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Geographical patterns in the distribution of naturalized plants in Central America

 Eduardo Chacón-Madrigal,  Julissa Rojas-Sandoval, Lilian Ferrufino-Acosta, Rodolfo Flores, Pablo Galán, AnaLu MacVean, Dagoberto Rodríguez Delcid, Iris Saldivar-Gómez, Yader Ruiz

1 **Geographical patterns in the distribution of naturalized plants in Central America**

2 Eduardo Chacón-Madrigal^{1,2}, Julissa Rojas-Sandoval^{3,4}, Lilian Ferrufino-Acosta⁵, Rodolfo
3 Flores^{6,7}, Pablo Galán⁸, AnaLu MacVean⁹, Dagoberto Rodríguez-Delcid¹⁰, Iris Saldivar-Gómez¹¹
4 & Yader Ruiz¹²

5 1. Herbario Luis Fournier Origgi, Centro de Investigación en Biodiversidad y Ecología Tropical
6 (CIBET), Universidad de Costa Rica

7 2. Herbario Nacional, Departamento de Historia Natural, Museo Nacional de Costa Rica

8 3. Institute of the Environment & Energy, University of Connecticut, Storrs, CT, USA.

9 4. Department of Geography, Sustainability, Community, and Urban Studies, University of
10 Connecticut, Storrs, CT, USA.

11 5. Escuela de Biología, Facultad de Ciencias, Universidad Nacional Autónoma de Honduras,
12 Honduras.

13 6. Los Naturalistas, P.O. Box 0426-01459, David, Chiriquí, Panamá

14 7. Departamento de Botánica, Universidad de Panamá, Estafeta Universitaria, Panama City

15 8. Herbario TECLA, Banco de Germoplasma, Centro Nacional de Tecnología Agropecuaria y
16 Forestal CENTA, El Salvador

17 9. Environmental Horticulture Department, York College of Pennsylvania, PA, USA

18 10. Herbario LAGU, Jardín Botánico La Laguna, La Libertad, El Salvador.

19 11. Centro de Investigación Capacitación y Formación Ambiental. CICFA, Nicaragua

20 12. Departamento de Ciencias Naturales, Facultad Multidisciplinaria Oriental, Universidad de El
21 Salvador.

22 Corresponding author: Eduardo Chacón Madrigal, Herbario Nacional, Departamento de Historia
23 Natural, Museo Nacional de Costa Rica, Email: edchacon@gmail.com

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26 **Abstract**

27 Non-native plant species are increasing globally, yet their distribution patterns and
28 environmental drivers remain poorly understood in biodiversity-rich but understudied regions
29 like Central America. In this region, non-native species increasingly affect biodiversity,
30 ecosystem integrity, and conservation efforts, especially when they become invasive. We
31 analyzed the spatial distribution of 751 naturalized plant species using over 42,000 occurrence
32 records across the seven countries in Central America. We evaluated the influence of
33 environmental variables, human population density, protected areas, and life zones on both
34 occurrence and species richness. Human population density emerged as the strongest predictor of
35 naturalized species occurrence and richness, highlighting the role of human activity in
36 facilitating invasions. Annual precipitation was positively associated with occurrences, while
37 species richness declined with increasing temperature and biodiversity integrity. Tropical
38 rainforests and other humid life zones had more naturalized species than expected by chance.
39 Protected areas had fewer naturalized species overall, but a higher species ratio per observation,
40 indicating both conservation value and vulnerability. Rare species were found outside protected
41 zones, particularly in disturbed and urbanized areas. Our findings highlight the need for early
42 detection, targeted management, and strengthened protection strategies, especially in mid-
43 elevation zones and densely populated areas. By identifying key environmental and
44 anthropogenic drivers, this study provides actionable insights for conservation planning and
45 invasive species management in one of the world's most biodiverse and socio-environmentally
46 vulnerable regions.

47 **Keywords:** Invasive Species, Biodiversity, Protected Areas, Central America, Population
48 Density

49

50 **Introduction**

51 Naturalized plant species, those that establish self-sustaining populations outside their native
52 range, play a complex and growing role in affecting global change and biodiversity conservation
53 (IPBES 2023). When naturalized species become invasive, they can severely impact ecosystem
54 function, economic stability, food and water security, human health, and even cultural identities

55 (Pimentel et al. 2001; Clements et al. 2021; Heringer et al. 2021; Bacher et al. 2023).
56 Understanding naturalized species' drivers and distribution patterns is essential for developing
57 effective management strategies for biodiversity conservation and ecological resilience. Despite
58 substantial progress in invasion ecology, significant knowledge gaps remain, particularly
59 regarding the interplay between environmental factors and human activities in shaping
60 naturalization patterns.

61 Central America is a biodiversity hotspot with exceptional species richness and endemism
62 (Coates 1999). However, the region also faces a high degree of ecological and socioeconomic
63 vulnerability, shaped by a long history of colonialism, political instability, inequality, poverty,
64 inadequate use of territory, and rapid unplanned urbanization (Morales-Marroquín et al. 2022).
65 These pressures are exacerbated by limited investment in scientific research, weak conservation
66 policies, and poor regulation of activities that affect native ecosystems (Harvey et al. 2005). For
67 instance, managing non-native invasive and naturalized species is rarely addressed in
68 environmental regulations and policies across Central American countries (Chacón-Madrigal et
69 al. 2022). Thus, empirical research on plant invasions in the region remains scarce compared to
70 other regions.

71 Recent studies have begun to address long-standing knowledge gaps in our understanding of
72 plant invasions in Central America (see Avalos et al. 2021; Chacón-Madrigal et al. 2022; Rojas-
73 Sandoval et al. 2022; MacVean and Zinn 2023). These efforts have provided valuable baseline
74 data, such as species inventories, introduction pathways, and invasion histories, and emphasize
75 the pressing need for more coordinated, region-wide research on plant invasions. A persistent
76 challenge identified across these studies is the lack of updated herbarium records, and systematic
77 botanical surveys focused specifically on non-native species. For example, our previous work
78 documented 1,228 species in the alien flora of Central America, including 835 naturalized
79 species and 393 casuals (Rojas-Sandoval et al. 2022). That study also identified major
80 introduction pathways, economic uses, and country-specific patterns. However, this region still
81 has limited fine-scale analyses of distributions of species and the environmental and human
82 factors influencing their spread. Addressing these gaps is crucial, as they constrain the capacity
83 of this region to develop predictive tools and proactive strategies for managing invasive species
84 and mitigating their ecological and socioeconomic impacts.

85 In this study, we evaluate the distribution patterns of naturalized species across Central America
86 in relation to key environmental and anthropogenic factors. Specifically, we assess the
87 distribution of these species across ecosystems and life zones, protected areas vs. unprotected
88 areas, gradients of human population density, and climatic variables such as annual mean
89 temperature and precipitation. By identifying the ecological and socioeconomic drivers of
90 naturalized species distributions, we aim to conduct future research, guide policy development,
91 and support evidence-based management strategies. Our findings provide crucial insight into the
92 role of protected areas in buffering against biological invasions and maintaining biodiversity, as
93 well as the vulnerability of disturbed landscapes and the potential for climatic and demographic
94 trends to shape future invasions. Finally, this research contributes to the scientific understanding
95 of biological invasions in the tropics and offers practical guidance for preserving the ecological
96 integrity of diverse landscapes in a region of global conservation significance.

