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## **Suppression of an invasive pine by a native shrub following a megafire**

 Víctor Escobedo, Persy Gomez,  Marco A. Molina-Montenegro, Ian S. Acuña-Rodríguez

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2 Víctor M. Escobedo<sup>1,2</sup>, Persy Gómez<sup>1</sup>, Marco A. Molina-Montenegro<sup>1,3</sup>, and Ian S. Acuña-  
3 Rodríguez<sup>1,2</sup>

4 <sup>1</sup>Centro de Ecología Integrativa, Instituto de Ciencias Biológicas, Universidad de Talca, Talca,  
5 Chile

6 <sup>2</sup>Dirección de Investigación, Vicerrectoría Académica, Universidad de Talca, Talca, Chile

7 <sup>3</sup>Centro de Investigación en Estudios Avanzados del Maule (CIEAM), Universidad Católica del  
8 Maule, Talca, Chile

9 For correspondence: [victor.escobedo@utalca.cl](mailto:victor.escobedo@utalca.cl), [ian.acuna@utalca.cl](mailto:ian.acuna@utalca.cl).

10

11 **Abstract**

12 Seedling density of the Chilean wineberry *Aristotelia chilensis* negatively correlates with the  
13 seedlings' abundance of an invasive pine *Pinus radiata*, particularly in post-fire areas. This pattern  
14 emerged following a megafire in Chile's Coastal Maulino Forest, a biodiversity hotspot facing  
15 increasing fire threats. This pattern, coupled with a high proportion of plots lacking pine seedlings,  
16 suggests that *A. chilensis* may play a role in limiting *P. radiata* invasion. The negative relationship  
17 was strongest in areas with moderate fire severity, likely reflecting differences in shade tolerance.  
18 *A. chilensis*, a light-demanding species with some degree of shade tolerance, can persist in partially  
19 shaded environments. In contrast, *P. radiata*, a more strictly light-demanding species, struggles to  
20 establish under significant shade. In high-severity fires, however, we found no significant  
21 relationship between these species, likely due to detrimental effects on both species, including  
22 potential microbiome dependence for *A. chilensis*. As *A. chilensis* shows successful establishment  
23 at low fire severity, enhancing its post-fire recruitment, particularly in moderately burned areas,  
24 could be a valuable strategy for mitigating *P. radiata* invasion and restoring fire-affected  
25 Mediterranean ecosystems.

26 **Keywords**

27 Invasion resistance, Fire severity, Post-fire establishment, Soil microbiome

28

## 29 Introduction

30 Wildfires pose a significant threat to biodiversity, disrupting ecosystem functions and threatening  
31 sensitive habitats worldwide. Their increased frequency and intensity are attributed to various  
32 factors, including climate change and land-use modifications (McLauchlan et al. 2020). The  
33 Coastal Maulino Forest, a biodiversity hotspot in central Chile (Myers et al. 2000), is facing more  
34 frequent and intense wildfires in last decades, driven by factors such as rising temperatures, a  
35 megadrought and the forestry plantations of non-native species (González et al. 2018, 2023), some  
36 of which could become invasive after fire disturbances. Primarily, the invasive species is *Pinus*  
37 *radiata* (Pinaceae) which covers approximately 60% of the country's 2.5 million hectares of forest  
38 plantations (Bustamante and Simonetti 2005, González et al. 2018). The devastating 2017 “Las  
39 Máquinas” megafire burned over 200,000 ha of the Coastal Maulino Forest, a stark reminder of  
40 the vulnerability of this ecosystem (Valencia et al. 2018). Despite ongoing active and passive  
41 restoration efforts in south-central Chile (Morales et al. 2021, Souza-Alonso et al. 2022),  
42 challenges persist, including the rapid arrival of post-fire pine regeneration that hinders restoration  
43 success (Gómez et al. 2019, González et al. 2020, 2023). This highlights the need for conservation  
44 and restoration practices tailored to this unique ecosystem.

