




PREPRINT

Author-formatted, not peer-reviewed document posted on 12/05/2025

DOI: <https://doi.org/10.3897/arphapreprints.e158652>

Metabarcoding arthropods in agroecosystems in Southern Ontario, Canada

 Dirk Steinke, Kate Perez, Sean Prosser, Jayme Sones, Jireh Agda, Stephanie deWaard,  Jeremy deWaard, Evgeny Zakharov, Sujeevan Ratnasingham,  Paul Hebert, John Fryxell

Metabarcoding arthropods in agroecosystems in Southern Ontario, Canada

Dirk Steinke[‡], Kate HJ Perez[‡], Sean W.J. Prosser[‡], Jayme E Sones[‡], Jireh RA Agda[‡], Stephanie L deWaard[‡], Jeremy R deWaard[§], Evgeny V Zakharov[‡], Sujeevan Ratnasingham[‡], Paul D.N. Hebert[‡], John Fryxell[|]

[‡] Centre for Biodiversity Genomics, University of Guelph, Guelph, Canada

[§] University of Guelph, Guelph, Canada

[|] Integrative Biology, University of Guelph, Guelph, Canada

Corresponding author: Dirk Steinke (dsteinke@uoguelph.ca)

Abstract

Background

Metabarcoding can generate large numbers of georeferenced occurrence data from bulk samples at low cost. Its integration into the practice of agricultural invertebrate biomonitoring currently lacks both standard methods and example datasets that allow the identification of potential challenges and uncertainties.

New information

For this study, we gathered metabarcoding data of terrestrial arthropods from Malaise trap samples across sites in southern Ontario, spanning a gradient from high production, intensely farmed areas to alternative land use farms with varying amounts of natural restoration of marginal lands. The result is one of the largest datasets available for comparison of how agricultural practices influence arthropod biodiversity.

Keywords

farmland biodiversity, pests, pollinators, insects, COI, DNA-based identifications

Introduction

Arthropod communities are important components of farm landscapes, providing critical ecosystem services related to nutrient recycling, pollination, and biological control (de Groot et al. 2002, Tilman 1999, Robertson and Swinton 2005, Tilman 1999). Alarminglly,

recent reports suggest significant declines in their biomass, with knock-on effects across ecosystems (Hallmann et al. 2017, Vogel 2017, Lister and Garcia 2018). The expansion of intensive agriculture has been identified as a key driver of this biodiversity loss (Fuller et al. 1995, Newton 2004, Sánchez-Bayo and Wyckhuys 2019, Tilman 1999), although the extent of these losses and the causes of decline are not fully understood a because arthropods have been infrequently included in biodiversity assessments (Robertson and Swinton 2005). One of the most pressing challenges is how to rapidly gather reliable quantitative data on organismal diversity and relative abundance over time and space. Such data are essential in discriminating between leading models of community assembly and dynamics, as well as monitoring how biodiversity varies in relation to natural processes. Both observation and quantification of change in ecosystems are fundamental tools for assessing the response of species communities to environmental alterations. Past studies have typically monitored the response of a few indicator species through repeated surveys of sites to infer impacts on the entire community, such as shifts in abundance or variation in alpha and beta diversity (e.g. Cunningham et al. 2022, Da Rocha et al. 2010). Although such studies can deliver a basic understanding of biodiversity, they fall short of providing the observational data needed to manage and protect it at larger scales. For instance, in 2012, European researchers developed a generic set of farmland biodiversity indicators which capture species and habitat diversity at the farm scale (Targetti et al. 2014). Their suggested approach to determine diversity requires the identification of many species representing different trophic levels in the ecosystem. Despite the importance of such information, the model is not feasible or scalable when using conventional approaches to species identification because of high cost. Only recent advances in biodiversity genomics allow the necessary repeatable measurement of organismal diversity. In particular, metabarcoding offers a compelling advantage over traditional approaches for tracking shifts in species distributions, as it can generate large volumes of georeferenced occurrence data from bulk samples at low cost (Braukmann et al. 2019, Steinke et al. 2022, Yu et al. 2012). The integration of metabarcoding into agricultural invertebrate biomonitoring has been lauded (Hawthorne et al. 2024), yet we lack both standard methods and example datasets to help identify key uncertainties in relation to new field metrics based on DNA-based methodologies and the integration within the field of biomonitoring practice. Consequently, the main objectives of this study were to utilize metabarcoding to establish a standardized methodology for biodiversity monitoring and environmental impact assessment, and to outline a framework for early warning of insect pest outbreaks and/or decline of beneficial insect species. In addition, we applied a standardized format to facilitate the sharing of metabarcoding data in accordance with FAIR principles.

