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Evaluating five primer pairs for environmental DNA metabarcoding of Central European fish species based on mock communities

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3 **Evaluating five primer pairs for environmental DNA**
4 **metabarcoding of Central European fish species**
5 **based on mock communities**
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38 Running head: Fish mock community for eDNA metabarcoding
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50 **Abstract**

51 Environmental DNA (eDNA) metabarcoding has become a powerful tool for examining
52 fish communities. The demand for methodological standardization and the
53 implementation of eDNA-based assessments into the regulatory monitoring (e.g.,
54 Water Framework Directive) are imminent. To ensure methodical accuracy and to meet
55 regulatory standards, various sampling, laboratory and bioinformatic workflows have
56 been established. However, a crucial prerequisite for a comprehensive fish monitoring
57 is the choice of suitable primer pairs to accurately depict the present fish fauna. Various
58 fish-specific primer pairs targeting different genetic marker regions were published
59 over the past decade. However, a dedicated study to evaluate performance of
60 frequently applied fish primer pairs to assess Central European fish species has not
61 yet been conducted. Therefore, we created an artificial community composed of DNA
62 from 45 Central European fish species and examined the discriminatory power and
63 reproducibility of five fish primer pairs. Our study highlights the effect of the primer
64 choice and bioinformatic filtering on the outcome of eDNA metabarcoding results. From
65 the five primer pairs evaluated in our study the tele02 (12S gene) primer pair proved
66 to be best choice for eDNA metabarcoding of Central European freshwater fish. Here,
67 the MiFish-U (12S) and SeaDNA-mid (COI) primer pairs also displayed good
68 discriminatory power and reproducibility. However, more general primer pairs (i.e.,
69 targeting vertebrates) were found to be less reliable and generated high numbers of
70 false-positive and false-negative detections. Our study illustrates how careful selection
71 of primer pairs and bioinformatic pipelines can make eDNA metabarcoding a more
72 reliable tool for fish monitoring.

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78 Introduction

79 Environmental DNA (eDNA) metabarcoding has become a valuable tool for monitoring
 80 fish species in different habitats (McDevitt et al. 2019, Wang et al. 2021, Miya 2022).
 81 Several studies have compared eDNA-based monitoring to traditional monitoring
 82 approaches, such as gillnetting or electrofishing, proving eDNA metabarcoding to be a
 83 reliable, fast, sensitive, non-invasive and cost-efficient method for fish detection (Pont
 84 et al. 2018, Fujii et al. 2019, Boivin-Delisle et al. 2021). However, applying eDNA
 85 metabarcoding comes with certain challenges such as the selection of appropriate
 86 sampling strategies and wet lab processing steps, completeness of reference
 87 databases, and choice of appropriate primers (Evans et al. 2017, Kumar et al. 2022).
 88 As a prerequisite for a comprehensive biodiversity monitoring, suitable primers are
 89 crucial to avoid false-negative detection and accurately depict the present fish fauna
 90 (Schenekar et al. 2020). Mock community metabarcoding is an efficient *in vitro*
 91 approach to test the performance of primers using an artificially composed DNA
 92 mixture representing the expected target community for biomonitoring (Hänfling et al.
 93 2016, Elbrecht et al. 2019). While different metabarcoding fish primers have been
 94 evaluated on natural communities, larger systematic tests of primers with fish mock
 95 communities are missing (Bylemans et al. 2018a, Miya et al. 2020, Zhang et al. 2020,
 96 Shu et al. 2021). These studies focused on the detection of Asian and Australian fish
 97 species, which are genetically divergent and differing in species composition from the
 98 Central European fish fauna. Primer pairs for European fish communities have for now
 99 only been evaluated for estuarine and costal eDNA samples (Collins et al. 2019) and
 100 on smaller scale for UK lake fish (Hänfling et al. 2016). Thus, especially for the
 101 implementation of fish eDNA metabarcoding in routine monitoring programs such as
 102 the European Water Framework Directive (Hering et al. 2018, Pont et al. 2021), it is
 103 crucial to evaluate suitable primer pairs regarding their detection ability of the most
 104 common European freshwater fish species and investigate false-positive and false-
 105 negative detections.
 106 In this study, we addressed this issue and evaluated five published fish eDNA
 107 metabarcoding primers (targeting fish and other vertebrates) by testing their
 108 performance on an artificial community composed of DNA from 45 Central European
 109 fish species. Here, we examined the discriminatory power and reproducibility of the
 110 five primer pairs, investigated their false-positive and false-negative detection rates,
 111 and investigated primer-specific biases. Finally, we conclude with a primer pair
 112 recommendation for eDNA metabarcoding approaches targeting fish in routine
 113 monitoring campaigns.

114 Methods

115 Fish swabs

116 Mucus samples of 66 specimens (45 species) were collected by fish bioassessment
 117 experts during electrofishing campaigns in autumn 2020 at five sites across Germany,
 118 covering both the Rhine and the Danube catchment. Each mucus sample was
 119 collected individually using sterile swabs (FLOQ Swab 80 mm, minitip, without
 120 medium, sterile sleeve; COPAN, Italy). All fish were handled as quickly as possible
 121 outside the water to keep the stress to a minimum, while a sterile swab was moved
 122 across the specimens' flank. Swabs were placed back into the sleeve and sealed. After
 123 field work, samples were stored at 4°C until delivery to the University of Duisburg-

124 Essen. Upon arrival the swabs were stored at -20°C overnight followed by DNA
125 extraction.

126 DNA extraction

127 Swab tips were clipped off at the handle and placed in a sterile 1.5 mL Eppendorf tube
128 before 1 mL TNES buffer and 15 µL Proteinase K (300 U/mL, 7BioScience, Neuenburg
129 am Rhein, Germany) was added to the sample. Samples were incubated at 55°C and
130 shaken at 1000 rpm for 3 h on an Eppendorf ThermoMixer C (Eppendorf AG, Hamburg,
131 Germany). Subsequently, DNA was extracted using an adapted NucleoMag tissue kit
132 (Macherey Nagel, Düren, Germany; Supplementary Material 1). In total, a volume of
133 400 µL per sample was extracted and DNA was eluted in a final volume of 50 µL elution
134 buffer. DNA concentration of each sample was measured using a Qubit dsDNA HS
135 Assay-Kit on a Qubit v2 fluorometer (Thermo Fisher Scientific).

136 Mock community composition

137 Two fish mock communities were created using the extracted fish swab DNA. In case
138 of several collected specimens per species, only the sample with the highest DNA
139 concentration was used for the composition of the mock community in order to
140 represent each species only by a single individual. The first normalized mock
141 community (MC1) was equimolarly pooled to 2 ng DNA per species. A second mock
142 community (MC2) was pooled using 1 µL of each extract to generate a mock
143 community with different DNA concentrations per species. MC2 was used to test for
144 potential correlation between DNA concentration and number of reads. Both mock
145 communities contained DNA of 45 fish species.