97 **Methods**

98 *Study site*

99 We analyzed seven countries in Central America: Belize, Costa Rica, El Salvador, Guatemala,
100 Honduras, Nicaragua, and Panama. These countries cover a total area of 525,300 km² (World
101 Bank 2022) and encompass a complex and diverse geography, including mountain ranges,
102 volcanoes, plains, and significant bodies of water. Elevation ranges from sea level to 4,220 m
103 a.s.l. at Tajumulco Volcano in Guatemala. The mean annual temperature is 24 °C, with average
104 minimum temperatures ranging from 0°C during the coldest month in the highlands and
105 maximum temperatures around 30°C during the hottest month in the lowlands (Taylor and
106 Alfaro 2005). Annual rainfall varies widely due to topographic and elevational gradients. For
107 instance, El Salvador and specific parts of Guatemala, Honduras, and Nicaragua receive less than
108 1,000 mm annually. In contrast, other regions in Guatemala, Panama, and Costa Rica receive
109 more than 2,500 mm. The region has distinct wet and dry seasons, with the wet season typically
110 extending from May to November (coinciding with increased hurricanes and tropical storm
111 activity), while the dry season extends from December to April. The Caribbean coast is generally
112 more humid and receives more rainfall than the Pacific coast (Taylor and Alfaro 2005).

113 Central America is a narrow land bridge connecting North and South America and separating the
114 Pacific Ocean from the Caribbean Sea. As a corridor between two major biogeographical

115 regions, it harbors ca. 7 % of the world's plant and animal species (Greenheck 2002). Despite its
 116 rich biodiversity and abundant natural resources, the region faces significant environmental
 117 challenges. It is considered a threatened biological hotspot with high conservation priority due to
 118 habitat degradation, soil erosion, water pollution, and climate change (Harvey et al. 2005).
 119 Deforestation, driven primarily by agriculture, is a major concern. Large portions of the forest
 120 have been cleared for cattle ranching, palm oil plantations, and other agricultural activities.
 121 Estimates suggest that less than 20% of the region retains dense forest cover, and much of the
 122 remaining forest is highly fragmented or at risk of conversion (Hoang and Kanemoto 2021).

123 The population of Central America is ca. 50 million people. The country with the highest
 124 population is Guatemala, and the smallest is Belize. The region has a high population growth
 125 rate, along with accelerating urbanization and migration. Economically, Central America relies
 126 heavily on agriculture, with coffee, bananas, sugarcane, and other crops as key exports (Grau and
 127 Aide 2008). However, recent years have shifted towards manufacturing and service industries,
 128 particularly in Costa Rica and Panama. According to the World Bank, nearly 30% of the
 129 population lives below the poverty line, with rural areas and indigenous communities especially
 130 vulnerable. The region is highly vulnerable to climate change due to its geographic location,
 131 socioeconomic inequalities, agricultural dependence, low educational levels, and weak
 132 infrastructure. Expected impacts of climate change include more frequent and intense hurricanes,
 133 floods, landslides, and droughts (Hannah et al. 2017; Donatti et al. 2019), leading to potential
 134 loss of life, displacement, and economic disruption. Rising sea levels may threaten coastal
 135 communities and infrastructure, and growing water scarcity could further exacerbate food
 136 insecurity (Hagen et al. 2022).

137 Due to its strategic geographic position, Central America is one of the most important global
 138 trade routes (e.g., the Panama Canal). As a result, the region is undergoing rapid economic
 139 development and increased international trade (Kerf 2021), which have ecological consequences,
 140 including pollution, resource degradation, and overexploitation. Infrastructure expansion, such as
 141 roads, ports, and urban areas, further contributes to environmental disturbances (ECLAC 2015)
 142 and may facilitate the introduction and spread of non-native species.

143 *Data collection and analysis*

144 This study builds on an open-access dataset of 1,228 non-native plant species occurring across
145 Central America compiled by Rojas-Sandoval et al. (Rojas-Sandoval et al. 2022) as part of the
146 FINCA Project (Flora Introduced and Naturalized in Central America:
147 <https://finca.collaboration.uconn.edu/>). In the FINCA dataset, species were classified as either
148 casuals or naturalized. We revised and expanded the original dataset for this study, incorporating
149 updated information for invasion status. Following Richardson et al. (2000) and Blackburn et al.
150 (2011), species were reclassified as cultivated, casual, naturalized, or invasive. For those species
151 classified as naturalized or invasive, we searched for occurrence records in the Global
152 Information Biodiversity Facility (GBIF: <https://www.gbif.org/>). For all the records in the
153 dataset, we applied the workflow by Seebens and Kaplan (2022), which includes data-cleaning
154 procedures to eliminate errors in geographic coordinates. Finally, we restricted our search to
155 records of each species in the country listed in the FINCA dataset.

156 We conducted correlation analyses between the number of occurrences and species by country
157 and three national-level variables: land area (km²), population size (people), and Gross Domestic
158 Product (GDP) (Billions US\$) for the year 2022. Country-level data were obtained from the
159 World Bank Data (World Bank 2022, 2025a, b). We calculate Pearson correlation coefficients
160 using R software (R Development Core Team 2022).

161 For independent variables, we use annual mean temperature (AMT), annual precipitation (AP),
162 human population density (human population), and a biodiversity integrity index. Climate data
163 (AMT and AP) at 5-minute resolution were obtained from WorldClim (Fick et al. 2017). Human
164 population density data were obtained from WorldPop Hub (2020) at 1 km resolution
165 (<https://hub.worldpop.org/>). Biodiversity integrity data were derived from Gassert et al. (2022).
166 All datasets were *resampled* to obtain the same resolution using bilinear interpolation with R's
167 *raster* package (Hijmans 2023). Species occurrence data were rasterized at a 5-minute resolution
168 to estimate both the number of naturalized species records and the number of different species
169 per cell. We used a mixed-effect regression model to assess the influence of independent
170 variables on the number of records. Latitude and longitude were included as random effects
171 using a Matérn correlation matrix to control spatial autocorrelation. The country was included as
172 a fixed effect. We used the *fitme* function from the package *spaMM* (Rousset and Ferdy 2014).
173 Species richness was analyzed with a similar model, incorporating the number of records as an

174 additional fixed effect. We generated maps of predicted records and predicted richness for each
175 model.

176 To evaluate whether life zones influence the number of occurrences of naturalized species, we
177 used a contemporary map of life zones from Elsen et al. (2021). We calculated the number of
178 raster cells per life zone and the number of cells with naturalized species. The proportion of
179 occupied cells was compared to a random distribution by sampling the map 1,000 times and
180 estimating mean and 95% confidence intervals. We plotted differences between observed and
181 expected proportions per life zone. We also use the *climenv* package (Tsakalos et al. 2023) to
182 visualize species occurrences within a climatic space diagram compared to the overall
183 availability in Central America.

184 Vector *shapefile* of protected areas for each Central American country were obtained
185 (Supplementary Information Table 1), and a 1 km² resolution raster map of protected areas was
186 created. We calculated the number of observations and species inside and outside protected areas
187 and determined the number of cells with naturalized species for each category (i.e., with and
188 without protection). Proportions were compared using a chi-square test with Yates's correction.
189 We also used rarefaction analysis with Hill numbers of order 0, 1, and 2 to compare species
190 richness between protected and unprotected cells (Chao et al. 2014) using the *iNEXT* package
191 (Chao et al. 2014). All the analyses were conducted in *R* (R Development Core Team 2022).

192 **Results**

193 We compiled 42,658 occurrence records for 751 naturalized plant species across Central
194 America. The number of records was positively associated with the number of species (SI. Fig.
195 1). However, records were unevenly distributed among countries; Costa Rica accounted for the
196 highest number of species (19,179), while Belize had the fewest (909; Table 1). Neither the
197 number of occurrence records nor the number of species per country was correlated with land
198 area, population size, or GDP (SI Fig. 2).