General description

Purpose: This dataset was generated as part of the Food from Thought Ecosystem Genomics Project (<https://foodfromthought.ca/research/ecosystems/genomic-indicators-of-agro-ecosystem-services>) whose overarching goal was to develop innovative methods

to link reliable field estimates of organismal abundance with evidence of biodiversity obtained through metabarcoding (this dataset), eDNA, and image analysis of bulk samples (Schneider et al. 2023, Schneider et al. 2022).

By gathering data through metabarcoding of Malaise trap samples across sites in southern Ontario that span a gradient from intense, highly productive corn-soybean farming to alternative land use farms with varying natural restoration of marginal lands, this work generated one of the deepest datasets available to examine factors influencing arthropod biodiversity in relation to farming practices, climatic variation, physical features, and landscape heterogeneity (Burgess et al. 2024, Castellanos-Labarcena et al. 2025, Gavlovski et al. 2025, MacDougall et al. 2025). Additionally, it provides a sentinel service for identifying outbreaks of agricultural pests or decline of beneficial insects long before such trends become pervasive.

Project description

Funding: Funding for research and fieldwork was provided by grants to JF and PDNH from the Canada First Research Excellence Fund to the University of Guelph's "Food From Thought" research program (Project 000054), as well as awards to PDNH from the Ontario Ministry of Economic Development, Job Creation and Trade, the Ontario Ministry of Colleges and Universities, the Canada Foundation for Innovation (MSI 42450), Genome Canada and Ontario Genomics (OGI-208), and the New Frontiers in Research Fund (NFRFT-2020-00073).

Sampling methods

Description:

Sample collection

ez-Malaise traps, Townes style, were placed in 32 farms and conservation areas in southern Ontario for three years (2018-2020) (Fig. 1). Two traps were deployed at each location during the growing season (May to October) and samples were collected on a bi-weekly basis. A total of 64 traps were deployed in 2018 and 2019, but only 54 traps could be installed in 2020 due to COVID-19 closures. Farms were categorized by management type, capturing sites with varying degrees of agricultural intensity, including conventional farms, mid-impact farms (similar to conventional farms but with a higher proportion of natural land, >20% of the farm area), ALUS farms (farms with restored habitat on marginal lands), and conservation areas.

All samples were metabarcoded 'except those compromised' in some way (e.g. reduced sampling duration, trap malfunctions). This resulted in a total of 1699 samples (699 in 2018, 595 in 2019, 405 in 2020).

DNA extraction and PCR

DNA extraction employed a membrane-based protocol (Ivanova et al. 2006) modified for bulk samples (Steinke et al. 2022). Specimens were removed from ethanol by filtration through a sterile Microfunnel 0.45 μm Supor Membrane Filter (Pall Laboratory) using a 6-Funnel Manifold (Pall Laboratory). The wet weight of each sample was then measured in grams to allow adjustment of the volume of lysis buffer to biomass. Each sample with added buffer was incubated overnight at 56°C while gently mixed on a shaker. Of the 1699 samples selected for analysis, 1346 samples were analyzed on the Ion Torrent S5 platform while the other 353 were analyzed on the Illumina NovaSeq platform. Both subsets were transferred into separate wells in 96-well microplates, with Ion Torrent bound plates containing 80 lysate samples (10 samples with 8 technical replicates), 8 technical replicates of a positive control (lysate from a bulk sample whose component specimens were individually Sanger sequenced – public BOLD dataset - dx.doi.org/10.5883/DS-RRNGS) and 8 negative controls. By comparison, plates for Illumina contained 90 lysate samples (30 samples with 3 technical replicates of each), 3 technical replicates of a positive control (AMPTk) and 3 negative controls. DNA extracts were generated using Acroprep 3.0 μm glass fiber/0.2 μm Bio-Inert membrane plates (Pall Laboratory). Each lysate was mixed with 100 μl of binding mix, transferred to a column plate, and centrifuged at 5000 g for 5 min. DNA was then purified with three washes; the first wash employed 180 μl of protein wash buffer centrifuged at 5000 g for 5 min. Each column was then washed twice with 600 μl of wash buffer centrifuged at 5000 g for 5 min. Columns were transferred to clean tubes and spun dry at 5000 g for 5 min before their transfer to clean collection tubes followed by incubation for 30 min at 56°C to dry the membrane. DNA was eluted by adding 60 μl of 10 mM Tris-HCl pH 8.0 followed by centrifugation at 5000 g for 5 min.

