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# **Feeding ecology of the invasive *Cynoscion regalis* on the Portuguese coast: Insights from a molecular approach**

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1 Feeding ecology of the invasive *Cynoscion regalis* on the Portuguese coast: Insights from  
2 a molecular approach

3

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17

## 18 **Abstract**

19

20 Invasive species can profoundly disrupt native ecosystems, as their self-sustaining  
21 populations alter niche occupation, increase competition, and exert predation pressure on  
22 indigenous species. Predation by invasive species, particularly on vulnerable life stages  
23 such as larvae and juveniles, can lead to population declines, loss of genetic diversity,  
24 disruption of food webs, extinction of keystone species, impaired ecosystem services, and  
25 delayed recovery of indigenous populations. *Cynoscion regalis* (Bloch and Schneider,  
26 1801), commonly known as the weakfish, is a sciaenid species native to the eastern coast  
27 of North America, where it acts as a generalist predator. Since 2009, its presence has also  
28 been documented in European waters, prompting concerns about its potential impact on  
29 local fish communities. This study presents the first molecular analysis of the feeding  
30 ecology of *C. regalis* outside its native range. Using molecular barcoding of prey items  
31 coupled with metabarcoding of gastric fluids, this research provides insights into the  
32 weakfish's diet in Portuguese waters (southwestern Europe) beyond those obtained in  
33 studies relying solely on visual stomach content identification. This underscores the value  
34 of molecular techniques in dietary analysis. Findings reveal that *C. regalis* exhibits

35 cannibalistic tendencies, with conspecifics forming a substantial part of its diet. Other  
36 dietary species include *Atherina presbyter*, *Diplodus annularis*, *Dicentrarchus punctatus*,  
37 *Engraulis encrasicolus*, *Sardina pilchardus*, and *SpondylIOSoma cantharus*. In addition,  
38 other prey were identified at the family level as Sparidae and Scorpaenidae. The presence  
39 of these species, many of which are common in estuarine habitats, particularly during  
40 juvenile stages, supports the establishment of weakfish populations in nursery areas,  
41 potentially resulting in adverse ecological impacts. This study contributes to  
42 understanding the dynamics of invasive species, highlighting the ecological interactions  
43 and disturbances they may cause. Effective monitoring and management of weakfish  
44 populations outside their native range are critical to mitigating potential ecological  
45 consequences.

46

47 **Keywords:** barcoding of stomach contents, invasive species, metabarcoding, weakfish

48

## 49 **Introduction**

50

51 Non-native species can have profound and far-reaching impacts on local ecosystems by  
52 forming stable populations and exerting pressure on native species through ecosystem  
53 disturbance. Recent studies predict a global increase of invasive species in the next  
54 decades, leading to an inevitable cascade of impacts like native species richness reduction  
55 and native population declines, ecosystem changes, and even consequences on human  
56 health, among several other effects (e.g., Seebens et al. 2020; IPBES 2023).

57

58 The weakfish, *Cynoscion regalis* (Bloch and Schneider, 1801), is a marine fish species  
59 believed to have been introduced to European waters from the Americas via ballast water  
60 (Morais and Teodósio 2016; Bañón et al. 2017a, 2017b). Its native distribution spans the  
61 western Atlantic, from southern Florida to Nova Scotia (Froese and Pauly 2025). The first  
62 documented occurrence in Europe was in 2009, with the capture of a juvenile in Belgium's  
63 Scheldt estuary (ANB 2011). Subsequent records followed in Spain's Gulf of Cadiz in  
64 2011 and Portugal's Sado, Tagus, and Mira estuaries between 2013 and 2016 (Morais and  
65 Teodósio 2016; Bañón et al. 2017a; Morais et al. 2017). The increasing presence of *C.*  
66 *regalis* in European waters is corroborated by fishing data, which report landings of  
67 hundreds of kilograms in recent years (Bañón et al. 2017a, 2017b). Once considered an  
68 incidental bycatch, the species has gained commercial interest and targeted fishing has

69 been suggested as a dual-purpose strategy, serving both as a population control method  
70 and as a means of developing a new fishery resource (Bañón et al. 2017a, 2017b; Béarez  
71 et al. 2016).