199 At the spatial cell level, population density and annual precipitation significantly explained the
200 number of occurrence records (Table 2). Cells with higher population density and precipitation
201 levels had more occurrences of naturalized species. In contrast, AMT and the biodiversity
202 integrity index did not significantly affect the number of records (Table 2). The species richness

203 of naturalized plants was significantly influenced by population density, AMT, and biodiversity
204 integrity index (Table 2). Specifically, species richness increased with population density, while
205 it decreased in areas with higher AMT and higher biodiversity integrity values (Table 2).

206 When comparing life zones, we found no significant deviation from random expectations in the
207 proportion of cells observed with naturalized species for tropical moist, tropical dry, and
208 premontane moist forests (Figs. 2 and 3). In contrast, all other life zones showed a higher-than-
209 expected proportion of cells with naturalized species. The tropical rainforest had the largest
210 positive deviation from random expectations (Figs. 2 and 3), with nearly 4% more occupied cells
211 than expected. These findings are consistent with the spatial model results.

212 Regarding protected areas, we found 33,211 occurrences of naturalized species outside protected
213 areas, compared to 8755 observations inside. In total, 729 naturalized species were recorded
214 outside protected areas, while 540 were found within protected areas (Fig. 4). Most species
215 occurring exclusively outside protected areas were observed at low frequencies (Fig. 4). No
216 significant differences were observed in the number of common or frequent species between
217 protected and unprotected areas (Fig. 4). The proportion of cells with naturalized species was
218 significantly different between protected and unprotected areas ($\chi^2 = 296$, $df = 2$, $p > 0.001$), with
219 higher numbers of naturalized species occurring outside protected areas (0.029) than inside
220 (0.019). However, the observation density was similar, with ca. 6.9 records per 100 km² outside
221 protected areas versus ca. 6.5 records per 100 km² inside. Interestingly, the ratio of species to
222 observations was higher within protected areas, averaging ca. six species per 100 records
223 compared to ca. two species per 100 records outside protected areas.

224 Discussion

225 Our findings reveal a clear pattern in the spatial distribution of naturalized plant species in
226 Central America, offering insight into future management of invasive species and the prevention
227 of biological invasions. Among all predictors, human population density emerged as a strong
228 determinant of both the number of occurrences and the richness of naturalized species. This is
229 consistent with global (Essl et al. 2019) and regional studies, e.g., China (Liu et al. 2005), South
230 Africa (Spear et al. 2013), and supports hypotheses linking invasion patterns to propagule
231 pressure (Simberloff 2009; De Jong and Fowler 2018), disturbance (Lozon and MacIsaac 1997;

232 González-Moreno et al. 2015), and deliberate or accidental human-mediated introductions
233 (Zimmermann et al. 2014).

234 While herbarium records are known to be spatially and temporally biased, favoring urban centers
235 and accessible areas, our models accounted for this by including both sampling effort (number of
236 occurrences) and country as fixed effects. Even after this correction, population density remained
237 a significant predictor, highlighting its robust association with invasion patterns. The tendency
238 for early records of non-native species to appear near urban areas (Aikio et al. 2012) may reflect
239 higher propagule pressure and the role of ornamental plant use in urban landscapes (Mayer et al.
240 2017; van Kleunen et al. 2018; Abonyo and Oduor 2025). Our rarefaction analysis reinforces the
241 pattern. Species richness was higher outside protected areas, although there was no difference in
242 the number of common or frequent species between protected and unprotected areas. This
243 suggests that less frequent species are more likely to be found in disturbed, non-protected
244 regions. These areas should be prioritized for early detection and rapid response strategies to
245 prevent the spread of potentially invasive plants into natural or protected ecosystems. The
246 negative association between species richness and biodiversity integrity index also supports the
247 idea that degraded environments, often associated with human activity and high propagule
248 pressure, are more susceptible to invasions (Marvier et al. 2004; Rojas-Sandoval and Ackerman,
249 2021; Rojas-Sandoval et al. 2024).

250 Interestingly, regions with high population density are not necessarily of low conservation value.
251 Many urban adjacent protected areas, such as Braulio Carrillo National Park (Costa Rica), La
252 Tigra National Park (Honduras), and El Boquerón National Park (El Salvador), safeguard critical
253 ecosystems and deliver essential services, including water supply, air filtering, and biodiversity
254 protection (Torres-Miranda et al. 2011, McDonald et al. 2008, Brassard et al. 2021). Although
255 our results show more naturalized species outside protected areas, the ratio of species per
256 observation was higher within protected areas. This suggests that protected areas support
257 relatively diverse non-native floras despite lower propagule pressure and reduced anthropogenic
258 disturbance, highlighting their vulnerability to biological invasions (Ackerman et al. 2017).

259 We also found that the physical environment played an important role in shaping the number of
260 occurrences and species richness of naturalized species. Annual precipitation was positively
261 associated with the number of occurrence records but not species richness, while mean annual

262 temperature had a weak negative effect on species richness alone. These findings align with
263 broader global patterns (Essl et al. 2019). A similar trend emerged from our life zone analysis:
264 life zones characterized by higher humidity and temperature, such as tropical rainforests and
265 tropical wet forests, contained more occupied cells than expected under a random distribution.
266 This pattern likely reflects regional biogeographic conditions, as much of Central America lies in
267 humid tropical life zones, where water availability strongly limits plant establishment (Ratcliffe
268 et al. 2024, Engelbrecht et al. 2007). Although the effect of precipitation on invasive species
269 abundance has been observed in some regions (Ratcliffe et al. 2024), it is not globally consistent.
270 In our study, Central America's tropical setting may explain the positive association between
271 precipitation and the occurrence of naturalized species. Many naturalized plants in Central
272 America originate from tropical zones (Rojas-Sandoval et al. 2022, MacVean and Zinn 2023),
273 making them well-adapted to high rainfall environments where high solar radiation increases
274 transpiration, making water availability a critical filter for plant survival. Historical colonization
275 patterns may partially explain the weak negative effect of temperature on species richness.
276 During the Spanish conquest and colonial period, most villages and towns were established in
277 mid-elevation zones in areas with climates more like those in Europe (Heckadon-Moreno 1997).
278 As a result, plant species introduced during that time became established in these mid-elevations.
279 These historical introduction patterns likely continue to influence the current distribution of
280 naturalized species across the region.

281 **Conclusions**

282 Our study highlights the strong influence of population density on the occurrence and the spread
283 of naturalized species in Central America. These findings call for integrating conservation
284 strategies into urban planning, focusing on green spaces, early detection systems, and public
285 awareness of the risks associated with invasive species. Because densely human-populated areas
286 support high species richness and higher occurrences of naturalized species, management
287 strategies must balance biodiversity conservation and invasive species control. The associations
288 between annual precipitation and naturalized plant occurrences emphasize the importance of
289 considering climatic suitability in invasion risk assessments. Similarly, mid-elevation areas with
290 moderate temperatures, likely reflecting ecological and historical introduction patterns, may act

291 as hotspots for naturalized plants and should be prioritized for targeted monitoring and
292 management.

293 This study also emphasizes the importance of early detection in disturbed, non-protected areas,
294 which are likely entry points for invasive species. At the same time, protected areas, especially
295 those near urban centers, play a vital role in maintaining native biodiversity and providing
296 ecosystem services. Conservation strategies should include strengthened protection and proactive
297 management of these areas to ensure long-term sustainability for people and natural ecosystems.