146 DNA amplification and sequencing

147 Both mock communities were assessed using five different published primer pairs for
148 DNA amplification: tele02 (Taberlet et al. 2018), MiFish-U (Miya et al. 2015), 12Sv5
149 (Riaz et al. 2011), SeaDNA-mid (Collins et al. 2019) and L2513/H2714 (Kitano et al.
150 2007). A two-step PCR approach (Bohmann et al. 2022) was applied for amplifying the
151 molecular marker genes and tagging of amplicons with barcodes and Illumina
152 sequencing adaptors. In the 1st-step PCR step, tagged versions of the five fish primer
153 pairs were used (Table 1).

154 In total, 60 1st-step PCR amplifications were conducted, including five replicates for
155 each mock communities (MC1 and MC2) and two negative PCR controls for each of
156 the 5 primer pairs. The reaction volume was 25 µL, consisting of 12.5 µL Multiplex
157 Mastermix (Qiagen Multiplex PCR Plus Kit, Qiagen, Hilden, Germany), 7 µL PCR-
158 grade water, 2.5 µL CoralLoad dye, 0.5 µL forward primer, 0.5 µL reverse primer (10
159 µM each), and 2 µL of DNA template. The 1st-step PCR included following steps: 5 min
160 95 °C initial denaturation, followed by 10 cycles of 30 s at 95 °C, 90 s at decreasing
161 annealing temperature (starting from annealing temperature +10 °C), and 30 s at 72
162 °C, followed by 25 cycles of 30 s at 95 °C, 90 s at the respective annealing temperature
163 (tele02: 52 °C, MiFish-U: 59 °C, SeaDNA-mid: 53 °C 12SV5: 52 °C, and LH16S: 55
164 °C), and 30 s at 72 °C. The final elongation was 10 min at 68 °C. Subsequently, PCR
165 products were size selected using magnetic beads (ratio 0.7,
166 [dx.doi.org/10.17504/protocols.io.36wgqj45xvk5/v2](https://doi.org/10.17504/protocols.io.36wgqj45xvk5/v2)) to remove excessive primers and
167 reduce subsequent primer dimer formation.

168 In the 2nd-step PCR, Illumina sequencing adapters with a dual twin-indexing system
169 were added (Buchner et al. 2021, Bohmann et al. 2022). For each sample, the 2nd-step

170 PCR mix contained, 7.5 μ L Multiplex Mix, 1.8 μ L PCR-grade water, 1.5 μ L CoralLoad
171 dye, 1.2 μ L combined primer (5 μ M), and 3 μ L 1st-step PCR product. The 2nd-step PCR
172 included the following steps: 5 min 95 °C initial denaturation, followed by 10 cycles of
173 30 s at 95 °C and 120 s at 72 °C. The final elongation was 10 min at 68 °C. The 2nd-
174 step PCR products were visualized on a 1% agarose gel to evaluate amplification
175 success. Negative controls did not produce bands on the gel. Then, PCR products
176 were size selected using magnetic normalization beads (ratio 0.7,
177 [dx.doi.org/10.17504/protocols.io.q26g7y859gwz/v1](https://doi.org/10.17504/protocols.io.q26g7y859gwz/v1)) to normalize samples and
178 remove excessive primers and primer dimers. Subsequently, all normalized PCR
179 products were pooled into one library. The pooled library was concentrated using a
180 NucleoSpin Gel and PCR Clean-up kit (Macherey Nagel, Düren, Germany) following
181 the manufacturer's protocol. The final elution volume of the library was 40 μ L. The
182 library was then analysed using a Fragment Analyzer (High Sensitivity NGS Fragment
183 Analysis Kit; Advanced Analytical, Ankeny, USA) to check for potential primer dimers
184 and co-amplification, and to quantify the DNA concentration of the library. The final
185 library was sequenced on a MiSeq 250bp PE V3 Illumina platform at CeGat (Tübingen,
186 Germany).

187 **Bioinformatics**

188 Raw reads were received as demultiplexed fastq files. All samples were processed
189 with the APSCALE-GUI pipeline v1.1.6 (Buchner et al. 2022), which is based on
190 VSEARCH (Rognes et al. 2016) and cutadapt (Martin 2011). Each primer pair was
191 processed separately. All settings were kept as default, and OTUs were clustered with
192 a 97% percentage similarity threshold. Subsequently, taxonomy was assigned using
193 the 'local BLAST' function in APSCALE with the Midori2 databases (v249 of CO1,
194 IrRNA and srRNA; Leray, Knowlton, and Machida 2022) as reference.
195 The taxonomic assignment of each OTU was filtered using APSCALE-GUI
196 (Supplementary Figure 1). Initially, taxonomic assignments were filtered by e-value
197 (hits with the lowest e-value are kept) and hits with the same taxonomy were
198 dereplicated. Subsequently, taxonomic assignments were adjusted according to
199 similarity thresholds (species \geq 97%, genus \geq 95%, family \geq 90%, order \geq 85%). If at
200 this point more than one taxon assigned to species level was remaining, additional
201 filtering and flag raising steps were performed as follows: All ambiguous taxa were
202 saved to a separate column in the taxonomy table. The number of occurrences per
203 remaining taxon was counted. If a dominant species was present, it was selected as
204 taxonomic assignment ("F1 - Dominant taxon"). Otherwise, if two species of the same
205 genus remained, the genus was saved with the two possible species names separated
206 by slash (e.g., *Leuciscus idus/leuciscus*; "F2 - Two species, one genus"). If more than
207 two species belonging to different genera remained, the number of genera was
208 counted. If one genus (and multiple species) was present the genus was saved (e.g.,
209 *Hucho* sp. with the ambiguous assignments *Hucho bleekeri*, *Hucho hucho*, and *Hucho*
210 *taimen*; "F3 - Multiple species of one genus"). Lastly, if more than one genus remained
211 and no dominant taxon was present, the taxonomic assignment was trimmed to the
212 most recent common taxon ("F4 - Multiple genera"). Both the taxonomy and read tables
213 were then converted to TaXon tables (Supplementary Material 2) for downstream
214 analyses in TaxonTableTools v1.4.7 (Macher et al. 2021a). To account for potential
215 contamination the sum of reads in the negative controls of each OTU was subtracted
216 from the number of reads for the respective OTU of each sample ('Negative control
217 subtraction' tool). Subsequently, all tables were filtered for fish and lamprey species
218 (Supplementary Material 3). Here, all OTUs with a \geq 97% similarity but without species