45 Invasive species often display rapid resource utilisation, potentially outcompeting native  
46 species and promoting more frequent fire events. This can create a positive invasion-fire feedback  
47 loop (Contreras et al. 2011, Taylor et al. 2017). *P. radiata* is a light-demanding and shade-intolerant  
48 species known for its aggressive post-fire regeneration through serotinous cones, which release  
49 large amounts of viable wind-dispersed seeds after fire events (Franzese and Raffaele 2017).  
50 Studies have shown a higher probability of fire ignition in areas dominated by *P. radiata*  
51 plantations compared to native forests in south-central Chile (Contreras et al. 2011, Gómez-  
52 González et al. 2019).

53 Previous research suggests limited success in controlling *P. radiata* invasion through  
54 overall native species diversity (Gómez et al. 2019, González et al. 2020). However, recent field  
55 studies provide evidence that the native wineberry species *Aristotelia chilensis* (Elaeocarpaceae)  
56 efficiently recolonises burnt areas even where *P. radiata* is present (Promis et al. 2019, Becerra et  
57 al. 2022). *A. chilensis* is a fast-growing, light-demanding, fleshy-fruited bird-dispersed tree  
58 species with a semi-dioecious leaf habit. These traits allow it to not only colonise clearings but

59 also persist after plantations replace native forests because it can exhibit some shade tolerance  
60 (Guerra et al. 2010, Salgado-Luarte and Gianoli 2012). This rapid establishment and fast growth  
61 of *A. chilensis* would align with the concept of the “pre-emptive resource effect” – a mechanism  
62 where early colonising native species can outcompete invasive plants by monopolising essential  
63 resources (Byun et al. 2013, Byun and Lee 2017, Delavaux et al. 2023). Additionally, studies  
64 suggest that *P. radiata*, being a shade-intolerant species, might struggle to establish into a darker  
65 understory dominated by *A. chilensis* and other native species (Gómez et al. 2019, Becerra and  
66 Simonetti 2020). The efficient colonisation and fast growth of *A. chilensis* suggest that it has the  
67 potential to act as a native plant competitor against *P. radiata* invasion in fire-affected ecosystems.

68 Building upon competition-based biotic resistance (Elton 1958) and the theory of limiting  
69 similarity, where native species can limit invasive plant establishment due to niche overlap, we  
70 hypothesised that *P. radiata* abundance would negatively correlate with increasing *A. chilensis*  
71 abundance. Specifically, we tested the relationship between the abundance of *A. chilensis* and *P.*  
72 *radiata* in plots affected by varying fire severity levels caused by the Las Máquinas mega-fire in  
73 the Maulino Coastal Forest. Additionally, we explored whether fire severity modulates this  
74 relationship. Moderate- or low-severity fires that increase light penetration while retaining  
75 understory vegetation could favour *A. chilensis* establishment, potentially strengthening its  
76 competitive effect on *P. radiata* (i.e., a negative relationship). In contrast, high-severity fires that  
77 create harsher conditions and potential soil disruption (i.e., depleting the soil microbiome) could  
78 hinder the establishment of both *A. chilensis* and *P. radiata*, obscuring any competitive effects. By  
79 elucidating these dynamics, we aim to provide valuable data to guide and enhance conservation  
80 and restoration efforts in fire-affected areas across the central Mediterranean region of Chile.

## 81 **Materials and methods**

### 82 *Study Site*

83 The study was conducted at El Porvenir (35°42' S, 72°22' W), located at the northern edge of the  
84 Coastal Maulino Forest in central-south Chile (Gómez et al. 2022). El Porvenir is a fragment of  
85 native mesic forest type, surrounded by large stands of planted *P. radiata* and *Eucalyptus globulus*  
86 (Myrtaceae). The dominant tree species include *Nothofagus glauca* (Nothofagaceae), *N.*  
87 *alessandrii*, *N. obliqua*, *Cryptocarya alba* (Lauraceae), *Aextoxicon punctatum* (Aextoxicaceae),  
88 *Gevuina avellana* (Proteaceae), and *A. chilensis*. Study area has a Mediterranean climate with a

89 mean annual precipitation of 918 mm and a mean annual temperature of 12.7 °C (Becerra and  
90 Simonetti 2020).