298 Addressing sampling bias in herbarium records and increasing data collection in
299 underrepresented or remote regions will improve our understanding of the geographic and
300 ecological dynamics of plant invasions. Equally important is the involvement of local
301 communities in monitoring, education, and invasive control efforts. A holistic, multidisciplinary
302 approach combining ecological research, urban planning, policy development, and public
303 participation is essential for effectively managing naturalized plant species and safeguarding
304 biodiversity throughout Central America.

305 **References**

- 306 Abonyo CRK, Oduor AMO (2025) Urban green spaces as reservoirs of exotic plant species with
307 invasion risk: A case study on the ornamental flora of Nairobi City, Kenya. *Perspectives in*
308 *Plant Ecology, Evolution and Systematics*: 125864.
309 <https://doi.org/10.1016/J.PPEES.2025.125864>
- 310 Ackerman JD, Tremblay RL, Rojas-Sandoval J (2017) Biotic resistance in the tropics: patterns of
311 seed plant invasions within an island. *Biol Invasions* 19, 315–328.
312 <https://doi.org/10.1007/s10530-016-1281-4>
- 313 Aikio S, Duncan RP, Hulme PE (2012) The vulnerability of habitats to plant invasion:
314 Disentangling the roles of propagule pressure, time and sampling effort. *Global Ecology*
315 *and Biogeography* 21: 778–786. <https://doi.org/10.1111/J.1466-8238.2011.00711.X>
- 316 Avalos G, Chacón-Madrigal E, Artavia-Rodríguez LG (2021) Invasive Plants of Costa Rica
317 Current Status and Research Opportunities. In: *Invasive Alien Species*. Wiley, 57–76.
318 <https://doi.org/10.1002/9781119607045.ch37>

- 319 Bacher S, Galil BS, Nuñez MA, Ansong M, Cassey P, Dehnen-Schmutz K, Fayvush G,
320 Hiremath AJ, Ikegami M, Martinou AF, McDermott SM, Preda C, Vilà M, Weyl OLF,
321 Fernandez RD, Ryan-Colton E (2023) Chapter 4: Impacts of invasive alien species on
322 nature, nature's contributions to people, and good quality of life. In: Roy HE, Pauchard A,
323 Stoett P, Renard Truong T (Eds), Thematic Assessment Report on Invasive Alien Species
324 and their Control of the Intergovernmental Science-Policy Platform on Biodiversity and
325 Ecosystem Services. IPBES Secretariat, Bonn, Germany.
326 <https://doi.org/10.5281/zenodo.7430731>
- 327 Brassard F, Leong CM, Chan HH, Guénard B (2021) High diversity in urban areas: How
328 comprehensive sampling reveals high ant species richness within one of the most
329 urbanized regions of the world. *Diversity* 13: 358. <https://doi.org/10.3390/d13080358>
- 330 Chacón-Madrigal E, Avalos G, Hofhansl F, Coronado I, Ferrufino-Acosta L, MacVean A,
331 Rodríguez D (2022) Biological invasions by plants in continental Central America. In:
332 Clements DR, Upadhyaya MK, Joshi S, Shrestha A (Eds), *Global Plant Invasions*. Springer
333 International Publishing.
- 334 Chao A, Gotelli NJ, Hsieh TC, Sander EL, Ma KH, Colwell RK, Ellison AM (2014) Rarefaction
335 and extrapolation with Hill numbers: A framework for sampling and estimation in species
336 diversity studies. *Ecological Monographs* 84: 45–67. <https://doi.org/10.1890/13-0133.1>
- 337 Clements DR, Upadhyaya MK, Joshi S, Shrestha A (2021) *Global Plant Invasions* Global Plant
338 Invasions. Springer Nature, Cham, Switzerland, 1–381 pp. [https://doi.org/10.1007/978-3-](https://doi.org/10.1007/978-3-030-89684-3)
339 [030-89684-3](https://doi.org/10.1007/978-3-030-89684-3)
- 340 Coates AG (1999) *Central America: A Natural and Cultural History*. Yale University Press, 332
341 pp. Available from: http://books.google.co.cr/books?id=jG_TV-akte8C.
- 342 De Jong GL, Fowler NL (2018) Duration of propagule pressure affects non-native plant species
343 abundances. *American Journal of Botany* 105:197–206. <https://doi.org/10.1002/AJB2.1026>
- 344 Donatti CI, Harvey CA, Martinez-Rodriguez MR, Vignola R, Rodriguez CM (2019)
345 Vulnerability of smallholder farmers to climate change in Central America and Mexico:

- 346 current knowledge and research gaps. *Climate and Development* 11: 264–286.
347 <https://doi.org/10.1080/17565529.2018.1442796>
- 348 ECLAC (2015) *Climate Change in Central America: Potential Impacts and Public Policy*
349 *Options*. LC/MEX/L.1196/Rev.
- 350 Elsen PR, Saxon EC, Simmons BA, Ward M, Williams BA, Grantham HS, Kark S, Levin N,
351 Perez-Hammerle K-V, Reside AE, Watson JEM, Perez-Hammerle K (2021) Data from:
352 Accelerated shifts in terrestrial life zones under rapid climate change.
353 <https://doi.org/10.5061/dryad.41ns1rnff>
- 354 Engelbrecht BMJ, Comita LS, Condit R, Kursar TA, Tyree MT, Turner BL, Hubbell SP (2007)
355 Drought sensitivity shapes species distribution patterns in tropical forests. *Nature* 447: 80–
356 82. <https://doi.org/10.1038/nature05747>
- 357 Essl F, Dawson W, Kreft H, Pergl J, Pyšek P, Van Kleunen M, Weigelt P, Mang T, Dullinger S,
358 Lenzner B, Moser D, Maurel N, Seebens H, Stein A, Weber E, Chatelain C, Inderjit,
359 Genovesi P, Kartesz J, Morozova O, Nishino M, Nowak PM, Pagad S, Shu WS, Winter M,
360 Burns J (2019) Drivers of the relative richness of naturalized and invasive plant species on
361 Earth. *AoB PLANTS* 11: 1–13. <https://doi.org/10.1093/aobpla/plz051>
- 362 Fick SE, Hijmans RJ (2017) WorldClim 2: new 1-km spatial resolution climate surfaces for
363 global land areas. *International Journal of Climatology* 37: 4302–4315.
364 <https://doi.org/10.1002/joc.5086>
- 365 Gassert F, Mazzarello J, Hyde S (2022) Global 100 m Projections of Biodiversity Intactness for
366 the years 2017-2020. Available from:
367 [https://ai4edatasetspublicassets.blob.core.windows.net/assets/pdfs/io-](https://ai4edatasetspublicassets.blob.core.windows.net/assets/pdfs/io-biodiversity/Biodiversity_Intactness_whitepaper.pdf)
368 [biodiversity/Biodiversity_Intactness_whitepaper.pdf](https://ai4edatasetspublicassets.blob.core.windows.net/assets/pdfs/io-biodiversity/Biodiversity_Intactness_whitepaper.pdf).
- 369 González-Moreno P, Diez JM, Richardson DM, Vilà M (2015) Beyond climate: Disturbance
370 niche shifts in invasive species. *Global Ecology and Biogeography* 24: 360–370.
371 <https://doi.org/10.1111/geb.12271>
- 372 Grau HR, Aide M (2008) Globalization and Land-Use Transitions in Latin America. *Ecology*
373 *and Society* 13: art16. <https://doi.org/10.5751/ES-02559-130216>