219 assignment were manually checked and adjusted if e.g., a hybrid or erroneous entry
 220 was preventing a species assignment (Supplementary Material 4). If the taxonomy was
 221 ambiguous due to the assignment to geographically clearly separated species with
 222 equal similarity values, the species which is reported from the area was selected. The
 223 distribution information was collected from the gbif database (www.gbif.org).
 224 Analyses were performed using custom python scripts and results were visualized
 225 using the plotly package (<https://plot.ly>). For all primer pairs, the OTU and read
 226 proportions of target taxa (i.e., fish and lamprey) and bycatch taxa (i.e., all other taxa)
 227 were calculated. Additionally, the number of ambiguous species-level OTUs and the
 228 number of occurrences of each flag was calculated. For all subsequent analyses the
 229 manually adjusted TaXon tables were used.
 230 First, the relative read abundances (%) for all species present in the mock community
 231 (i.e., true positive species) and all non-target species (i.e., false-positive species) were
 232 calculated. Here, for each species the number of positive detections and the standard
 233 deviation of the relative read abundances across the five primer pairs were calculated.
 234 Second, Venn diagrams comparing the detected species of each primer pair to the
 235 original fish mock community composition were created. Additional Venn diagrams
 236 were created to compare the pre-adjusted TaXon tables. Oversplitting rates (i.e.,
 237 number of additional OTUs) were calculated for all species and each primer pair. PCR
 238 replicates were investigated by calculating the mean, minimum, and maximum Jaccard
 239 index of all five technical replicates per primer pair. The log transformed number of
 240 reads and the log transformed DNA concentration (ng/μL) were plotted and Spearman
 241 rho coefficients were calculated. Also, the log transformed number of reads per species
 242 of MC2 were plotted against the log transformed reads per species of MC1 and a
 243 Spearman rho coefficient was calculated. Lastly, the number of taxonomically assigned
 244 OTUs and unique species per family and the number of false-positive and false-
 245 negative assignments for each primer pair were calculated and plotted in two
 246 heatmaps.

247 Results

248 According to the fishbase database (fishbase.org), 123 fish species are reported from
 249 Germany (occurrence categories: “endemic”, “introduced”, “native”, “not established”,
 250 “questionable”, and “stray”). Here we manually added the round goby (*Neogobius*
 251 *melanostomus*) and the rainbow trout (*Oncorhynchus mykiss*), as they are both
 252 invasive species in Germany, but were not present in the fishbase list. Consequently,
 253 our fish mock community of 45 Central European freshwater fish species represents
 254 about 36.6% of fish reported from Germany (Supplementary Table 2). In detail, our
 255 mock community accounts for 50% of “native”, 26% of “introduced”, 22.2% of
 256 “questionable”, and 8% of “not established” fish species in Germany.
 257 Sequencing yielded a total of 8,254,293 raw reads across all primer pairs. In total,
 258 7,745,593 quality-filtered reads were clustered into 140 (tele02), 105 (MiFish-U), 120
 259 (12SV5), 111 (SeaDNA-mid), and 142 (LH16S) OTUs, respectively. Nearly all primer
 260 pairs showed little amplification of non-fish OTUs (between 96 to 98% fish OTUs),
 261 except for the SeaDNA-mid primer pair, which exhibited 50% non-fish OTUs, (Figure
 262 1A). However, only few reads were assigned to non-target OTUs for all primer pairs
 263 (between 98.2 and 100% fish OTUs; Figure 1B).
 264 The proportions of flagged taxonomic assignments varied between the five different
 265 primer pairs. Here, both the MiFish-U and tele02 primer pairs had the highest
 266 proportion of supported species-level OTUs (both 60%), followed by the SeaDNA-mid
 267 (49%), 12SV5 (46%), and LH16S (41%) primer pairs (Figure 1C). The first flag (‘Two

268 species, one genus') was most prominent in the SeaDNA-mid (25%) and least
 269 prominent in the 12SV5 primer pair (10%). For the second flag ('Multiple species of
 270 one genus') again the SeaDNA-mid showed the highest proportions (16%), while both
 271 the tele02 and MiFish-U primer pairs had the fewest cases (5%). Furthermore, the
 272 12SV5 primer pair showed the highest proportion of the third flag ('Dominant taxon')
 273 with 27% assigned species-level OTUs, while again the tele02 primer pair showed the
 274 fewest (9%). The SeaDNA-mid primer did not have any cases of flag four ('Multiple
 275 genera'), while the LH16S primer pair had the most (9%). Overall, the most abundant
 276 ambiguous assignment was *Leuciscus idus/leuciscus* (10 total occurrences), followed
 277 by *Sander canadensis/lucioperca* (8), *Blicca bjoerkna* (7), *Proterorhinus*
 278 *semilunaris/marmoratus* (6), and *Cyprinus carpio* and *Hucho* sp. with each 5 cases
 279 (Supplementary Table 3). Overall, the genera *Leuciscus* and *Sander* showed the
 280 highest number of ambiguous taxonomic assignments (14 and 13, respectively).
 281 After removal of bycatch taxa and curation of ambiguous taxonomic assignments, the
 282 12SV5 primer pair (45) included most species, followed by LH16S (40), tele02 (39
 283 species), MiFish-U (37), and SeaDNA-mid (36). In comparison to the original mock
 284 community fish species composition, the tele02 dataset showed the highest
 285 congruence (2 false-positive species, 37 true positive, and 8 false-negative), followed
 286 by the MiFish-U (2, 35, 10) and SeaDNA-mid (3, 33, 12). Both the 12SV5 (18, 27, 18)
 287 and LH16S primer pair (17, 23, 22) were less congruent to the original mock community
 288 composition (Figure 2). The 12SV5 and LH16S primer pairs resulted in OTUs assigned
 289 to several marine fish taxa, which were not part of the mock community, including
 290 Acanthuridae (surgeon fishes), Kyphosidae (sea chubs), Ophidiidae (cusk-eel),
 291 Peristediidae (armoured sea robins), Pholidae, and Zoarcidae (eelpouts; Table 2).
 292 Regarding the number of false-positive and false-negative assignments per family, the
 293 LH16S primer pair showed high incongruencies to the mock community, particularly
 294 for the Leuciscidae (4 false-positive / 10 false-negative) and Percidae (2/2). Similarly,
 295 the 12SV5 primer pair had various false-positive and false-negative assignments for
 296 the Leuciscidae (6/5), Cyprinidae (4/0), or Gobionidae (3/1). The SeaDNA-mid primer
 297 showed only a moderate number of incorrect assignments in the Leuciscidae (2/6).
 298 Lastly, the tele02 and MiFish-U primer pairs were overall the least prone to false-
 299 positive assignments and only showed false-positive assignments in Leuciscidae
 300 (*Leuciscus aspius*) and Salmonidae (*Parahucho perryi* and *Brachymystax lenok*).
 301 As a measure of primer bias the standard deviation of relative read abundances was
 302 across primer pairs. Here the standard deviation varied between the primer pairs
 303 ranging from an average of < 0.01% (*Barbatula barbatula*, *Leucaspius delineatus*,
 304 *Neogobius melanostomus*, *Phoxinus phoxinus*, and *Romanogobio albiginnatus*) to a
 305 maximum of 7.5% (*Pungitius pungitius*; Table 3A). While most species were detected
 306 with at least four primer pairs (29 mock community species), 10 species were detected
 307 with three or less primer pairs. In total, six species were not detected by any of the
 308 primer pairs, namely *Cottus gobio*, *Gymnocephalus schraetser*, *Lampetra fluviatilis*,
 309 *Rutilus pigus*, *Umbra krameri*, and *Zingel zingel*. Most false-positive species were
 310 unique to one primer pair (34 of 37 species; Table 3B), while only three species were
 311 detected with two or more primer pairs, namely *Leuciscus aspius* (4 occurrences),
 312 *Pungitius platygaster* (2), and *Umbra pygmaea* (2).
 313 In total, 48 cases of oversplitting (in our case species with more than one OTU
 314 assigned) were observed (Supplementary Table 4). Most over-split species-level
 315 assigned OTUs were found with the tele02 primer pair (12), while all other primer pairs
 316 showed 9 cases of oversplitting. The highest oversplitting rate was observed in
 317 *Gymnocephalus cernua* (7-fold OTU to species ratio, tele02 primer pair) and *Tinca*
 318 *tinca* (7-fold, 12SV5). While no over-split species was found in all five or even four of