72

73 Concerning its ecology, *Cynoscion regalis* is a gregarious demersal fish species that  
74 inhabits coastal and estuarine ecosystems, typically preferring depths of 10-26 meters  
75 over sandy or muddy substrates (Froese and Pauly 2025). Known as a generalist predator,  
76 its diet includes fish, molluscs, and crustaceans, showcasing its adaptability to diverse  
77 food sources (e.g., Wilk 1979; Schultz 2004). A study examining the feeding strategy of  
78 *C. regalis* outside its native range through stomach content analysis conducted in  
79 Portugal's Sado Estuary (Cerveira et al. 2021) confirmed that the species retains its  
80 generalist feeding behaviour. It also revealed a significant trophic overlap with the meagre  
81 (*Argyrosomus regius*), suggesting potential competition for food and habitat resources,  
82 which may adversely affect native fish populations.

83

84 This study presents the first molecular analysis of the feeding ecology of *C. regalis* in its  
85 non-native range, focusing on its fish prey in Portuguese waters. Using barcoding of prey  
86 items and metabarcoding of gastric fluids, our research provides a more detailed  
87 perspective of the weakfish's diet compared to visual stomach content analyses. This  
88 approach highlights the utility of molecular techniques in dietary studies, offering insights  
89 into the ecological interactions and potential disturbances caused by an invasive species  
90 in the native fish populations in the same area. By improving our understanding of the  
91 feeding ecology and potential ecological impact of *C. regalis*, this study underscores the  
92 importance of effective monitoring and management strategies of non-native species to  
93 mitigate potential ecological consequences in recipient ecosystems.

94

## 95 **Materials and methods**

96

### 97 *Sample collection*

98

99 Thirty individuals of *C. regalis* from the Tagus estuary were collected, labelled, measured  
100 (total length), weighed (total weight) and carefully dissected to extract the digestive tracts,  
101 which were examined for stomach content analysis. Semi-digested fish prey items were  
102 selected for molecular identification via molecular barcoding, and small muscle tissue

103 samples were carefully extracted, preserved in 96% ethanol and stored in individually  
 104 labelled Eppendorf tubes. The stomachs and their contents were also preserved in 96%  
 105 ethanol to facilitate metabarcoding analysis of gastric fluids. Vouchers for all specimens  
 106 and tissue samples were deposited in the MARE-Ispa tissue collection, ensuring proper  
 107 documentation and accessibility for future reference.

108

109 *Laboratory procedures*

110

111 Genomic DNA of the undigested contents was extracted using the REExtract-N-Amp  
 112 Tissue PCR kit from Sigma-Aldrich, following the manufacturer's instructions. For the  
 113 DNA-barcoding of undigested tissue, the mitochondrial DNA cytochrome c oxidase I  
 114 (COI) gene was used. PCR amplification of COI was conducted using primers COI-F1  
 115 (5'-TCAACCACCCACAAAGACATTGGCAC-3') and COI-R2 (5'-  
 116 ACTTCAGGGTGACCGAAGAATCAGAA-3') (Ward et al. 2005). PCR conditions  
 117 used were: 2 min initial denaturation at 95 °C, followed by 35 cycles of denaturing at  
 118 94 °C for 30 s, annealing at 52 °C for 30 s, extension at 72 °C for 1 min, and final  
 119 extension of 10 min at 72 °C. PCR products were purified and sequenced in STABVIDA  
 120 (Sanger sequencing, <https://www.stabvida.net/>).

121

122 Concerning metabarcoding of gastric fluids, four samples plus a blank were extracted  
 123 with the Qiagen DNeasy Blood and Tissue kit following the manufacturer's instructions,  
 124 including two elutions of 100 µl each with EB buffer. Metabarcoding was performed  
 125 using two-rounds of amplification via the polymerase chain reaction (PCR) with the  
 126 MiFish-U primer pair (Miya et al. 2015) plus a newly designed, species-specific reverse  
 127 blocking primer Creg-MiFishUR-  
 128 blockCCCAGTTTGTGTCATAGCTTTCGTGGGTTCAAAGTTG(C3spacer). The first  
 129 round of amplification was performed in 10 µl reactions including 5 µl of Qiagen  
 130 Multiplex Mastermix, 0.5 µl of each of the MiFish-U primers (10 µM), 2.5 µl of  
 131 Creg\_MiFish-UR\_block (100 µM; 50x final concentration) and 1.5 µl of DNA from  
 132 stomach content extractions. The temperature profile followed a touch-down PCR  
 133 consisting in an initial denaturing step of 95 °C for 15 min; 11 cycles of 95 °C for 20 s,  
 134 65 °C for 15 s (-0.5 °C/cycle), and 72 °C for 15 s; followed by 19 cycles of 98 °C for 20  
 135 s, 60 °C for 15 s and 72 °C for 15 s; and a final extension step of 72 °C for 5 min. All  
 136 PCRs were performed in triplicate for each sample (including the extraction blank) and a