- 374 Greenheck FM (2002) *Naturaleza, gente y bienestar: Mesoamérica en cifras*. Centroamericana,
375 Comisión Centroamericana de Ambiente y Desarrollo (CCAD) y Sistema de la
376 Integración.
- 377 Hagen I, Huggel C, Ramajo L, Chacón N, Ometto JP, Postigo JC, Castellanos EJ (2022) (2022)
378 Climate change-related risks and adaptation potential in Central and South America during
379 the 21st century. *Environmental Research Letters* 17: 033002.
- 380 Hannah L, Donatti CI, Harvey CA, Alfaro E, Rodriguez DA, Bouroncle C, Castellanos E, Diaz
381 F, Fung E, Hidalgo HG, Imbach P, Läderach P, Landrum JP, Solano AL (2017) Regional
382 modeling of climate change impacts on smallholder agriculture and ecosystems in Central
383 America. *Climatic Change* 141: 29–45. <https://doi.org/10.1007/s10584-016-1867-y>
- 384 Harvey C, Alpiñar F, Chacón M, Madrigal R (2005) Assessing linkages between agriculture and
385 biodiversity in Central America: Historical overview and future perspectives.
386 Mesoamerican & Caribbean Region, Conservation Science Program. The Nature
387 Conservancy (TNC), San José, Costa Rica: 1–162. Available from:
388 http://www.efdnitiative.org/sites/default/files/linking20agriculture20and20biodiversity20in20ca202004_0.pdf (March 23, 2022).
389
- 390 Heckadon-Moreno S (1997) Spanish rule, independence, and the modern colonization frontiers.
391 In: Coates AG (Ed.), *Central America: A Natural and Cultural History*. Yale University
392 Press, 177–274.
- 393 Heringer G, Angulo E, Ballesteros-Mejia L, Capinha C, Courchamp F, Diagne C, Duboscq-Carra
394 VG, Andrés Nuñez M, Zenni RD (2021) The economic costs of biological invasions in
395 Central and South America: a first regional assessment. *Neobiota* 67: 401–426.
396 <https://doi.org/10.3897/neobiota.67.59193i>
- 397 Hijmans RJ (2023) *_raster: Geographic Data Analysis and Modeling_*. Available from:
398 <https://cran.r-project.org/package=raster>.
- 399 Hoang NT, Kanemoto K (2021) Mapping the deforestation footprint of nations reveals growing
400 threat to tropical forests. *Nature Ecology and Evolution* 5: 845–853.
401 <https://doi.org/10.1038/s41559-021-01417-z>

- 402 IPBES (2023) Summary for Policymakers of the Thematic Assessment Report on Invasive
403 Alien Species and their Control of the Intergovernmental Science-Policy Platform on
404 Biodiversity and Ecosystem Services. In: Roy HE, Pauchard A, Stoett P, Renard Truong T,
405 Bacher S, Galil BS, Hulme PE, Ikeda T, Sankaran K V, McGeoch MA, Meyerson LA,
406 Nuñez MA, Ordonez A, Rahlao SJ, Schwindt E, Seebens H, Sheppard AW, Vandvik V
407 (Eds), IPBES Secretariat, Bonn, Germany. <https://doi.org/10.5281/zenodo.7430692>
- 408 Kerf M (2021) The ideal trilogy for better growth in Central America and the Dominican
409 Republic. Available from: [https://blogs.worldbank.org/latinamerica/ideal-trilogy-better-](https://blogs.worldbank.org/latinamerica/ideal-trilogy-better-growth-central-america-and-dominican-republic)
410 [growth-central-america-and-dominican-republic](https://blogs.worldbank.org/latinamerica/ideal-trilogy-better-growth-central-america-and-dominican-republic).
- 411 van Kleunen M, Essl F, Pergl J, Brundu G, Carboni M, Dullinger S, Early R, González-Moreno
412 P, Groom QJ, Hulme PE, Kueffer C, Kühn I, Máguas C, Maurel N, Novoa A, Parepa M,
413 Pyšek P, Seebens H, Tanner R, Touza J, Verbrugge L, Weber E, Dawson W, Kreft H,
414 Weigelt P, Winter M, Klöner G, Talluto M V., Dehnen-Schmutz K (2018) The changing
415 role of ornamental horticulture in alien plant invasions. *Biological Reviews* 93: 1421–
416 1437. <https://doi.org/10.1111/brv.12402>
- 417 Liu J, Liang SC, Liu FH, Wang RQ, Dong M (2005) Invasive alien plant species in China:
418 Regional distribution patterns. *Diversity and Distributions* 11: 341–347.
419 <https://doi.org/10.1111/j.1366-9516.2005.00162.x>
- 420 Lozon JD, MacIsaac HJ (1997) Biological invasions: Are they dependent on disturbance?
421 *Environmental Reviews* 5: 131–144. <https://doi.org/10.1139/a97-007>
- 422 MacVean AL de, Zinn H (2023) Plantas introducidas a Guatemala. In: Schuster JC, Yoshimoto J,
423 Sierra JM (Eds), *Biodiversidad de Guatemala*. Universidad del Valle de Guatemala, 368–
424 384.
- 425 Marvier M, Kareiva P, Neubert MG (2004) Habitat destruction, fragmentation, and disturbance
426 promote invasion by habitat generalists in a multispecies metapopulation. *Risk Analysis*
427 24: 869–878. <https://doi.org/10.1111/J.0272-4332.2004.00485.X>
- 428 Mayer K, Haeuser E, Dawson W, Essl F, Kreft H, Pergl J, Pyšek P, Weigelt P, Winter M,
429 Lenzner B, van Kleunen M (2017) Naturalization of ornamental plant species in public

- 430 green spaces and private gardens. *Biological Invasions* 19: 3613–3627.
431 <https://doi.org/10.1007/s10530-017-1594-y>
- 432 McDonald RI, Kareiva P, Forman RTT (2008) The implications of current and future
433 urbanization for global protected areas and biodiversity conservation. *Biological*
434 *Conservation* 141. <https://doi.org/10.1016/j.biocon.2008.04.025>
- 435 Morales-Marroquín JA, Solis Miranda R, Baldin Pinheiro J, Zucchi MI (2022) Biodiversity
436 Research in Central America: A Regional Comparison in Scientific Production Using
437 Bibliometrics and Democracy Indicators. *Frontiers in Research Metrics and Analytics* 7.
438 <https://doi.org/10.3389/FRMA.2022.898818/FULL>
- 439 Pimentel D, McNair S, Janecka J, Wightman J, Simmonds C, O'connell C, Wong E, Russel L,
440 Zern J, Aquino T, others (2001) Economic and environmental threats of alien plant,
441 animal, and microbe invasions. *Agriculture, Ecosystems & Environment* 84: 1–20.
442 Available from: <http://www.sciencedirect.com/science/article/pii/S016788090000178X>.
- 443 R Development Core Team (2022) R: A language and environment for statistical computing.
444 Available from: <http://www.rproject.org/>.
- 445 Ratcliffe H, Kendig A, Vacek S, Carlson D, Ahlering M, Dee LE (2024) Extreme precipitation
446 promotes invasion in managed grasslands. *Ecology* 105: e4190.
447 <https://doi.org/10.1002/ecy.4190>
- 448 Rojas-Sandoval J, Ferrufino-Acosta L, Flores R, Galán P, López O, MacVean AL, Rodríguez
449 Delcid D, Ruiz Y, Chacón-Madrigal E (2022) Flora introduced and naturalized in Central
450 America. *Biological Invasions*: 1–15. [https://link.springer.com/article/10.1007/s10530-](https://link.springer.com/article/10.1007/s10530-022-02968-3)
451 [022-02968-3](https://link.springer.com/article/10.1007/s10530-022-02968-3)
- 452 Rojas-Sandoval J, Ackerman JD (2021) Ornaments lead the way: global influences on plant
453 invasions in the Caribbean. *NeoBiota* 64: 177-197
454 <https://doi.org/10.3897/neobiota.64.62939>
- 455 Rojas-Sandoval J, Ackerman JD, Dueñas MA, Velez J, Díaz-Soltero, H (2024) Habitat affiliation
456 of non-native plant species across their introduced ranges on Caribbean islands. *Biological*
457 *Invasions* 26, 2237–2249. <https://link.springer.com/article/10.1007/s10530-024-03307-4>