319 the primer pairs, six species were over split in three primer pairs (i.e., *Abramis brama*,
 320 *Blicca bjoerkna*, *Ctenopharyngodon idella*, *Gymnocephalus cernua*, *Hucho hucho*, and
 321 *Sander lucioperca*).
 322 PCR replicates were highly consistent for all investigated primer pairs. The 12SV5
 323 primer pair showed the highest reproducibility (mean Jaccard similarity of 0.99),
 324 followed by LH16S (0.98), SeaDNA-mid (0.96), tele02 (0.96), and MiFish-U (0.95). No
 325 correlations between log transformed input DNA concentration (ng/ μ L) and log
 326 transformed reads of the second mock community (MC2) were found for most of the
 327 primer pairs (Spearman's rho between 0.21 and 0.34, $p \geq 0.05$) except for the MiFish-U
 328 primer pair ($p \geq 0.05$), which showed a moderate positive correlation (Spearman's rho =
 329 0.41) (Supplementary Figure 2). However, when comparing the number of log
 330 transformed reads per species between MC1 and MC2, significant correlations for the
 331 tele02 (rho=0.79, $p \leq 0.05$), MiFish-U (0.78, $p \leq 0.05$), 12SV5 (0.85, $p \leq 0.05$), SeaDNA-
 332 mid (0.81, $p \leq 0.05$), and LH16S (0.6, $p \leq 0.05$) primer (Supplementary Figure 3) were
 333 found.

334 Discussion

335 Discriminatory power, and reproducibility

336 Our primer evaluation based on mock communities of 45 European freshwater fish
 337 species confirmed the previously reported high discrimination power for two primer
 338 pairs (MiFish-U and tele02) belonging to the MiFish primer group (Bylemans et al.
 339 2018a, Taberlet et al. 2018, Collins et al. 2019, Polanco F. et al. 2021). The tele02
 340 primer pair (a modified version of the MiFish-U primer pair) performed particularly well
 341 in our study and clearly showed the highest species specificity and discriminatory
 342 power for European freshwater species. Until now the tele02 primer pair was evaluated
 343 *in silico* (Taberlet et al. 2018, Collins et al. 2019) as well as for water samples from
 344 Beijing, where it exhibited outstanding detection success of fish diversity in comparison
 345 with other fish-specific primers tested (Zhang et al. 2020). Accordingly, our results
 346 show that the tele02 primer pair recovered most true-positive species while producing
 347 the lowest number of false-positive and negative detections. From all primer pairs
 348 tested in this and other studies, the tele02 primer pair is arguably the best currently
 349 available choice for fish eDNA metabarcoding of European freshwater fish. While the
 350 SeaDNA-mid primer pair, targeting the COI gene, showed comparable good
 351 discriminatory power (i.e., true-positive detections), the co-amplification of non-fish
 352 taxa with this primer pair might be of concern. The fish mucus likely accumulates eDNA
 353 molecules and thus also contains DNA from other organisms than the fish itself. Here,
 354 the SeaDNA-mid primer pair was the only primer pair that showed high numbers of
 355 non-target OTUs. While non-target OTUs were observed in low read abundances for
 356 the mock communities, co-amplification issues could be more pronounced when
 357 applying the SeaDNA-mid primer pair on environmental samples. Here, comparably
 358 deeper sequencing depths might be required to detect all present fish species in an
 359 environmental sample with more non-target DNA, which would reduce the cost-
 360 efficiency per sample. The remaining two primer pairs 12SV5 and LH16S were
 361 designed to generally amplify vertebrate DNA (Kitano et al. 2007, Riaz et al. 2011,
 362 Hänfling et al. 2016, Harper et al. 2019). We decided to include these primer pairs
 363 since they have the potential for more holistic monitoring approaches, e.g., targeting
 364 the whole vertebrate community associated to a freshwater habitat (Pertoldi et al. 2021,
 365 Dou et al. 2023). However, the broader target range resulted in a drastically lower
 366 detection rate of fish species for the vertebrate primer pairs.

367 Overall, all primer pairs generated highly reproducible taxa lists among the PCR
 368 replicates for the fish mock communities. However, this reproducibility might not be
 369 achieved for environmental samples. Here, a generally lower reproducibility is
 370 expected and it is recommended to consider sufficient field and laboratory replicates
 371 to maximise species detection and minimize stochastic sampling effects (Sato et al.
 372 2017, Bylemans et al. 2018b, Macher et al. 2021b, Rojahn et al. 2021). Particularly the
 373 SeaDNA-mid primer pair might suffer from lower reproducibility for environmental
 374 samples due to the strong co-amplification.

375 **Primer bias**

376 Generally, several cases of potential primer amplification bias were observed, where
 377 certain species exhibited over-proportional read abundance in comparison to other
 378 primers. Despite our mucus samples most likely did not solely contain DNA of the
 379 target species, the amount of target input DNA was equally biased for all primer pairs.
 380 Thus, under optimal conditions without amplification biases, equal relative read
 381 abundances per species are to be expected. However, we observed distinct
 382 differences in the relative read abundances per primer pair. While no general
 383 amplification bias trend was observed, several species showed significantly higher
 384 read abundances for one of the primer pairs, such as *Perca fluviatilis* (12SV5: 12.65%
 385 to an average of other primer pairs of 1.04%), *Hucho hucho* (SeaDNA-mid: 17.36% to
 386 2.5%), or *Pungitius pungitius* (SeaDNA-mid: 17.49 to 0.65%). Using reads as a proxy
 387 for fish biomass has been addressed in various studies (Takahara et al. 2012, Kelly et
 388 al. 2019, Muri et al. 2020). However, next to uncertainties about the fate and state of
 389 eDNA in the environment, primer-specific PCR amplification biases as observed in our
 390 results can drastically affect read counts, depending on the choice of primer pair. This
 391 can lead to ambiguous or false conclusions about biomass estimates for the
 392 investigated fish community. While trends might exist, the interpretation of reads as
 393 proxy for biomass should be taken with care.