91 In January 2017, the Las Máquinas megafire affected El Porvenir, which experienced fire  
92 severity ranging from low to high (see Valencia et al. 2018, Gómez et al. 2022). Following the fire,  
93 several species exhibited regeneration at different levels, with high seedling recruitment for the  
94 invasive *P. radiata* and the native *A. chilensis* (Gómez et al. 2022).

#### 95 *Plot establishment and seedling survey*

96 To assess the potential role of *A. chilensis* in limiting *P. radiata* invasion, we established twenty-  
97 three 625 m<sup>2</sup> plots across El Porvenir. These plots were randomly distributed within areas  
98 experiencing low (n = 8), moderate (n = 10), and high (n = 5) fire severity (see Gómez et al. 2022).  
99 Seedling surveys were conducted at 8 and 24 months following the Las Máquinas mega-fire  
100 (hereafter 2017 and 2019). To estimate the density of *A. chilensis* and *P. radiata* seedlings, three 1  
101 m<sup>2</sup> sub-plots were randomly located within each plot to search for regenerating *A. chilensis* and *P.*  
102 *radiata* seedlings under 60 cm in height. The average number of seedlings per 3 m<sup>2</sup> sampled area  
103 was  $21.98 \pm 20.34$  for *A. chilensis* and  $7.46 \pm 12.78$  for *P. radiata*. To confirm that the seedlings  
104 originated from seeds and not resprouts, we collected at least three random plant samples per  
105 species from each sub-plot for root system examination.

#### 106 *Data analysis*

107 We performed a negative binomial Generalised Linear Mixed-effects Model (NB GLMM) to  
108 analyse the relationship between the abundance of *A. chilensis* and *P. radiata* seedlings. This  
109 statistical method is suitable for counting data with overdispersion, a common characteristic of  
110 ecological data. Here, we account for the potential influence of sampling time at each plot by  
111 including time since the fire as a random factor nested within fire severity. This nested structure  
112 considers the variation in fire severity across the landscape while acknowledging the potential  
113 influence of sampling time within each fire severity category (see above). Additionally, we  
114 conducted separate NB GLMM analyses for each fire severity level, including time sampling as a  
115 random factor.

#### 116 **Results and Discussion**

117 Our analysis revealed a negative relationship between *A. chilensis* and *P. radiata* abundance across  
118 the study site ( $\chi^2_{(1,46)} = 8.0707$ ,  $p < 0.01$ ; Fig. 1). Thus, areas with higher numbers of *A. chilensis*  
119 seedlings have fewer *P. radiata* seedlings, potentially indicating a suppressive effect of native  
120 species on invasive tree establishment.

121 Furthermore, our results suggest that the strength of this negative relationship varied  
122 depending on fire severity. A significantly negative relationship between *A. chilensis* and *P. radiata*  
123 was found in areas with moderate fire severity ( $\chi^2_{(1,20)} = 16.385$ ,  $p < 0.01$ ; Fig. 2). In contrast, these  
124 two species had no significant relationship in plots with high or low fire severity (Fig. 2). This  
125 pattern hints that fire severity might play a role in mediating the interaction between *A. chilensis*  
126 and *P. radiata*. Wildfire severity plays a crucial role in shaping post-fire succession and ecosystem  
127 dynamics. Understanding the severity-specific effects of fires is essential for developing effective  
128 forest restoration and conservation management strategies.

129 The observed negative correlation between the abundance of *A. chilensis* and *P. radiata*  
130 suggests that the former's presence, as a component of the pre-fire native flora, can be considered  
131 a predictor variable influencing *P. radiata* establishment in post-fire areas. Given that *A. chilensis*  
132 was already present in these ecosystems before the fires, its abundance at the time of the fire event  
133 likely influenced the available resources and habitat conditions for *P. radiata* establishment.  
134 Several mechanisms could explain this phenomenon, including the priority effect by pre-empting  
135 resources and habitat filtering (Byun et al. 2013, Byun and Lee 2017). First, *A. chilensis* is a fast-  
136 growing, light-demanding species. In areas with a higher abundance of *A. chilensis*, competition  
137 for light, water, or nutrients could be hindering the successful establishment of *P. radiata* seedlings.  
138 Future studies that quantify resource availability and seedling performance concerning *A. chilensis*  
139 density could provide stronger evidence for this hypothesis. Second, fire can have profound and  
140 different effects on plant community assembly depending on its severity (McLauchlan et al. 2020).  
141 The environmental conditions created by moderate fire severity may be more favourable for the  
142 establishment of native compared to invasive species. Specifically, these fires create a more open  
143 canopy with increased light availability in the understory, typically forming a patchy mosaic of  
144 burned and unburned areas rather than eliminating the entire canopy. While *A. chilensis* is a light-  
145 demanding species with some degree of shade tolerance (Guerra et al. 2010, Salgado-Luarte and  
146 Gianoli 2012), *P. radiata* is a strictly shade-intolerant species (Gómez et al. 2019). Thus, this

