ( $Area_s$ ).

220 We then calculate the density  $D_s$  (n livestock/ha) of each category of livestock for  
 221 each cadastral sheet in a municipality, assuming a homogeneous distribution of  
 222 livestock within each sheet. This was done according to a proportional allocation  
 223 process, as follow:

224 Eq. 2

$$225 \quad D_s = \sum_{i=1}^n \frac{L_{im}}{Area_{im}} \times \frac{Area_{is}}{Area_s}$$

226 In the case of Gagliano Aterno municipality, we only had information on which  
 227 cadastral sheets were grazed, but not on the number of livestock grazing for each  
 228 sheet. Therefore, we calculated the overall density for the entire grazed area of the  
 229 municipality. Then, for each livestock category, we simply divided the total number of  
 230 livestock heads ( $L_i$ ) of each farm  $i$  by the total grazed area of the municipality  
 231 ( $Area_{gm}$ ), as follow:

232 Eq. 3

$$233 \quad D_m = \sum_{i=1}^n \frac{L_i}{Area_{gm}}$$


234 To calculate the total grazing pressure, we used the LSU (Livestock Unit) conversion  
 235 factor. The LSU has the purpose of synthetically expressing the livestock load, so  
 236 that the environmental impact of different farmed animals can easily be compared.  
 237 We referred to the conversion values of the Commission Implementing Regulation  
 238 (EU) 2016/669 (European Commission, 2016) (Table 1), as these were the same  
 239 coefficients used by the municipalities. Finally, we associated these densities of each  
 240 livestock category to the vector map of the cadastral sheets of the study area.

241

242

243

244 Table 1 - Conversion rates of free and semi-free ranging animals to livestock units  
 245 referring to the European Commission Implementing Regulation 2016/669.

	<b>Conversion rates of animals to livestock units (“LSU”)</b>	
	Bulls, cows and other bovine animals over two years and equine animals over six months	1 LSU
	Bovine animals from six months to two years	0.6 LSU
	Bovine animals below six months	0.4 LSU
	Sheep and goats	0.15 LSU