394 Another issue was the detection of multiple OTUs for certain species. While this does
 395 not affect the analysis when working on species level (i.e., OTUs of the same species
 396 are merged), the OTU alpha diversity is artificially inflated. Here we observed that
 397 particularly the Leuciscidae showed drastically higher numbers of OTUs than species.
 398 If the analysis of OTUs is of particular interest, this issue can be tackled by e.g., using
 399 a post-clustering curation algorithm, such as LULU filtering (Frøsløv et al. 2017), which
 400 should give more reliable biodiversity estimates e.g., when taxonomic references are
 401 lacking.

402 **False-negative assignments**

403 The here used Midori2 database is a curated version of the larger GenBank database
 404 and can be used as a reliable source for taxonomic assignment of fish OTUs. All
 405 species present in the mock community have reference sequences available for at
 406 least one genetic marker. However, seven species were not detected at all.

407 Amongst these was *Cottus gobio*, a common fish species in Central Europe for which
 408 117 reference sequences comprising all three investigated markers are deposited in
 409 GenBank. Although a taxonomic assignment was possible, no primer pair detected *C.*
 410 *gobio* in the mock communities. Since this species is frequently detected with eDNA
 411 metabarcoding from various sites and samples (Macher et al., unpublished data; tele02
 412 primer pair), it is likely that the *C. gobio* sample itself was the reason for the false-
 413 negative detection, as it might not have contained *C. gobio* DNA in sufficient

414 concentration or due to sampling or laboratory errors (e.g., specimen misidentification,
 415 swab inaccurately taken or DNA degradation).
 416 The stripped ruffe (*Gymnocephalus schraetser*) only has 13 reference sequences
 417 available in the Midori2 database, none of which is a 12S sequence. Consequently,
 418 the lack of reference for the 12S marker prevents a species level assignment for the
 419 tele02, MiFish-U, and 12SV5 primer pairs. However, all 12S primer pairs included
 420 OTUs assigned to *Gymnocephalus* that were trimmed to genus level due to low
 421 reference similarity threshold (< 97%). While no primer pair was able to detect *G.*
 422 *schraetser*, the SeaDNA-mid COI primer contained one ambiguous OTU assigned to
 423 *G. schraetser/cernua*. Thus, it remains unclear if the stripped ruffe can be distinguished
 424 from *G. cernua*, using eDNA primer pairs.
 425 Furthermore, various species are known to be indistinguishable with the short target
 426 fragment lengths used for eDNA metabarcoding. Particularly the two common lamprey
 427 species *Lampetra fluviatilis* and *L. planeri* could not be distinguished with any of the
 428 used primer pairs. The species status of these two 'sister species' has puzzled
 429 scientists for decades and while a genome-wide divergence can be observed (Mateus
 430 et al. 2013), they are known to share mitochondrial haplotypes (Espanhol et al. 2007).
 431 Considering that most eDNA primer pairs target short mitochondrial fragments of
 432 approximately 180 bp, a distinction of these species with eDNA metabarcoding will
 433 most likely not be possible in the foreseeable future.
 434 The zingel (*Zingel zingel*) was not detected by any primer pair despite the availability
 435 of various whole genome shotgun, COI, and sRNA reference sequences in the Midori2
 436 database. The closely related Danube zingel (*Zingel streber*) has various COI and 16S
 437 reference sequence available and was detected by the SeaDNA-mid and LH16S
 438 primer pair. Thus, the most likely explanation for the absence of *Zingel zingel* is errors
 439 in sampling or laboratory handling that led to the sample failure.

440 **Ambiguous assignments**

441 In several instances, the distinction between true-positive, false-positive, and false-
 442 negative assignments was very narrow. For several species, we observed
 443 misidentification with closely related species, which resulted in false-positive and false-
 444 negative assignments in single cases. For example, a species that was not detected
 445 by any primer pair is *Rutilus pigus*, the Danube roach. This species is closely related
 446 to the cactus roach (*R. virgo*) which was once considered a subspecies (*Rutilus pigus*
 447 subsp. *virgo* (Heckel, 1852)) and occurs in the same habitats. However, since
 448 molecular data showed that *R. pigus* and *R. virgo* are separate species (Pourshabanan
 449 et al. 2022), either the reference taxonomy is incorrect, which can occur in a non-
 450 curated database such as Genbank, or the specimen that was sampled for the mock
 451 community was actually *R. virgo*. For both species COI reference sequences are
 452 available in the Midori2 database, however, no 12S or 16S reference sequences are
 453 present. Here, the tele02 (1 OTU, 96.5%) and MiFish-U (1 OTU, 96.0%) both detected
 454 OTUs assigned to the genus *Rutilus*, besides *Rutilus rutilus* (which was present in the
 455 mock community), rendering these false-negative assignments as result of missing
 456 12S reference sequences. Furthermore, the false-positive *Rutilus virgo* assignment by
 457 the SeaDNA primer pair was most likely not a false-positive detection due to primer
 458 bias or lack of reference sequences but rather a lack of species name harmonisation
 459 or misidentification.
 460 For the European mudminnow (*Umbra krameri*), only 10 reference sequences (for 12S,
 461 COI, 16S or whole genome) are available in GenBank and it was not detected by any
 462 primer pair in our study. However, the SeaDNA-mid and 12SV5 primer pairs false-

463 positively detected the closely related species *Umbra pygmaea* and the teleo2, MiFish-
464 U, and LH16S detected *Umbra limi/pygmaea*. Both *U. limi* (Central mudminnow) and
465 *U. pygmaea* (Eastern mudminnow) are native to North America, and particularly the
466 latter has been introduced to Western and Central Europe. One explanation for the
467 incorrect assignments could be a misidentification of the specimen from which the
468 mucus sample was taken. If so, the specimen identified as European mudminnow was
469 truly an invasive Eastern mudminnow. This case should be further investigated since
470 the European mudminnow is listed as ‘vulnerable’ (IUCN Red List of Threatened
471 Species in 2010) and should ideally be distinguishable from the invasive Eastern
472 mudminnow with eDNA metabarcoding.