137 PCR blank was run per batch of PCR replicates. The second round of PCR added sample-  
138 and replicate-specific indexes and Illumina-compatible barcodes as described in Miya et  
139 al. (2015). Indexed libraries were cleaned using Ampure Beads (0.6X), quantified and  
140 pooled equimolar. Final insert size was checked on the TapeStation (Agilent  
141 Technologies) using the DNA D1000 High Sensitivity assay kit (Agilent Technologies).  
142 The final pooled library was validated on a real-time PCR machine using the KAPA  
143 Library Quantification Kit (KAPA BIOSYSTEM) for Illumina platforms, adjusted to 12  
144 pM, and sequenced on an Illumina MiSeq using the MiSeq Reagent Nano Kit (500 cycles)  
145 aiming at 20000 reads per sample.

146

#### 147 *DNA sequence data analyses*

148

149 The sequences obtained from individual fish prey items were visually inspected and  
150 edited with Codon Code Aligner (Codon Code Corporation,  
151 <https://www.codoncode.com/>). Molecular identification was performed by comparing  
152 sequences against those available in GenBank in the nucleotide (nt) database (online,  
153 accessed on 01.06.2024), using megablast in BLAST (Basic Local Alignment Search  
154 Tool). Species identification was confirmed if the sequence exhibited at least 97% query  
155 coverage and  $\geq 99\%$  identity at the species level (e.g., Hebert et al. 2003; Wong and  
156 Hanner 2008).

157

158 In metabarcoding analysis, raw reads were filtered for quality ( $-q\ 30$ ), trimmed of  
159 adaptors and primers, and discarded if shorter than 100 bp using 'cutadapt' (Martin 2011).  
160 The cleaned reads were subsequently filtered with 'dada2' (Callahan et al. 2016) if the  
161 maximum expected errors  $\geq 2$  and if ambiguous bases (i.e., N) were present. Forward and  
162 reverse reads were dereplicated and merged, allowing no mismatches ( $\text{maxMismatch}=0$ )  
163 to reconstruct the corresponding amplicon sequence variants (ASVs). The set of ASVs  
164 was inspected for and filtered from chimeric sequences (i.e., sequences resulting from 2  
165 or more biological sequences that were incorrectly aligned) using the method 'consensus'.  
166 The final set of ASVs was screened across samples and read counts were estimated per  
167 sample. Molecular identification of the final ASVs was performed using BLAST searches  
168 (megablast algorithm) against the nucleotide (nt) database on NCBI (accessed on 27.12.  
169 2022), using default parameters and a maximum of 10 hits per search. Given the short  
170 length of the MiFish-U barcode (~170 bp), stringent cut-offs were considered for taxon-

171 level identifications, namely:  $\geq 97\%$  query coverage and  $\geq 99\%$  identity at the species  
 172 level,  $\geq 97\%$  at the genus level and  $\geq 95\%$  at the family level. ASVs with read counts  
 173  $< 0.03\%$  of total read count per sample were considered as false positives and excluded  
 174 from each sample's taxon list (Calderón-Sanou et al. 2020).

175

176 **Results**

177

178 The *C. regalis* individuals sampled (N=30) had an average total length of  $439.9 \text{ mm} \pm$   
 179  $39.7 \text{ mm}$  (max: 540 mm; min: 370 mm) and an average total weight of  $893.9 \pm 243.34 \text{ g}$   
 180 (max: 1567.6 g; min: 498.1 g).

181

182 Visual inspection of stomach contents enabled the identification of crustaceans (Fig. 1A)  
 183 in the majority of the examined individuals. Pieces of fish muscle and nearly intact preyed  
 184 fish were also observed (Fig. 1B). A total of 20 fish prey specimens, originating from the  
 185 stomach contents of 13 weakfish individuals were visually identified and/or barcoded  
 186 through tissue analysis. The remaining fishes had only digested material or crustaceans.

187



188 **Figure 1.** **A** Undigested crustacean in the stomach of *Cynoscion regalis* from the Tagus  
 189 estuary (left) **B** *Cynoscion regalis* from the Tagus estuary and respective semi-digested  
 190 fish prey found in stomach contents (right).

191

192 The COI barcodes identified 5 species belonging to 4 families. Among the sampled prey,  
 193 half were barcoded as *C. regalis* (N=10), while the remaining were attributed to the  
 194 European anchovy *Engraulis encrasicolus* (Eugraulidae, N=7), the annular seabream  
 195 *Diplodus annularis* (Sparidae, N=1), the black seabream *Spondyllosoma cantharus*  
 196 (Sparidae, N=1), and the sand smelt *Atherina presbyter* (Atherinidae, N=1).