- 458 Rousset F, Ferdy J-B (2014) Testing environmental and genetic effects in the presence of spatial
459 autocorrelation. *Ecography* 37: 781–790.
- 460 Seebens H, Kaplan E (2022) DASCO: A workflow to downscale alien species checklists using
461 occurrence records and to re-allocate species distributions across realms. *NeoBiota* 74: 75-
462 91 74: 75–91. <https://doi.org/10.3897/NEOBIOTA.74.81082>
- 463 Simberloff D (2009) The role of propagule pressure in biological invasions. *Annual Review of*
464 *Ecology, Evolution, and Systematics* 40: 81–102.
465 <https://doi.org/10.1146/ANNUREV.ECOLSYS.110308.120304>
- 466 Spear D, Foxcroft LC, Bezuidenhout H, McGeoch MA (2013) Human population density
467 explains alien species richness in protected areas. *Biological Conservation* 159: 137–147.
468 <https://doi.org/10.1016/j.biocon.2012.11.022>
- 469 Taylor MA, Alfaro EJ (2005) Central America and the Caribbean, Climate of. In: *Encyclopedia*
470 *of World Climatology*. Springer Netherlands, 183–189. [https://doi.org/10.1007/1-4020-](https://doi.org/10.1007/1-4020-3266-8_37)
471 [3266-8_37](https://doi.org/10.1007/1-4020-3266-8_37)
- 472 Torres-Miranda A, Luna-Vega I, Oyama K (2011) Conservation biogeography of red oaks
473 (*Quercus*, Section *Lobatae*) in Mexico and Central America. *American Journal of Botany*
474 98: 290–305. <https://doi.org/10.3732/ajb.1000218>
- 475 Tsakalos JL, Smith MR, Luebert F, Mucina L (2023) climenv: Download, extract and visualise
476 climatic and elevation data. *Journal of Vegetation Science* 34.
477 <https://doi.org/10.1111/jvs.13215>
- 478 World Bank (2022) Land Area (sq. km) - Latin America & Caribbean. Available from:
479 <https://data.worldbank.org/indicator/AG.LND.TOTL.K2>.
- 480 World Bank (2025a) GDP (Gross Domestic Product) data.
481 <https://doi.org/https://data.worldbank.org/indicator/NY.GDP.MKTP.CD>
- 482 World Bank (2025b) Population, total.
483 <https://doi.org/https://data.worldbank.org/indicator/SP.POP.TOTL>

484 Zimmermann H, Brandt P, Fischer J, Welk E, von Wehrden H (2014) The Human Release
485 Hypothesis for biological invasions: Human activity as a determinant of the abundance of
486 invasive plant species. F1000Research 3. <https://doi.org/10.12688/f1000research.3740.1>

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494 **Table 1.** Geographical and socioeconomic data of the Central American countries. Land area
 495 (km²), human population (people), Gross Domestic Product 2022 (GDP), number of occurrence
 496 records, and naturalized plant species.

Country	Land area (km ²)	Population	GDP (Billions US\$)	Occurrences	Species
Belize	22299.42	405272	2.8305	909	138
Costa Rica	51144.33	5180829	69.244	19179	545
El Salvador	20539.22	6336392	31.989	3261	364
Guatemala	108811.24	17357886	95.003	2904	273
Honduras	112236.73	10432860	31.426	3416	283
Nicaragua	128691.43	6948392	15.65	5373	287
Panama	74530.31	4408581	76.523	7616	286

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515 Table 2. Coefficients of the mixed-effect regression model predicting the number of occurrences
 516 and the species richness of naturalized species in Central America using as predictors human
 517 population, annual mean temperature, annual precipitation, and biodiversity integrity index (see
 518 methods).

Variable	Estimate	SE	Sq-Mean	Num df	Den df	F-value	Pr (>F)
Number of occurrences							
(Intercept)	-15.399	24.477					
Population density	0.118	0.004	210.303	1	1520.1	804.155	<0.001***
Annual mean temperature	-0.683	0.612	0.326	1	460.34	1.247	0.265
Annual precipitation	0.008	0.003	1.329	1	94.6	5.083	0.026*
Biodiversity integrity	19.281	19.479	0.256	1	973.03	0.980	0.323
Species richness							
(Intercept)	22.441	5.321					
Population density	0.019	0.001	1.0817	1	1669.8	525.337	<0.001***
Annual mean temperature	-0.509	0.132	0.0308	1	808.24	14.980	<0.001***
Annual precipitation	0.001	0.001	0.0039	1	215.58	1.913	0.16805
Biodiversity integrity	-10.065	4.023	0.0128	1	1411.5	6.258	0.01247*