473 Furthermore, we observed several cases of “difficult” taxonomic assignments. Here,
474 particularly OTUs assigned to the genera *Hucho*, *Sander* and *Leuciscus* caused
475 ambiguities. The Danube salmon (*Hucho hucho*) was initially only detected by the
476 SeaDNA-mid and LH16S primer pairs. The three 12S primer pairs faced ambiguities
477 caused by hits to the Sichuan taimen (*Hucho bleekeri*) and the Siberian taimen (*Hucho*
478 *taimen*), which all share identical 12S sequences. However, since the Danube salmon
479 is the only present species of the genus *Hucho* in Central Europe, *H. bleekeri* and *H.*
480 *taimen* were ruled out for the tele02, MiFish-U and 12SV5 primer pairs. Similarly, the
481 pikeperch (*Sander lucioperca*) is geographically clearly separated from the sauger (*S.*
482 *canadensis*), but the two species are not genetically distinguishable with the
483 investigated markers, leading to flag 1 ambiguities (“Two species, one genus”). In this
484 case, however, based on the current distribution ranges, one can account for this
485 ambiguity, similarly to the Danube salmon. Nevertheless, if one of the *Hucho* or *Sander*
486 species were to be introduced to Central Europe, not all primer pairs could distinguish
487 the native species, which could be of concern for invasive species monitoring. The
488 common dace (*Leuciscus leuciscus*) and ide (*L. idus*), however, are highly prone to
489 causing flag 1 ambiguities. This can be caused by several reasons: for instance,
490 species of the family Leuciscidae are known to commonly hybridize, such as the bleak
491 (*Alburnus alburnus*) and chub (*Leuciscus cephalus*) (Wheeler 1978) or chub and roach
492 (*Rutilus rutilus*) (Wheeler and Easton 1978). This can lead to mitochondrial
493 introgression, causing reference sequences of different species to be identical.
494 Another reason is the wide distribution of common dace across Europe and its habitus
495 typical for the family Leuciscidae. This can result in false species identification, that is
496 propagated to incorrect database entries, which ultimately can lead to ambiguous
497 assignments. Here, a sophisticated curation of the Midori2 database, or the usage of
498 a custom reference database, including reference sequences from known source,
499 might help to reliably distinguish *L. leuciscus* and *L. idus*. Another reason for false-
500 negative assignments may occur in the automated taxonomic assignment of OTUs due
501 to unclear species status or the use of synonyms. For example, we were aware from
502 previous eDNA metabarcoding datasets that *Rhodeus amarus* and *R. sericeus* are
503 used synonymously and we corrected our dataset for this issue (*Rhodeus*
504 *amarus/sericeus*).

505 While in this study we used the Midori2 database, which is a curated version of the
506 Genbank database, another widely used reference library for mitochondrial sequences
507 is the MitoFish database (Sato et al. 2018). While reference sequences for most fish
508 are available in the MitoFish database, some species cannot be assigned due to the
509 absence of e.g., whole genome sequences (e.g., *Romanogobio albipinnatus*).
510 Additionally, the comparably lower overall number of reference sequences might be of
511 concern in light of intraspecific variation and could lead to false-negative assignments.

512 False-positive assignments

513 The detection of false-positives is of particular concern since it drastically reduces the
 514 robustness of taxa lists. Particularly the more general vertebrate primer pairs were
 515 prone to produce comparably high numbers of false-positive assignments. Here,
 516 12SV5 and LH16S were the only datasets that included marine fish taxa, which were
 517 not present in the mock community of Central European freshwater fish. Since no
 518 marine samples have been processed in this laboratory, cross-contaminations can be
 519 ruled out. The most likely explanation for these false-positive assignments is the
 520 placement of target fragments in conserved regions to amplify a broader taxonomic
 521 range (e.g., vertebrates). However, this will ultimately decrease the taxonomic
 522 resolution for specific taxa within that group (e.g., fish species). For the here
 523 investigated primer pairs most likely the short fragment length (12SV5 primer pair; 106
 524 bp) or the fragment location for the LH16S primer pair the number of substitutes is too
 525 low for reliable fish identification.

526 Furthermore, incorrect assignments of closely related species were observed for the
 527 more general vertebrate primer pairs 12SV5 and LH16S. These included the Asian
 528 *Chondrostoma prespense* instead of *C. nasus*, the North American *Thymallus arcticus*
 529 instead of *T. thymallus*, or *Pungitius platygaster* instead of *P. pungitius*. Again, the
 530 conserved regions amplified by the 12SV5 and LH16S primer pairs could have led to
 531 these false-positive assignments. Particularly phylogenetically ‘young’ species that
 532 have not been separated long and e.g., share mitochondrial haplotypes (Espanhol et
 533 al. 2007) or closely related species that exhibit hybridisation and introgression (Hata et
 534 al. 2019, De Santis et al. 2021) are potentially not distinguishable with short and
 535 conserved target fragments.

536 However, also the tele02, MiFish-U and SeaDNA-mid primer pairs showed false-
 537 positive assignments. Even though the asp (*Leuciscus aspius*) was not included in the
 538 mock community, it was detected by all three primer pairs. Since it was consistently
 539 detected by the tele02 (2 OTUs, 98% similarity to reference sequence, 8578 reads,
 540 10/10 samples), MiFish-U (2 OTUs, 98%, 7246 reads, 10/10 samples), and the
 541 SeaDNA primer pair (1 OTU, 100%, 156 reads, 9/10 samples), the most likely
 542 explanation for the detection of *L. aspius* is be a misidentification during sampling (e.g.,
 543 another closely related cyprinid species). Another explanation is that the DNA of one
 544 species can be found in the mucus of another species’ mucus, which could potentially
 545 also contain eDNA traces from other fish that were present during sampling. Another
 546 case of false-positive detection is the Japanese huchen (*Parahucho perryi*), which was
 547 detected in low read abundances by the tele02 primer pair (1 OTU, 98%, 114 reads,
 548 9/10 samples). The Japanese huchen is not recorded from Central Europe but is
 549 related to both the huchen (*Hucho hucho*) and brown trout (*Salmo trutta*), which were
 550 both present in the mock community. The most likely explanation is that this false-
 551 positive assignment originates from huchen or brown trout DNA that is amplified by the
 552 tele02 primer pair followed by misassignment. The low read abundance observed in
 553 this dataset and its occurrence in combination with the brown trout in other eDNA
 554 metabarcoding datasets using the tele02 primer pair (Macher et al., unpublished data)
 555 hints towards a systematically false-positive detection of the Japanese huchen in the
 556 presence of the brown trout. A similar case is the detection of the Asian sharp-snouted
 557 lenok (*Brachymystax lenok*) with the MiFish-U primer pair, which is a salmonoid
 558 species related to trouts.