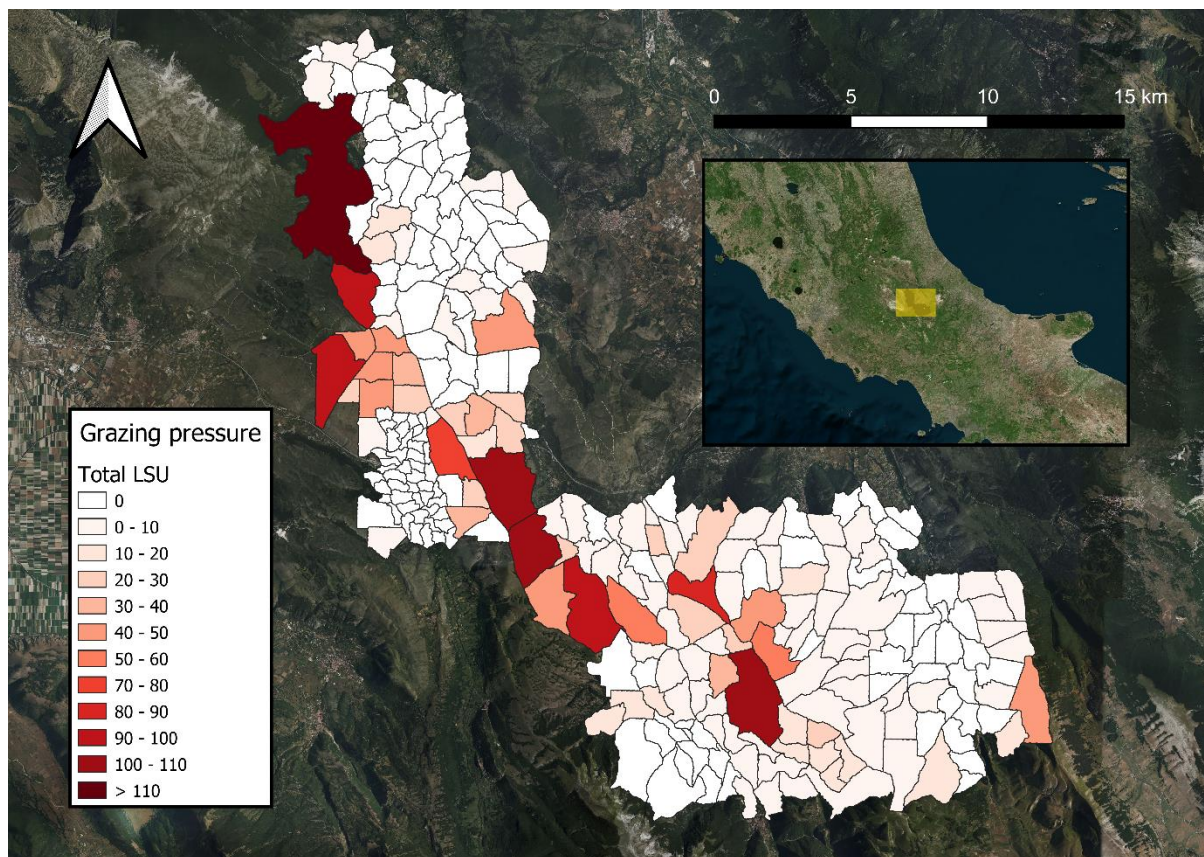
246

247 We compared our total livestock unit (LSU) data at the municipal level with that of  
 248 Malek et al. (2024a), which calculated LSUs for each European administrative unit.  
 249 Additionally, we compared our total LSUs at the cadastral sheet level with Malek et  
 250 al. (2024b) LSU estimates for semi-natural or managed grazed areas, aggregating  
 251 their data at the cadastral sheet level (Fig. S1 and S2). In doing both comparisons,  
 252 we excluded horses from our total LSU, as the other datasets do not consider  
 253 equines. We verified the correlation between our data and data from both studies  
 254 (Spearman test).

## 255 Results

256 We obtained a shapefile of grazing pressure, divided into eight livestock categories  
 257 at cadastral sheet resolution (LSU, LSU density, equines, equines density, cattle,  
 258 cattle density, sheep+goats and sheep+goats density; Fig. 4). The shapefile attribute  
 259 table contains 11 columns, each indicating the number and density of each type of  
 260 livestock listed above, for each sheet of each municipality.





261

262 *Figure 4 - Map of total grazing pressure in each cadastral sheet, in terms of livestock*  
 263 *unit (LSU).*

264

265 We found that Corridor 1 is generally more grazed than Corridor 2, both in terms of  
 266 absolute numbers and relative density. Among the municipalities with the highest  
 267 LSU counts, Gagliano Aterno ranked first (461.8 LSU), followed by Ortona dei Marsi  
 268 (373.2 LSU), Anversa degli Abruzzi (350.7 LSU) and Scanno (295.2 LSU).  
 269 Regarding cattle, Gagliano Aterno showed the highest numbers (366 cows), followed  
 270 by Ortona dei Marsi (204) and Cocollo (166). However, in terms of cattle density,  
 271 Ortona dei Marsi has the highest grazing intensity (2.03 cattle/ha in grazed cadastral  
 272 sheets), followed by Scanno (0.96 cattle/ha).

273 For sheep and goats, Anversa degli Abruzzi showed the highest values (1,173  
 274 animals with a density of 8.21 individuals/ha in grazed cadastral sheets), followed by



275 Bugnara and Pescina in terms of absolute numbers, but with Scanno ranking third in  
276 terms of density (5.38 individuals/ha). Regarding equines, Ortona dei Marsi has both  
277 the highest number (117 individuals) and the highest density (1.62 individuals/ha),  
278 followed by Scanno and Gagliano Aterno.

279 When comparing our data with Malek et al., 2024a and Malek et al., 2024b, we found  
280 our estimates being generally higher than those extracted from both studies, except  
281 for some municipalities (i.e., Pescina, Pettorano sul Gizio, Ortona dei Marsi,  
282 Scanno), which showed higher values in the comparison study of Malek et al., 2024a  
283 (Fig. S2). This was due to the methodological differences and the use of different  
284 coefficients for LSU calculations. Overall we found a moderate correlation between  
285 our values and the ones extracted from Malek et al. 2024a (Spearman = 0.47,  $p =$   
286 0.08), but a weak correlation with Malek et al. 2024b (Spearman = 0.14,  $p = 0.02$ ).

## 287 Discussion

288 Many studies simply compare 'grazed' to 'ungrazed' conditions and the measures of  
289 grazing intensity at a local scale come in different forms, almost always generalised  
290 without a distinction between different types of livestock, thus making the  
291 comparison across studies difficult (Briske et al. 2011, Schieltz and Rubenstein  
292 2016). In order to compensate for this lack of standardisation of grazing pressure  
293 measurements, and to obtain precise information on the actual distribution of free  
294 and semi-free ranging livestock, the introduction of a well-structured protocol for  
295 mapping grazing pressure with a standardised data collection method represents an  
296 important tool in this sense.