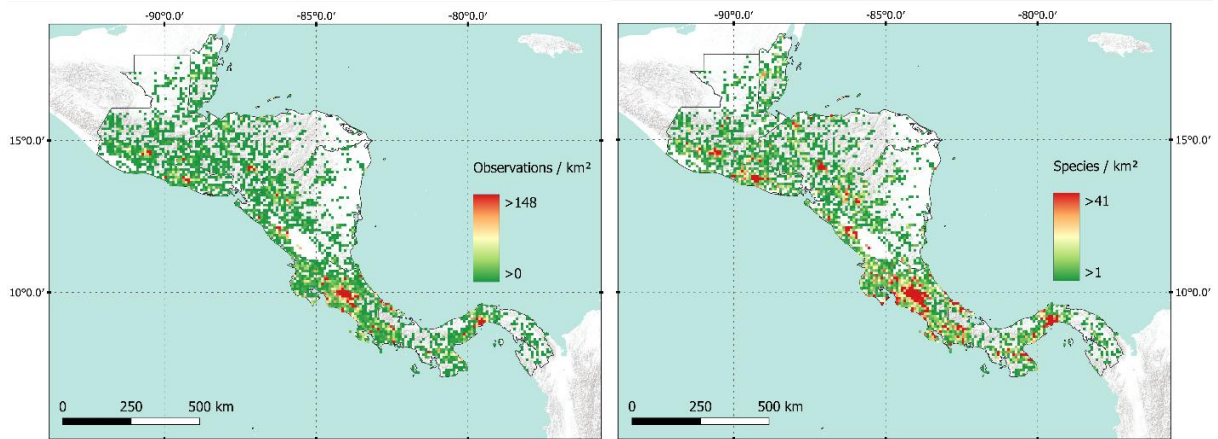
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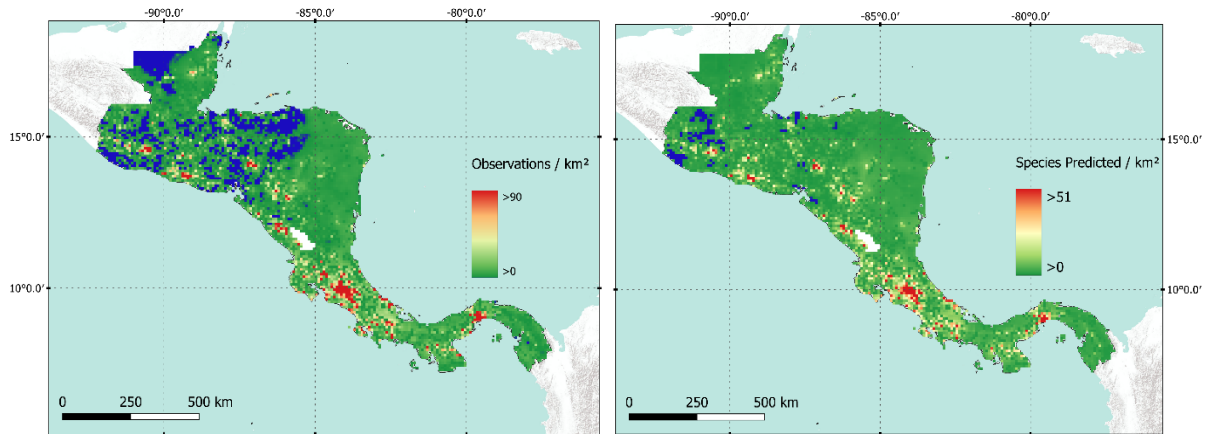
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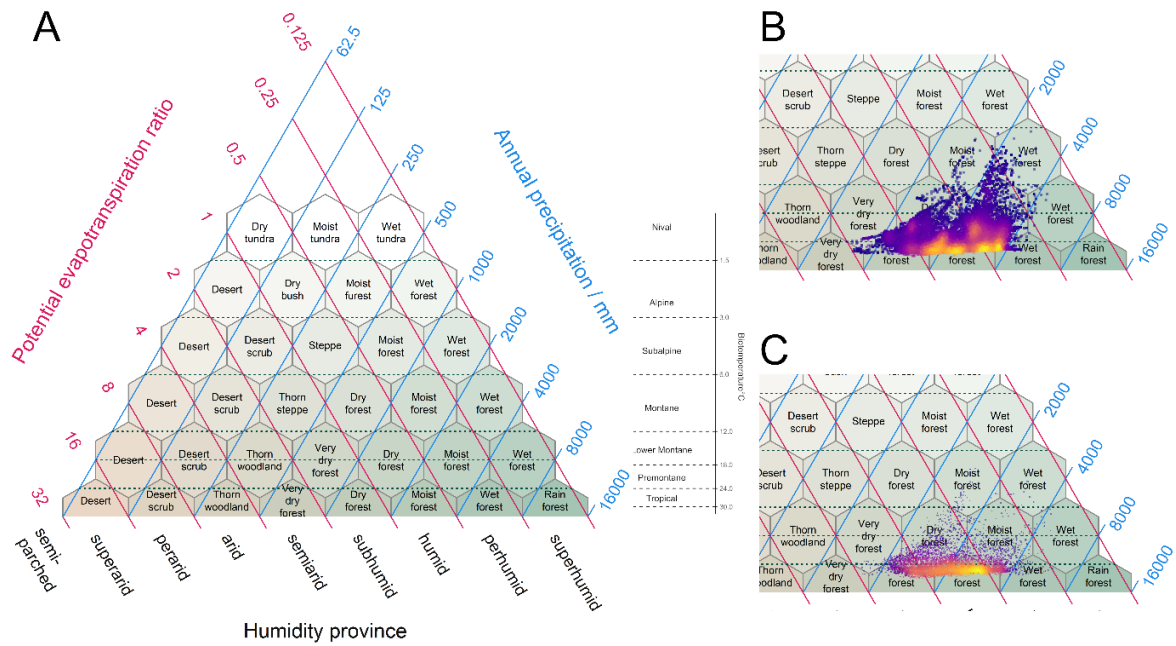
526 Figure 1. Maps of the number of occurrence records of naturalized species (above-left) and
527 species richness (above-right). Map of the number of occurrences predicted (below-left),
528 according to population density and precipitation, and map of species richness (below-right)
529 predicted according to population density, temperature, and biodiversity integrity with a fixed
530 number of 300 occurrences.

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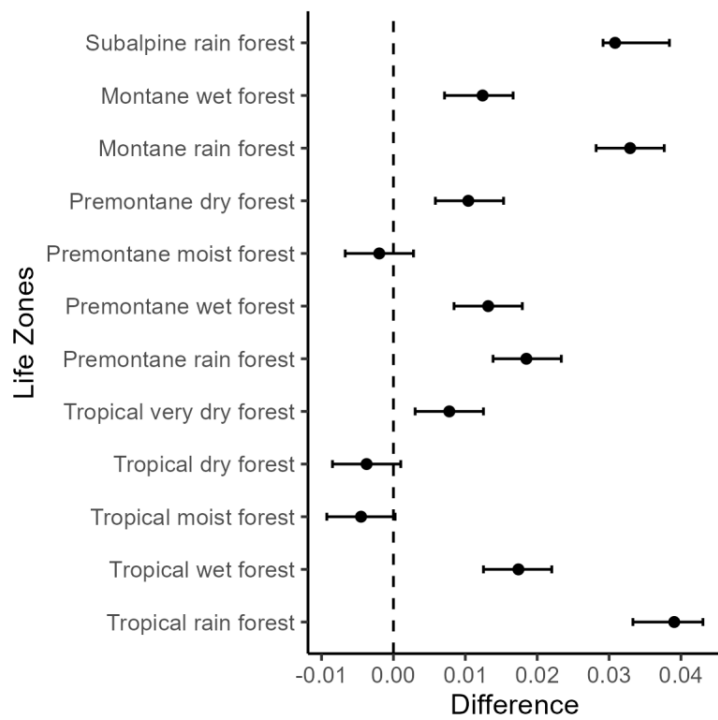
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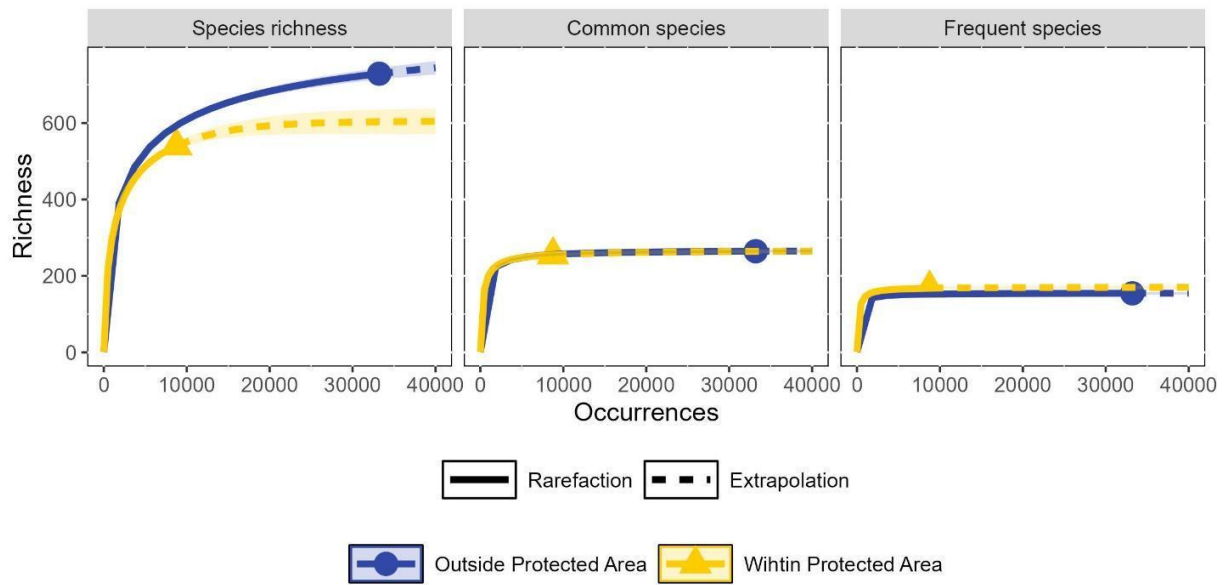
536 Fig. 2. Holdridge life zones scheme based on biotemperature, annual precipitation, and potential
 537 evapotranspiration ratio (A). (B) Environmental space is available within the Central American
 538 Region within the dimensions of the Holdridge life zone variables. (C) Environmental space
 539 where naturalized species have been recorded within the dimensions of the Holdridge life zones.
 540 The blue points represent low-density points, and the yellow points are high-density points.