147 variation in light availability could favour *A. chilensis* over *P. radiata* establishment in suitable  
 148 microsites within the burned landscape. In low-severity fires with more remaining canopy cover,  
 149 *P. radiata* pine showed very low establishment (only two plots with 9 and 15 seedlings), likely due  
 150 to limited light availability for germination and seedling growth. In contrast, *A. chilensis*, which  
 151 can exhibit some shade tolerance, could persist and thrive, with an average of 16 seedlings per plot  
 152 and up to 67 in one case (Fig. 2). High-severity fires present a vastly different scenario, where  
 153 most or all vegetation is fire-consumed and heat sterilises the soil, eliminating vital microbes and  
 154 disrupting biogeochemical processes. These harsh conditions are detrimental to the establishment  
 155 of both species, resulting in the lack of relationship observed in Fig. 2. In this line, *A. chilensis*'s  
 156 lower establishment suggests a dependence on healthy soil microbes (Escobedo et al.  
 157 unpublished), which are eliminated by high-severity fires. *P. radiata*, meanwhile, sometimes  
 158 showed higher abundance in these areas, potentially benefiting from the open and *A. chilensis*-free  
 159 conditions since its establishment and survival are less reliant upon microbe communities  
 160 (Escobedo et al. unpublished). This resilience disparity highlights the threat of high-severity fires  
 161 to the Coastal Maulino forest.

162 Our findings suggest that native *A. chilensis* might play a role in limiting *P. radiata*  
 163 invasion, potentially through competition and/or habitat filtering. However, this beneficial effect  
 164 might be compromised in high-severity fire areas, where *A. chilensis* establishment is hampered  
 165 due to its dependence on a healthy soil microbiome. High-severity fires that disrupt soil microbial  
 166 communities pose a significant threat to native plant communities and their ability to resist  
 167 invasion. Promoting the establishment of native species like *A. chilensis*, particularly in areas with  
 168 moderate fire severity, could be a valuable strategy for mitigating tree invasion and fostering the  
 169 recovery of fire-affected Mediterranean ecosystems like the Maulino forest in central Chile.

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## 174 **Data Availability**



175 The data supporting this study's findings will be available in a repository upon manuscript  
176 acceptance.

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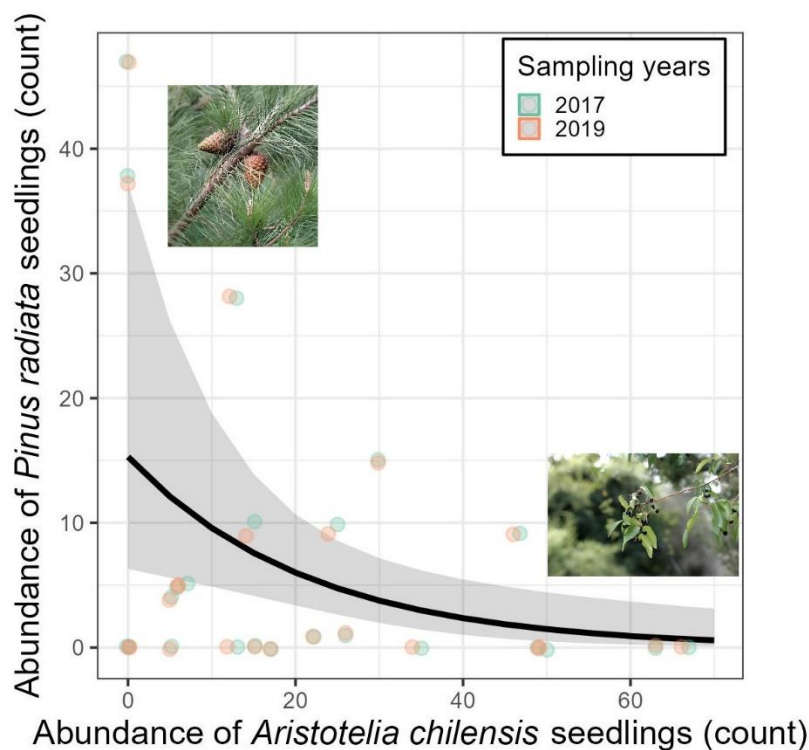
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295