559 While most false-positive assignments can be easily corrected, primer pairs that are
 560 not prone to false-positive assignments, such as the tele02, MiFish-U and the
 561 SeaDNA-mid primer pairs, are to be preferred over the more general 12SV5 and

562 LH16S primer pairs when investigating fish communities based on eDNA
563 metabarcoding.

564 Conclusion

565 In conclusion, our study highlights how the choice of primer has a major effect on the
566 outcome of eDNA metabarcoding analysis. The tele02 primer pair proved to be best
567 choice for eDNA metabarcoding of Central European freshwater fish, showing the
568 highest discriminatory power and good reproducibility with fewest false-positive and
569 false-negative detections of the here tested primer pairs. We also observed that gaps
570 in reference libraries can still lead to false-negative detections and thus should be
571 addressed. Through careful selection of the primer pair, laboratory protocol, and
572 bioinformatic pipeline, eDNA metabarcoding is becoming an increasingly reliable tool
573 for fish monitoring.

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772 **Disclaimer**

773 The collection of mucus samples is not categorized as animal experiment and did not
774 require further authorisation. All sampling events were coordinated with local
775 authorities. Fish specimens were solely caught during sampling events for monitoring
776 campaigns and were handled by experts.

777 **Data accessibility**

778 The raw data were deposited at the European Nucleotide Archive
779 (<https://www.ebi.ac.uk/ena/browser/home>) under the accession number PRJEB60937.

780 **Author contributions**

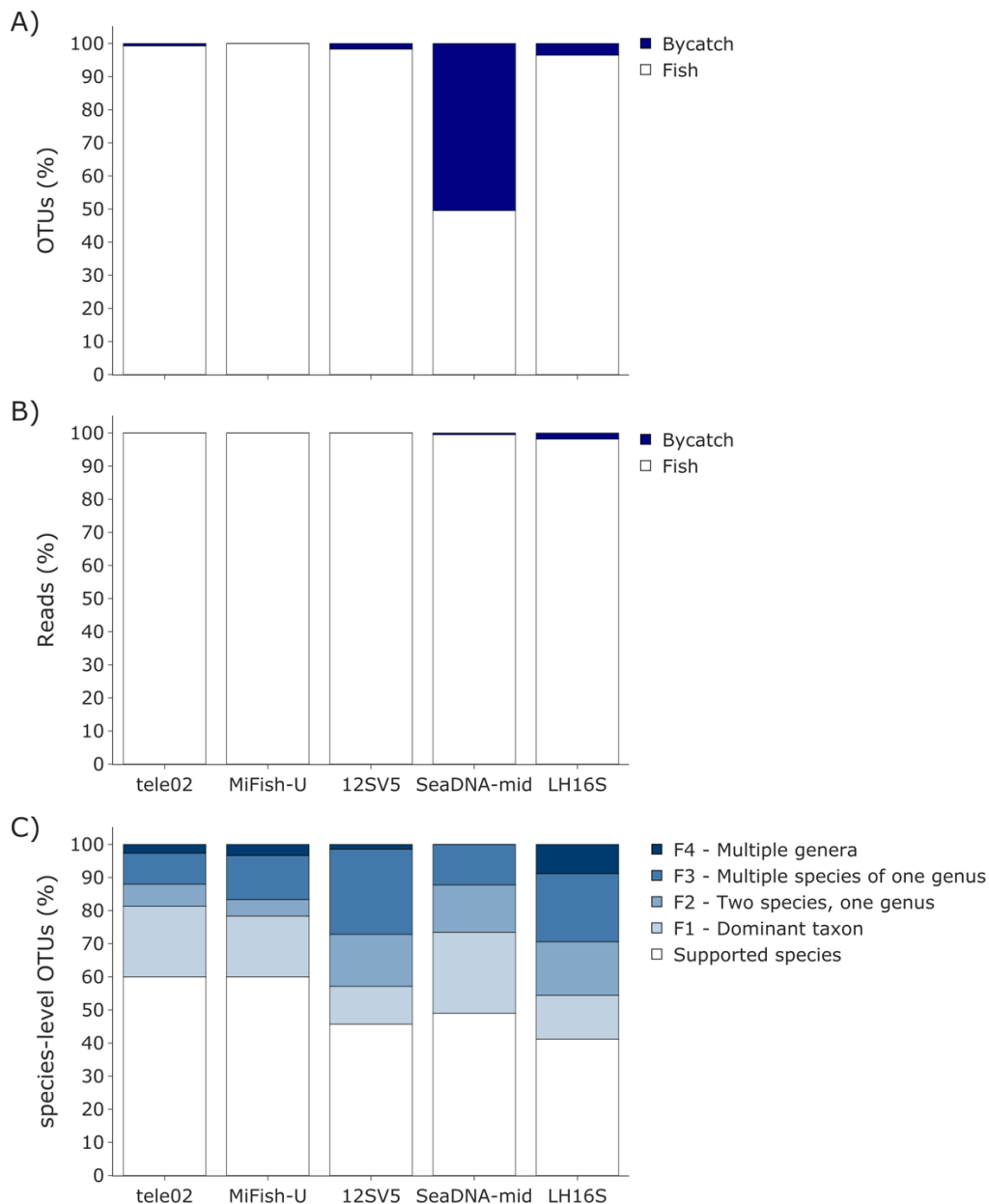
781 **Till-Hendrik Macher:** Conceptualization, Methodology, Formal analysis, Investigation,
782 Visualization, Writing - original draft, Writing - review & editing; **Robin Schütz:**
783 Conceptualization, Methodology, Formal analysis, Investigation, Writing - original draft,
784 Writing - review & editing; **Atakan Yildiz:** Methodology, Writing - review & editing; **Arne**
785 **J. Beermann:** Conceptualization, Validation, Supervision, Writing - review & editing;
786 **Florian Leese:** Conceptualization, Resources, Supervision, Project administration,
787 Funding acquisition, Writing - review & editing.