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542 Fig. 3. Difference between the observed and expected proportion of cells with naturalized
 543 species across each Holdridge life zone in Central America. The observed proportion represents
 544 the number of cells per life zone, while the expected proportion was generated by randomly
 545 assigning the same number of occupied cells ($n = 14,125$) across all life zones randomized 1,000
 546 times. The error bars represent the standard error of the expected distribution.

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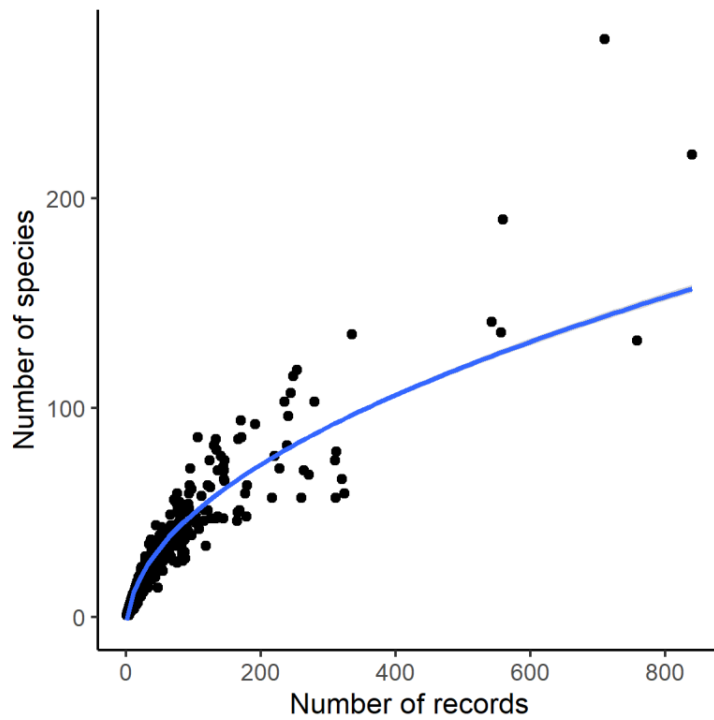
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549 Fig. 4. Rarefaction curves of the naturalized plant species in Central America occurring within
550 and outside protected areas based on Hill's numbers 0 (species richness), 1 (common species),
551 and 2 (frequent species). Curves represent the estimated diversity for each category, allowing a
552 comparison of species accumulation and community structure between protected and unprotected
553 areas.

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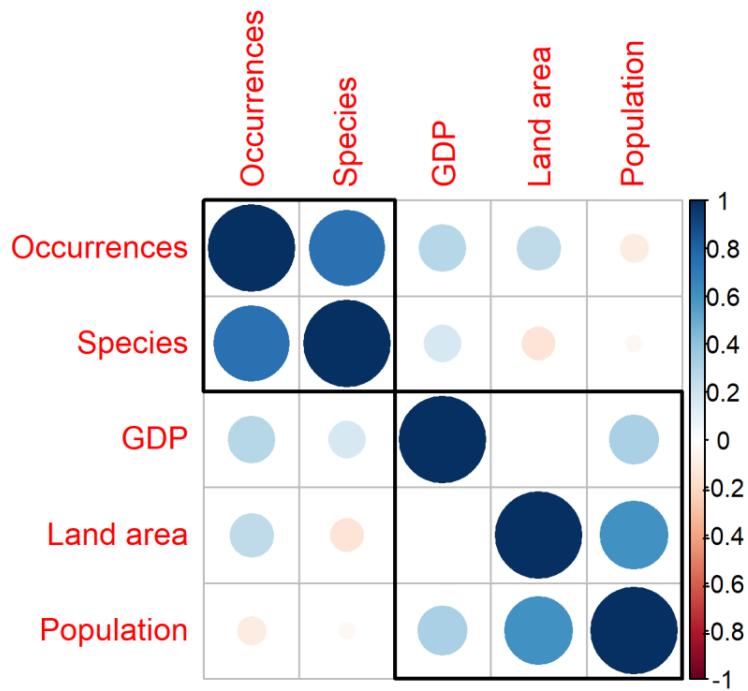
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556 Supplementary Information.



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558 SI. Fig. S1 Relationship between the number of occurrence records in the Global Biodiversity
559 Information Facility (GBIF database) and the number of naturalized plant species across Central
560 America. Regression model: $y = 11.48 \pm 0.16 + (1,377.25 \pm 8.69) x^{-233.57 \pm 8.69} \chi^2$, d.f. =
561 2868, $p < 0.001$, $r^2 = 0.9$.



562

563 SI. Fig 2. Correlation between the number of occurrence records of naturalized species and land
564 surface, human population (2022), and Gross Domestic Product (GDP, 2022) across Central
565 American countries.

566

567 SI. Table 1. Sources of layers of Protected Areas for each Central American country used in the
 568 analysis.

Country	Source
Belize	Meerman, J.C. (2019) Bm-Belize-Central America-all protected areas. Caribbean Marine Atlas. EPSG:26716. https://www.caribbeanmarineatlas.net/layers/cma_geonode_data:geonode:belize_protected_areas_all
Costa Rica	SINAC (2024) Areas silvestres protegidas. Sistema Nacional de Areas de Conservación. EPSG:5367. http://geoslpne.sirefor.go.cr/wfs
El Salvador	World Bank Data (2009) Protected natural areas in El Salvador. EPSG:32616 https://datacatalog.worldbank.org/search/dataset/0042341/Protected-Natural-Areas-in-El-Salvador
Guatemala	Cuque,D. Pérez, G. Unidad de datos e información estratégica. (2023). Mapa del sistema guatemalteco de áreas protegidas de la República de Guatemala. [mapa digital]. Guatemala. INVERSE(ESRI):103598. https://sie.url.edu.gt/capas-geograficas/
Honduras	Universidad Nacional Autónoma de Honduras (2020) Límite de áreas protegidas de acuerdo a la base de datos del Instituto de Conservación y Desarrollo Forestal. EPSG:32616. https://territoriosenriesgo.unah.edu.hn/layers/geonode:areas_protegidas
Nicaragua	UNEP-WCMC and IUCN (2025), Protected Planet: The world database on protected areas (WDPA) and world database on other effective area-based conservation measures (WD-OECM) [Online], January 2025, Cambridge, UK: UNEP-WCMC and IUCN.
Panama	Solano, M. (2022) Comprehensive Panama protected areas dataset (2022) Smithsonian Institution. https://stridata-si.opendata.arcgis.com/datasets/9c7deb2cb1e24dcc89214e6194f16fe3_0/about

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