PCR reactions employed a standard protocol (Braukmann et al. 2019). Briefly, each reaction included 5% trehalose (Fluka Analytical), 1 \times Platinum Taq reaction buffer (Invitrogen), 2.5 mM MgCl_2 (Invitrogen), 0.1 μM of each primer (Integrated DNA Technologies), 50 μM of each dNTP (KAPA Biosystems), 0.3 units of Platinum Taq (Invitrogen), 2 μl of DNA extract, and Hyclone ultra-pure water (Thermo Scientific) for a final volume of 12.5 μl . Two-stage PCR was used to generate amplicon libraries for sequencing on Ion Torrent S5. The first round of PCR used the primer combination AncientLepF3 (Prosser et al. 2016) and LepR1 (Hebert et al. 2004) to amplify a 463 bp fragment of COI. Prior to the second PCR, first round products were diluted 2x with ddH₂O. Fusion primers were then used to attach platform-specific unique molecular identifiers (UMIs) along with the sequencing adaptors required for Ion Torrent S5 libraries. Both rounds of PCR employed the same thermocycling conditions: initial denaturation at 94 °C for 2 min, followed by 20 cycles of denaturation at 94°C for 40 sec, annealing at 51°C for 1 min, and extension at 72 °C for 1 min, with a final extension at 72°C of 5 min. For samples prepared for Illumina sequencing, a fusion primer based two-step PCR protocol was employed that amplifies target fragments in the first step and attaches in-line tags and Illumina TruSeq library sequence tails during the second PCR (Elbrecht and Steinke 2018). This was done using in-line tags of different lengths and sequenced amplicon pools in mixed orientation. For the first PCR step, we used a protocol similar to that described above, only with a fixed annealing temperature of 46°C for each primer pair

(BF3/BR2 – Elbrecht et al. 2019) and 24 cycles. We used 1 μ L PCR product of each primer set as template for the second PCR step (with no quantification or reaction cleanup) under similar PCR conditions except we increased the extension time to 2 minutes and reduced the number of cycles to 14. PCR products were cleaned using SPRIselect (Beckman Coulter, CA, USA) with a sample to volume ratio of 0.76x. DNA concentration was quantified using a Qubit fluorometer, High Sensitivity dsDNA Kit (Thermo Fisher Scientific, MA, USA).

Sequencing library construction

For each plate, labelled amplicons were pooled prior to sequencing. In total, 135 libraries were assembled. Samples, along with positive and negative controls, were pooled after UMI tagging to create a library that was analyzed on a 530 chip (35 chips in total). Amplicon libraries were prepared on an Ion Chef (Thermo Fisher Scientific) and sequenced on an Ion Torrent S5 platform at the Centre for Biodiversity Genomics following manufacturer's instructions (Thermo Fisher Scientific). The single Illumina library was sequenced on one lane of an Illumina NovaSeq SP chip at the Sick Kids Hospital Sequencing Centre in Toronto.

Data analysis

Reads were uploaded to mBRAVE (<http://mbrave.net/>) for quality filtering and subsequent queries using several reference libraries in an open reference approach. Reads were mapped against a Canadian Reference library (unpublished data). Reads were also queried against five system libraries on mBRAVE: bacteria (SYS-CRLBACTERIA) to screen for potential contamination, e.g., by endosymbionts such as *Wolbachia*; chordates (SYS-CRLCHORDATA); insects (SYS-CRLINSECTA); non-insect arthropods (SYS-CRLNONINSECTARTH); non-arthropod invertebrates (SYS-CRLNONARTHINVERT). All non-arthropod reads were discarded from further analysis. Sequences were only included in this analysis if they met a minimum length of >350 bp and the following three quality criteria: mean QV >20; <25% positions with a QV <20; <5% positions with QV <10. Reads were trimmed 30 bp from their 5' terminus with a set trim length filter of 450 bp. Reads were matched to sequences in each reference library with an ID distance threshold of 3% but were only retained for further analysis if at least five reads matched a BIN in the reference database. This number is based on earlier benchmarking of the assignment algorithm on mBRAVE, where IonTorrent generated sequences provided the best compromise between removing error and retaining real matches (Steinke et al. 2022). All reads failing to match any sequence in the five reference libraries were clustered at an OTU threshold of 1% with a minimum of five reads per cluster, again a value based on initial benchmarking.

Using mBRAVE, we generated BIN (Ratnasingham and Hebert 2013) tables, including all library queries for each individual plate/run. Read counts for any BINs recovered from the negative control on a plate were subtracted from the counts for the same BIN in the non-control wells in the run. When this reduced the read count for a BIN to zero, its

occurrence was removed. This step aimed to reduce the effects of rare tag switching (Elbrecht and Steinke 2018) as well as background contamination.