297 It is important to point out that the preliminary data collection method may vary  
298 across different countries, as the legislation in force may require the registration of  
299 individual livestock on different databases and regulate grazing activity in different  
300 ways. Based on the level of data accessibility, information on the distribution and  
301 actual size of the grazing livestock load may be completed and refined with specific  
302 interviews at livestock farms or by consulting different databases. In any case, the  
303 applicability of this protocol is linked to the processing and subsequent mapping of  
304 this data according to a precise map unit, selected based on the precision of the data  
305 collected and the spatial resolution desired. Another limitation pertains to the limited  
306 temporal validity of the results obtained with this protocol as well, as the concession  
307 of municipal lands to farms is annual and may change from year to year (at least in  
308 Italy). In order to overcome the problems related to the different spatial resolutions of  
309 the collected data, we decided to assume a homogeneous distribution of livestock  
310 within the farms, and to group the data provided at the resolution of cadastral  
311 particles within the related cadastral sheets. This assumption allowed us to use data  
312 with different spatial precision but might not necessarily hold because food resources  
313 for grazing animals may not be equally distributed in the territories granted to the  
314 farms or because some animals, such as sheep and goats, may move in herds,  
315 concentrating the grazing pressure in specific areas.

316 We compared our grazing assessment with two previous studies conducted at a  
317 larger spatial scale (Fig. S1, S2), and identified key differences. Our analysis  
318 revealed differences in both calculation methods and final data, which are reflected  
319 in the applicability of the dataset. For instance, both of Malek's studies excluded  
320 horses from their grazing assessments, a livestock category which could significantly  
321 influence grazing dynamics in many areas.

322 We excluded livestock held on private grazing lands from our analysis because this  
323 data was unavailable from the WMS of the “Agenzia delle Entrate” nor by individual  
324 municipalities. Private grazing lands constitute only a small portion of the total  
325 grazing area within our study region, hence any underestimation of grazing intensity  
326 is likely small in our case. However, we acknowledge that private and grazing could  
327 be an element of higher importance, and deserving deeper investigation, if  
328 transferring our framework to other areas.

329 Mapping grazing intensity allows us to quantify livestock pressure on ecosystems,  
330 which can then serve different purposes. Our data can support conservation  
331 strategies that help local communities, their activities, and wildlife to coexist. Our  
332 study can be used in combination with land-cover maps, to obtain a further finer  
333 grazing allocation and thus a more accurate density estimates actual pasturelands in  
334 each cadastral sheet (i.e., mapped at a higher resolution than we did here). By  
335 identifying areas where human activities occur in natural and semi-natural  
336 environments, and evaluating the biodiversity conditions, it is possible to identify and  
337 promote sustainable agriculture and pastoralism practices.

338 This knowledge is particularly relevant for our study area in the central Apennines,  
339 identified as a corridor for the Marsican brown bear. Here, we identified  
340 municipalities such as Scanno or Gagliano Aterno as areas that need an  
341 assessment of potential overgrazing, together with the monitoring of the interactions  
342 between pasture activity and natural systems. In this sense, our study is a useful tool  
343 to preserve bear’s suitable habitats from excessive disturbance and degradation.  
344 Additionally, our work is a useful tool to assess eventual zoonotic risks, as there is  
345 increasing interest in examining the impact of livestock-borne pathogens on the

346 bear's health (Fico et al. 2019), and more in general the risk of a two-way pathogen  
347 transmission between farmed and wild animals. Our study is also critical to support  
348 conservation strategies of many other species living in these areas, including those  
349 living in ecotonal semi-natural environments, shaped and maintained by extensive  
350 grazing (Dragonetti et al. 2024).

351 Considering the influence that livestock has on the temporal and spatial behaviour of  
352 wildlife (Schieltz and Rubenstein 2016), in pathogen transmission (Jori et al. 2021) or  
353 on changes in vegetation structure and cover (Augustine and McNaughton 1998),  
354 accurate mapping of grazing livestock is an important tool in land management  
355 planning and biodiversity conservation.

356

## 357 Data Availability

358 All maps are available in GeoTIFF format and are freely accessible in the Zenodo  
359 repository: <https://zenodo.org/records/14832257>.

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