## 296 Figures

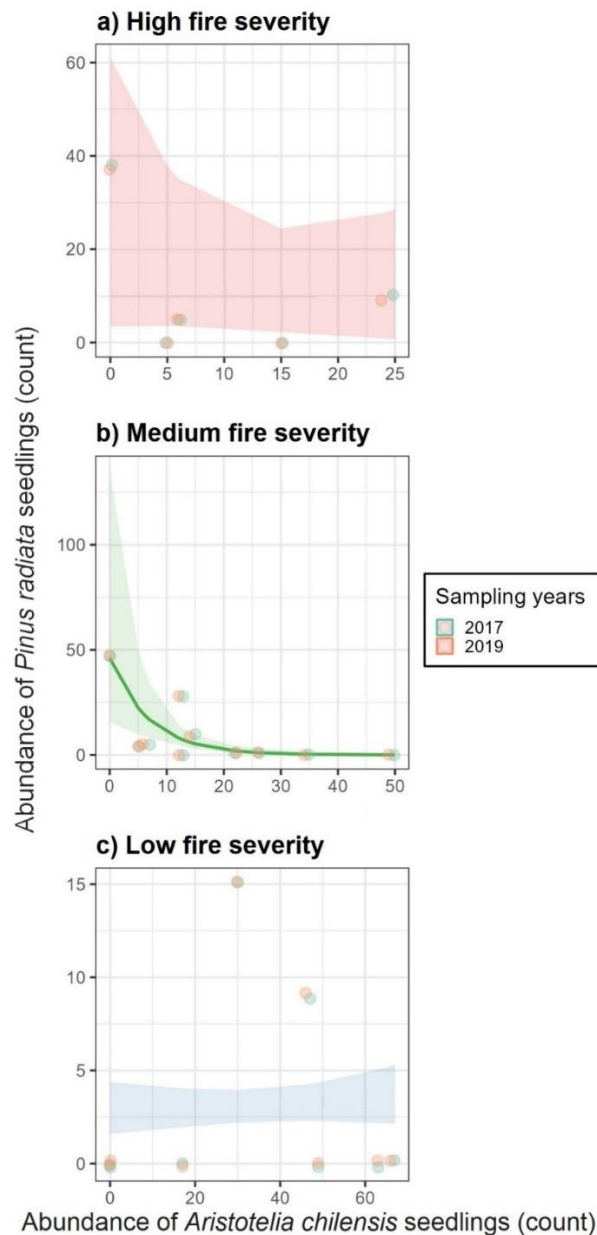
297 **Figure 1.** Model-predicted relationship between *Aristotelia chilensis* and *Pinus radiata* seedlings  
298 abundance for two sampling times (2017 and 2019). Line indicates a statistically significant  
299 negative relationship ( $p < 0.05$ ) based on a negative binomial GLMM.



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302 **Figure 2.** Model-predicted relationships between *Aristotelia chilensis* and *Pinus radiata* seedling  
 303 abundance across fire severity levels (a, low; b, medium; c, high) for two sampling times (2017  
 304 and 2019). The solid line in the middle panel (b, moderate-severity fire area) indicates a  
 305 statistically significant negative relationship ( $p < 0.05$ ) based on a negative binomial GLMM.  
 306 Relationships were not statistically significant in high- or low-fired-severity areas.



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