Datasets downloaded from mBRAVE were converted into OTU tables and presence/absence matrices for further analysis using a R script. To determine the completeness of sampling, we calculated rarefaction curves and Hill numbers (Chao et al. 2014) using the *iNEXT* package (Hsieh et al. 2016). All analyses were performed for both the entire dataset and management type subsets. Chord diagrams visualizing overlap between management types were generated using the *circlize* package (Gu et al. 2014). Since the overall dataset was skewed towards ALUS farms, this calculation was done by using a random selection of six ALUS sites. Treemaps of taxonomic distribution were generated using the *treemap* package (Tennekes 2017). In addition, the dataset was screened against Canadian Libraries of Pest and Pollinator species assembled in Padhye et al (in prep). All analyses were performed in R v.4.1.1 (R Core Team 2021).

Geographic coverage

Description: The study was carried out at farms and conservation areas in Southern Ontario (Fig. 1, Suppl. material 1)

Coordinates: 42.2606 and 43.7785 Latitude; -83.0684 and -80.1416 Longitude.

Taxonomic coverage

Description: The metabarcoding analysis generated a total of 1,449,919 occurrence records. The overall dataset included 28,667 BINs belonging to 45 arthropod orders (Fig. 2). Among them, Diptera (43%), Hymenoptera (25%), Lepidoptera (9%), Coleoptera (7%) and Hemiptera (6%) represented the highest percentages of the BINs, while the remaining 40 orders represented less than 3% each. The taxonomic composition did not vary significantly over the three years (Fig. 2). The greatest number of BINs were found in ALUS farms, followed by conservation areas, conventional farms, and mid-impact management (Table 1). Extrapolations based on Hill numbers (Chao et al. 2014) suggest that nother 3,000–4,000 BINs await detection (Fig. 3). Suppl. material 1 provides BIN counts per individual farm along with counts per farm for pollinator and registered pest species. Overall, a total of 418 pest species (Suppl. material 2) and 2,692 pollinator species (Suppl. material 3) were detected.

Temporal coverage

Notes: Malaise trap samples were collected from May–October in 2018, 2019, and 2020.

Usage licence

Usage licence: Creative Commons Public Domain Waiver (CC-Zero)

Data resources

Data package title: Agricultural Monitoring 2018-2020

Number of data sets: 4

Data set name: Arthropod monitoring at ALUS Farms 2018

Download URL: <https://www.ncbi.nlm.nih.gov/bioproject/?term=PRJNA877241>

Data format: Genomic Standard Consortium

Description: DNA sequence data have been deposited on NCBI SRA under accession number PRJNA877241.

Data set name: Arthropod monitoring at ALUS Farms 2019

Download URL: <https://www.ncbi.nlm.nih.gov/bioproject/?term=PRJNA856887>

Data format: Genomic Standard Consortium

Description: DNA sequence data have been deposited on NCBI SRA under accession number PRJNA856887.

Data set name: Arthropod monitoring at ALUS Farms 2020

Download URL: <https://www.ncbi.nlm.nih.gov/bioproject/?term=PRJNA873715>

Data format: Genomic Standard Consortium

Description: DNA sequence data have been deposited on NCBI SRA under accession number PRJNA873715.

Data set name: Arthropod monitoring at ALUS Farms 2018-2020

Download URL: [10.5281/zenodo.15345117](https://doi.org/10.5281/zenodo.15345117)

Data format: Darwin Core Archive

Description: The dataset representing DNA-based occurrences available as an occurrence dataset with the DNA-derived extension table based on GBIF recommendations (Abarenkov et al. 2023).

Column label	Column description
basisOfRecord	Nature of data record - "MaterialSample"
occurrenceID (Occurrence core)	A unique identifier for the occurrence.
eventID (Occurrence core)	An identifier for the set of information associated with an Event (sample number).
eventDate (Occurrence core)	Date when the sample was retrieved from trap.
recordedBy (Occurrence core)	Organization or persons responsible for recording occurrence "Centre for Biodiversity Genomics"
organismQuantity (Occurrence core)	Number of reads of this OTU in this sample.
organismQuantityType (Occurrence core)	"DNA sequence reads".
sampleSizeValue (Occurrence core)	Total number of reads in sample
sampleSizeUnit (Occurrence core)	"DNA sequence reads"
materialSampleID (Occurrence core)	Biosample ID obtained from NCBI SRA
samplingProtocol (Occurrence core)	Sampling method "Malaise Trap"
decimalLatitude (Occurrence core)	The geographic latitude where the dwc:Event occurred (exact locality of the sample collection).
decimalLongitude (Occurrence core)	The geographic longitude where the dwc:Event occurred (exact locality of the sample collection).
country (Occurrence core)	A name of the country where the sampling occurred ("Canada").
stateProvince (Occurrence core)	A name of the province where the sampling occurred ("Ontario").
locationID (Occurrence core)	Code for a specific location
geodeticDatum (Occurrence core)	The geodetic datum ("WGS84").
scientificName (Occurrence core)	identifier from BOLD (BIN)