197

198 Metabarcoding of the gastric fluids revealed the presence of the following species:  
199 *Cynoscion regalis*, *Engraulis encrasicolus*, *Sardina pilchardus* and *Dicentrarchus*  
200 *punctatus*. In addition, other preys were identified at family-level as Sparidae  
201 and Scorpaenidae. This adds at least Clupeidae, Moronidae and Scorpaenidae to the list  
202 of families preyed on by the weakfish.

203

## 204 **Discussion**

205

206 The molecular approach employed in this study allowed the identification of the  
207 following prey species of *C. regalis*: *Atherina presbyter*, *Diplodus annularis*,  
208 *Spondylisoma cantharus*, *Engraulis encrasicolus*, *Sardina pilchardus* and  
209 *Dicentrarchus punctatus*. Sparidae and Scorpaenidae were also detected at family-level,  
210 adding Clupeidae, Moronidae and Scorpaenidae to the list of families preyed on by the  
211 weakfish outside its native area. It also confirmed the cannibalistic nature of *C. regalis*  
212 already documented in previous studies and its preference for pelagic preys (e.g.,  
213 Merriner 1975; Cerveira 2019, 2021) such as *Atherina presbyter*, *Engraulis encrasicolus*  
214 and *Sardina pilchardus*, which are common pelagic fish in the Portuguese estuaries.

215

216 Considering the prey species listed, they are all found in estuaries, at least in some part  
217 of their life cycle (for details on the species occurring in the Tagus estuary, see Cabral et  
218 al. 2001; Brandão 2021). These findings from the Tagus estuary align with the study by  
219 Cerveira et al. (2021) in Sado, further supporting evidence of trophic niche maintenance  
220 in non-native habitats. Cerveira et al. (2021) reported the presence of the black goby  
221 *Gobius niger* and horse mackerel *Trachurus trachurus*, absent from this study, but failed  
222 to identify 3% of the fish species due to the advanced state of digestion, where molecular  
223 methods offer a powerful tool.

224

225 Taken together, these results underscore the adaptability of *C. regalis* in exploiting  
226 available resources, which may amplify its ecological impact on estuarine ecosystems,  
227 critical as feeding grounds and nurseries. The large size the adults can attain, and the high  
228 fecundity of this species raise additional concerns about the potential effects of *C. regalis*'  
229 presence on local biodiversity (Bañón et al. 2017a, 2017b).

230



231 This study has limitations that should be considered when interpreting the findings. The  
232 relatively small sample size limits the representativeness of the results for the broader  
233 population, while collecting samples from a single estuary (Tagus) at a specific time  
234 introduces potential temporal and spatial biases, neglecting seasonal or geographical  
235 variations in diet. The reliance on a single genetic marker (COI for tissue sequencing and  
236 12S for metabarcoding) may not fully resolve taxonomic ambiguities, especially in  
237 closely related species, and the metabarcoding approach may favour the detection of  
238 species with less degraded or more abundant DNA. Additionally, the exclusion of  
239 amplicon sequence variants (ASVs) with low read counts (<0.03%) could result in the  
240 omission of rare prey items, which might play ecologically significant roles. The study's  
241 focus on dietary analysis excludes broader ecological interactions, such as prey  
242 availability, competition, and predator-prey dynamics, and it lacks complementary  
243 methods, such as stable isotope analysis, which could provide insights into long-term  
244 dietary trends. Furthermore, the primers used in metabarcoding may not amplify all prey  
245 taxa equally, leading to potential biases in diet composition, and the selection of semi-  
246 digested prey for tissue sequencing might overlook prey that is too degraded to yield  
247 usable DNA. These limitations and the need to include other prey items, like crustaceans,  
248 in future molecular approaches to *C. regalis* diet highlight the need for more  
249 comprehensive studies.

250

251 Despite these caveats, this study contributes to a deeper understanding of the ecological  
252 dynamics of invasive species. By incorporating molecular methods into the assessment  
253 of feeding ecology, we have demonstrated their value in providing detailed insights into  
254 dietary patterns, which are critical for understanding the ecological impacts of invasive  
255 species. Future monitoring and management of *C. regalis* populations in Portuguese  
256 waters are essential to mitigate unforeseen ecological consequences and preserve  
257 biodiversity in estuaries.

258

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260

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