kingdom (Occurrence core)	The scientific name of the kingdom in which the BIN is classified
phylum (Occurrence core)	The scientific name of the phylum in which the BIN is classified.
class (Occurrence core)	The scientific name of the class in which the BIN is classified.
order (Occurrence core)	The scientific name of the order in which the BIN is classified.
family (Occurrence core)	The scientific name of the family in which the BIN is classified.
subfamily (Occurrence core)	The scientific name of the subfamily in which the BIN is classified.
genus (Occurrence core)	The scientific name of the genus in which the BIN is classified.
verbatimIdentification (Occurrence core)	The taxonomic identification as it appeared in the original record in which the BIN was classified
habitat (Occurrence core)	A category or description of the habitat in which the dwc:Event occurred ("ALUS" - Alternative Landuse Farming, "CONV" - Conventional Farming, "CONS" - Conservation Area, "MID" - mid-impact Farming).
ID (DNA-derived extension)	A unique identifier for the occurrence refers to the occurrence table (occurrenceID).
sop (DNA-derived extension)	Standard operating procedures used in assembly and/or taxonomic annotation of reads.
target_gene (DNA-derived extension)	Targeted gene or marker name for marker-based studies (COI).
target_subfragment (DNA-derived extension)	Name of subfragment of a gene (COI-barcode region).
pcr_primer_forward (DNA-derived extension)	Forward PCR primer ("TTATAATTGGDGGWTTTGGWAATTG", "CCHGAYATRGCHTTYCCHCG").
pcr_primer_reverse (DNA-derived extension)	Reverse PCR primer ("TAAACTTCTGGATGTCCAAAAATCA", "TCDGGRTGNCCRAARAAYCA").
pcr_primer_name_forward (DNA-derived extension)	Name of the forward PCR primer ("AncientLepF3", "BF3").
pcr_primer_name_reverse (DNA-derived extension)	Name of the reverse PCR primer ("LepR1", "BR2").
pcr_primer_reference (DNA-derived extension)	DOI Reference for the primers (https://doi.org/10.1093/gigascience/giac040 , https://doi.org/10.7717/peerj.7745).
lib_layout (DNA-derived extension)	The configuration of reads ("single", "paired").
seq_meth (DNA-derived extension)	Sequencing method used ("Ion Torrent", "Illumina NextSeq").
otu_class_appr (DNA-derived extension)	Approach/algorithm and clustering level ("mBRAVE").

otu_seq_comp_appr (DNA-derived extension)	Tool and thresholds used to assign "species-level" names to OTUs ("mBRAVE").
otu_db (DNA-derived extension)	Reference database: Canadian Reference library (Pentinsaari et al.) "DS-CANREF22"

Additional information

Sequence analysis of the 1,699 samples produced 2,380,695,937 reads across 142 S5 runs (mean reads per run = 13.4 million) and one NovaSeq SP lane (423,128,504 reads). Over two thirds of these reads were filtered, leaving 809,436,619 reads that could be assigned to a BIN. Nearly all reads (99.5%) found a BIN match on BOLD. Those that failed to do so were *de novo* clustered using mBRAVE with a 99% similarity threshold. The latter analysis recognized an average of 12 additional OTUs per sample, but >98% reflected sequencing/PCR errors (e.g., chimeras, sequences with multiple indels) or NUMTs so they were excluded from the data set.

Acknowledgements

We thank all staff and students who helped to deploy and maintain traps for this project, particularly: Gustavo Betini, Marie Gutgesell, Connor Warne, and Patrick Burgess. Special thanks to the participants and members of the ALUS Norfolk and Elgin, as well as Ojibway Provincial Park, Hawk Cliff Woods, Backus Heritage, Galt Gardens, rare Charitable Research Reserve, and Watson Pond.

Author contributions

Conceptualization: JF, JRD, DS. Data curation: KHJP, JES, JRAA, SR. Formal analysis: DS, KHJP, SP. Funding acquisition: JF, PDNH. Investigation: SLD, SWJP, EVZ. Project administration: JF, DS, JRD. Supervision: JF, DS, JES. Validation: DS, KHJP. Visualization: DS. Writing - original draft: DS. Writing - review and editing: KHJP, JES, JRAA, PDNH, EVZ, SR.

References

- Abarenkov K, Andersson AF, Bissett A, Finstad AG, Fossøy F, Grosjean M, Hope M, Jeppesen TS, Kõljalg U, Lundin D, Nilsson RN, Prager M, Provoost D, Schigel D, Suominen S, Svenningsen C, Frøslev TG (2023) Publishing DNA-derived data through biodiversity data platforms, v1.3. GBIF Secretariat <https://doi.org/10.35035/doc-vf1a-nr22>
- Braukmann T, Prosser S, Ivanova N, Elbrecht V, Steinke D, Ratnasingham R, deWaard J, Sones J, Zakharov E, Hebert P (2019) Metabarcoding a diverse arthropod mock community. *Molecular Ecology Resources* 19: 711-727.

- Burgess P, Betini GS, Cholewka A, deWaard JR, deWaard S, Griswold C, Hebert PD, MacDougall A, McCann KS, McGroarty J, Miller E, Perez K, Ratnasingham S, Reisinger C, Steinke D, Wright E, Zakharov E, Fryxell JM (2024) Spatial and seasonal determinants of arthropod community composition across an agro-ecosystem landscape. *FACETS* 9: 1-15. <https://doi.org/10.1139/facets-2023-0051>
- Castellanos-Labarcena J, Adamowicz SJ, Hanner R, Steinke D (2025) Insect richness is promoted by complex and heterogeneous habitats in agroecosystem landscapes. *Ecological Indicators*. In review.
- Chao A, Gotelli N, Hsieh TC, Sander E, Ma KH, Colwell R, Ellison A (2014) Rarefaction and extrapolation with Hill numbers: a framework for sampling and estimation in species diversity studies. *Ecological Monographs* 84 (1): 45-67. <https://doi.org/10.1890/13-0133.1>
- Cunningham M, Tran L, McKee C, Ortega Polo R, Newman T, Lansing L, Griffiths J, Bilodeau G, Rott M, Marta Guarna M (2022) Honey bees as biomonitors of environmental contaminants, pathogens, and climate change. *Ecological Indicators* 134 <https://doi.org/10.1016/j.ecolind.2021.108457>
- Da Rocha JRM, De Almeida JR, Lins GA, Durval A (2010) Insects as indicators of environmental changing and pollution: A review of appropriate species and their monitoring. *Holos Environment* 10 (2). <https://doi.org/10.14295/holos.v10i2.2996>
- de Groot RS, Wilson MA, Boumans RJ (2002) A typology for the classification, description and valuation of ecosystem functions, goods and services. *Ecological Economics* 41 (3): 393-408. [https://doi.org/10.1016/s0921-8009\(02\)00089-7](https://doi.org/10.1016/s0921-8009(02)00089-7)
- Elbrecht V, Steinke D (2018) Scaling up DNA metabarcoding for freshwater macrozoobenthos monitoring. *Freshwater Biology* <https://doi.org/10.7287/peerj.preprints.3456v4>
- Elbrecht V, Braukmann TWA, Ivanova NV, Prosser SWJ, Hajibabaei M, Wright M, Zakharov EV, Hebert PDN, Steinke D (2019) Validation of COI metabarcoding primers for terrestrial arthropods. *PeerJ* 7: e7745. <https://doi.org/10.7717/peerj.7745>
- Fuller RJ, Gregory RD, Gibbons DW, Marchant JH, Wilson JD, Baillie SR, Carter N (1995) Population declines and range contractions among lowland farmland birds in Britain. *Conservation Biology* 9 (6): 1425-1441. <https://doi.org/10.1046/j.1523-1739.1995.09061425.x>
- Gavlovski A, Fryxell J, Steinke D, deWaard J (2025) Sampling requirements for standardized insect biodiversity monitoring vary with abundance. *FACETS*: In review.
- Gu Z, Gu L, Eils R, Schlesner M, Brors B (2014) *circIize* implements and enhances circular visualization in R. *Bioinformatics* 30 (19): 2811-2812. <https://doi.org/10.1093/bioinformatics/btu393>
- Hallmann CA, Sorg M, Jongejans E, Siepel H, Hofland N, Schwan H, Stenmans W, Müller A, Sumser H, Hörren T, Goulson D, de Kroon H (2017) More than 75 percent decline over 27 years in total flying insect biomass in protected areas. *PLoS one* 12 (10): e0185809. <https://doi.org/10.1371/journal.pone.0185809>
- Hawthorne BJ, Cuff J, Collins L, Evans D (2024) Metabarcoding advances agricultural invertebrate biomonitoring by enhancing resolution, increasing throughput and facilitating network inference. *Agricultural and Forest Entomology* 27 (1): 50-66. <https://doi.org/10.1111/afe.12628>
- Hebert PN, Penton E, Burns J, Janzen D, Hallwachs W (2004) Ten species in one: DNA barcoding reveals cryptic species in the neotropical skipper butterfly *Astraptes fulgerator*.

- Proceedings of the National Academy of Sciences 101 (41): 14812-14817. <https://doi.org/10.1073/pnas.0406166101>
- Hsieh TC, Ma KH, Chao A (2016) iNEXT: an R package for rarefaction and extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution* 7 (12): 1451-1456. <https://doi.org/10.1111/2041-210x.12613>
 - Ivanova N, deWaard J, Hebert PN (2006) An inexpensive, automation-friendly protocol for recovering high-quality DNA. *Molecular Ecology Notes* 6 (4): 998-1002. <https://doi.org/10.1111/j.1471-8286.2006.01428.x>
 - Lister B, Garcia A (2018) Climate-driven declines in arthropod abundance restructure a rainforest food web. *Proceedings of the National Academy of Sciences* 115 (44). <https://doi.org/10.1073/pnas.1722477115>
 - MacDougall A, Esch E, Dolzetal A, Kamm C, Carroll O, Tosi M, K M, Nessel M, Wilcox A, Ellsworth L, Mazzorato A, Noble D, Pavusa M, Ramirez S, Arce B, Gutgesell M, McCann K, Fraser E, Fryxell J, Gilvesy B, Balpataky K, Levinson J, Biswas A, Dunfield K, Rooney N, Maharali H, Newmann A, Husband B, Steinke D, deWaard J, Ali G, Prosser R, Young A, Earl H, Sulik J, Harvey E, Campbell M (2025) Ecosystem services on retired marginal farmland. *Frontiers in Ecology and the Environment* in press.
 - Newton I (2004) The recent declines of farmland bird populations in Britain: an appraisal of causal factors and conservation actions. *Ibis* 146 (4): 579-600. <https://doi.org/10.1111/j.1474-919x.2004.00375.x>
 - Prosser SJ, deWaard J, Miller S, Hebert PN (2016) DNA barcodes from century-old type specimens using next-generation sequencing. *Molecular Ecology Resources* 16 (2): 487-497. <https://doi.org/10.1111/1755-0998.12474>
 - Ratnasingham S, Hebert PN (2013) A DNA-based registry for all animal species: The Barcode Index Number (BIN) System. *PLoS ONE* 8 (7). <https://doi.org/10.1371/journal.pone.0066213>
 - R Core Team (2021) R: A language and environment for statistical computing. , Vienna, Austria. URL . 4.4.2. R Foundation for Statistical Computing, Vienna, Austria. URL: <https://www.R-project.org/>
 - Robertson GP, Swinton S (2005) Reconciling agricultural productivity and environmental integrity: A grand challenge for agriculture. *Frontiers in Ecology and the Environment* 3 (1). <https://doi.org/10.2307/3868443>
 - Sánchez-Bayo F, Wyckhuys KG (2019) Worldwide decline of the entomofauna: A review of its drivers. *Biological Conservation* 232: 8-27. <https://doi.org/10.1016/j.biocon.2019.01.020>
 - Schneider S, Taylor G, Kremer S, Burgess P, McGroarty J, Mitsui K, Zhuang A, deWaard J, Fryxell J (2022) Bulk arthropod abundance, biomass and diversity estimation using deep learning for computer vision. *Methods on Ecology and Evolution* 13 (2): 346-357. <https://doi.org/10.1111/2041-210X.13769>
 - Schneider S, Taylor GW, Kremer SC, Fryxell JM (2023) Getting the bugs out of AI: Advancing ecological research on arthropods through computer vision. *Ecology letters* 26 (7): 1247-1258. <https://doi.org/10.1111/ele.14239>
 - Steinke D, deWaard SL, Sones JE, Ivanova NV, Prosser SWJ, Perez K, Braukmann TWA, Milton M, Zakharov EV, deWaard JR, Ratnasingham S, Hebert PDN (2022) Message in a bottle—Metabarcoding enables biodiversity comparisons across ecoregions. *GigaScience* 11 <https://doi.org/10.1093/gigascience/giac040>

- Targetti S, Herzog F, Geijendorffer IR, Wolfrum S, Arndorfer M, Balázs K, Choisis JP, Dennis P, Eiter S, Fjellstad W, Friedel JK, Jeanneret P, Jongman RH, Kainz M, Luescher G, Moreno G, Zanetti T, Sarthou JP, Stoyanova S, Wiley D, Paoletti MG, Viaggi D (2014) Estimating the cost of different strategies for measuring farmland biodiversity: Evidence from a Europe-wide field evaluation. *Ecological Indicators* 45: 434-443. <https://doi.org/10.1016/j.ecolind.2014.04.050>
- Tennekes M (2017) treemap: Treemap visualization. R package version 2.2-4.
- Tilman D (1999) Global environmental impacts of agricultural expansion: The need for sustainable and efficient practices. *Proceedings of the National Academy of Sciences* 96 (11): 5995-6000. <https://doi.org/10.1073/pnas.96.11.5995>
- Vogel G (2017) Where have all the arthropods gone? *Science* 356: 576-579.
- Yu D, Ji Y, Emerson B, Wang X, Ye C, Yang C, Ding Z (2012) Biodiversity soup: metabarcoding of arthropods for rapid biodiversity assessment and biomonitoring. *Methods in Ecology and Evolution* 3 (4): 613-623. <https://doi.org/10.1111/j.2041-210x.2012.00198.x>

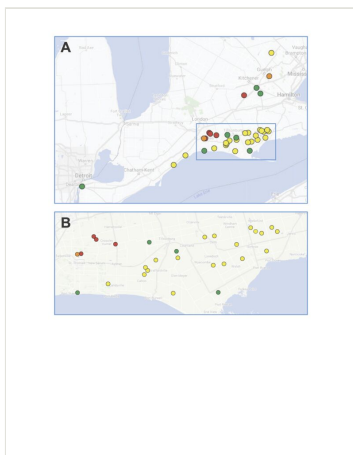


Figure 1.

Sampling locations. What is the difference between A and B- blow up? Colours indicate farm management type (yellow-ALUS farm, red-conventional farm, orange-mid-impact farm, green-conservation area)

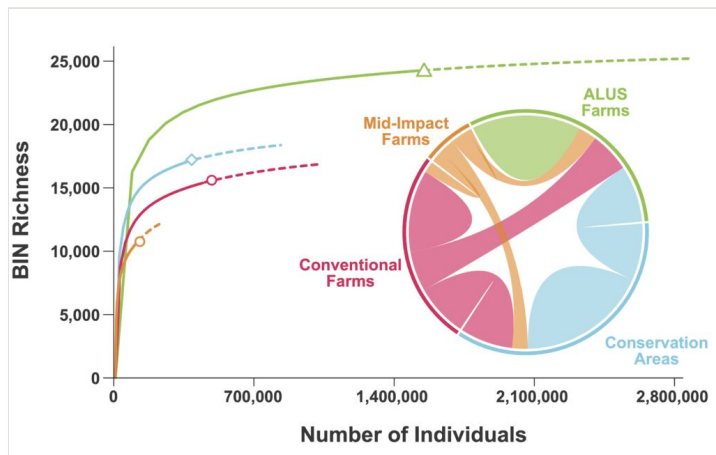


Figure 3.

BIN accumulation curves by management type for 1699 metabarcoded samples collected at 32 farms from 2018-2020. The chord diagram shows BIN overlap between management types using a random selection of 6 ALUS sites.

Table 1.

BIN counts per management type

	2018	2019	2020	Total
ALUS Farm	21,242	18,709	16,550	25,710
Conservation Area	13,828	12,876	7,191	18,225
Conventional Farm	12,332	10,801	9,266	16,163
Mid-impact farm	6,648	6,706	6,213	10,849
Total	23,457	21,807	19,218	28,667

Supplementary materials

Suppl. material 1: Overview trap locations and BIN counts

Authors: Dirk Steinke

Data type: csv

Brief description: BIN, pest and pollinator counts for each trap per year

[Download file](#) (6.16 kb)

Suppl. material 2: Pest Species

Authors: Dirk Steinke

Data type: csv

Brief description: Pest species detected with BIN and full taxonomy as well as feeding guild

[Download file](#) (48.32 kb)

Suppl. material 3: Pollinator Species

Authors: Dirk Steinke

Data type: csv

Brief description: Pollinator species detected with BIN and full taxonomy as well as feeding guild

[Download file](#) (276.35 kb)