788 Figures and tables

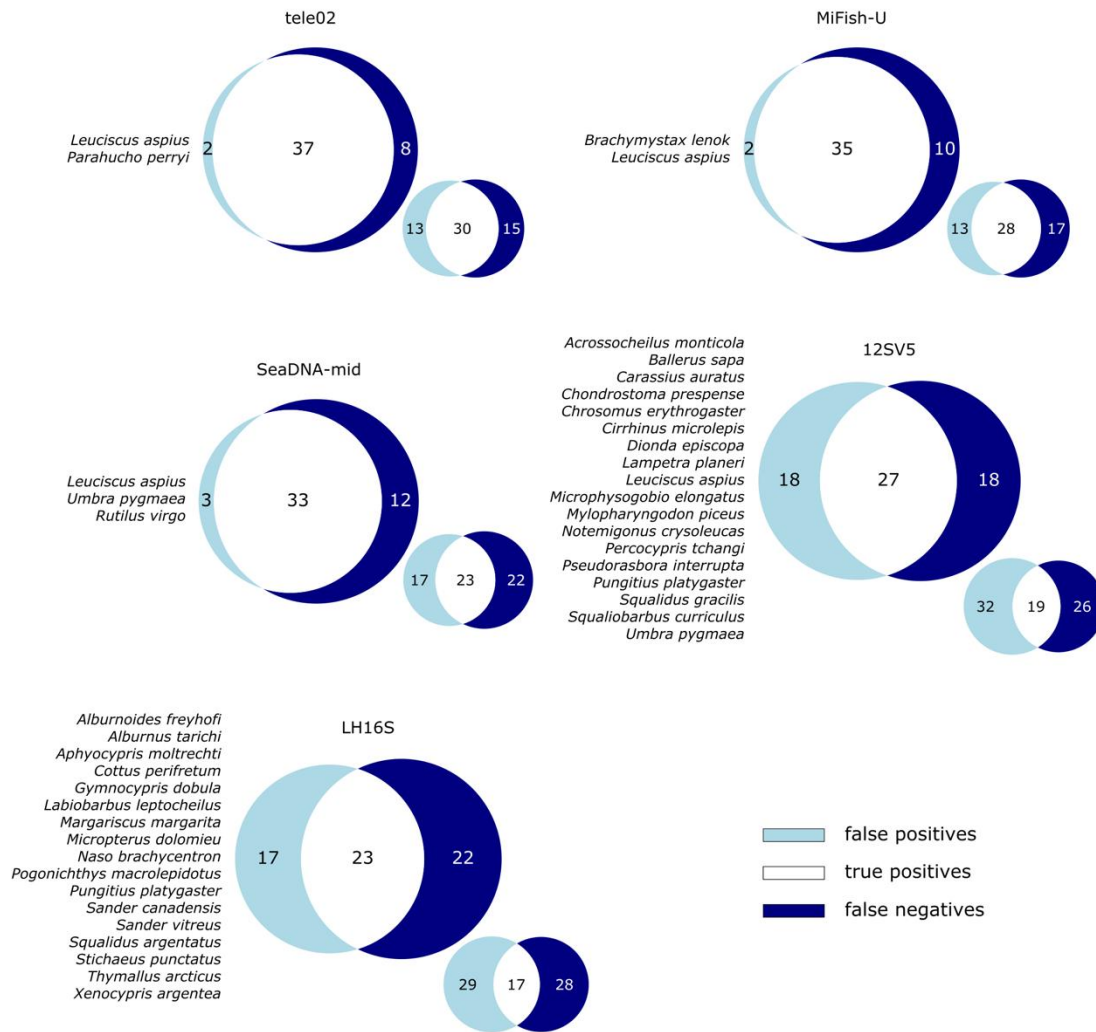
789 **Table 1:** Primer pairs used for PCR amplification of the fish mock community.

Name	Gene	Primer pair	Forward sequence (5'-3')	Reverse sequence (5'-3')	Annealing temp.	Target length	Publication
tele02	12S	tele02_fw/tele02_rv	AAACTCGTGCCAGCCACC	GGGTATCTAATCCCAGTTTG	52 °C	~ 167 bp	Taberlet et al. 2018
MiFish-U	12S	MiFish-U_fw/MiFish-U_rv	GTCGGTAAACTCGTGCCAGC	CATAGTGGGGTATCTAATCCCAGTTTG	59 °C	~ 170 bp	Miya et al. 2015
SeaDNA-mid	COI	coi.175f/coi.345r	GGAGGCTTTGGMAAYTGRYT	TAGAGRRGGGTARACWGTYCA	53 °C	~ 130 bp	Collins et al. 2019
12SV5	12S	12S-V5f/12S-V5r	ACTGGGATTAGATACCCC	TAGAACAGGCTCCTCTAG	52 °C	~ 106 bp	Riaz et al. 2011
LH16S	16S	L2513/H2714	GCCTGTTTACCAAAAACATCAC	CTCCATAGGGTCTTCTCGTCTT	55 °C	~ 220 bp	Kitano et al. 2007

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793 **Figure 1:** Proportions of fish and non-fish OTUs (A) and read proportions (B) detected with the five
794 different primer pairs (A), and the proportions of ambiguous taxonomic assignments (flags 1-4) for
795 all species-level OTUs (C), based on the pre-adjusted datasets.
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 798 **Figure 2:** Comparison of the fish mock community species composition to the detected species
 799 with each primer pair for both the adjusted (large Venn diagrams) and the pre-adjusted datasets
 800 (small Venn diagrams). All species declared as false-positive detections are listed on the left-hand
 801 side of the respective Venn diagram.
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822 **Table 2:** The overall number of fish species and the respective number of OTUs (in brackets) per
 823 family is shown in subplot A) for each primer pair. The number of false-positive (n/) and false-
 824 negative (/n) fish species detections compared to the original fish mock community composition is
 825 presented in subplot B).

Family	A)					B)				
	tele02	MiFish-U	12SV5	SeaDNA-mid	LH16S	tele02	MiFish-U	12SV5	SeaDNA-mid	LH16S
Acanthuridae	0 (0)	0 (0)	0 (0)	0 (0)	1 (2)	0/0	0/0	0/0	0/0	1/0
Acheilognathidae	1 (1)	1 (1)	1 (2)	1 (2)	1 (4)	0/0	0/0	0/0	0/0	0/0
Anguillidae	1 (1)	1 (1)	1 (1)	1 (1)	1 (1)	0/0	0/0	0/0	0/0	0/0
Centrarchidae	0 (0)	0 (0)	0 (0)	0 (0)	1 (1)	0/0	0/0	0/0	0/0	1/0
Cobitidae	1 (1)	1 (1)	0 (2)	1 (1)	0 (1)	0/0	0/0	0/1	0/0	0/1
Cottidae	1 (2)	1 (1)	0 (3)	1 (1)	1 (1)	0/1	0/1	0/2	0/1	1/2
Cyprinidae	3 (8)	3 (6)	7 (12)	3 (4)	4 (15)	0/0	0/0	4/0	0/0	2/1
Esocidae	1 (3)	1 (1)	1 (9)	1 (1)	1 (1)	0/0	0/0	0/0	0/0	0/0
Gasterosteidae	2 (6)	2 (2)	2 (2)	2 (5)	3 (5)	0/0	0/0	1/1	0/0	1/0
Gobiidae	2 (2)	1 (1)	2 (2)	2 (4)	1 (10)	0/0	0/1	0/0	0/0	0/1
Gobiionidae	3 (4)	3 (4)	5 (8)	2 (3)	3 (11)	0/0	0/0	3/1	0/1	1/1
Kyphosidae	0 (0)	0 (0)	0 (1)	0 (0)	0 (0)	0/0	0/0	0/0	0/0	0/0
Leuciscidae	11 (53)	10 (40)	13 (23)	8 (14)	6 (16)	1/2	1/3	6/5	2/6	4/10
Lotidae	1 (1)	1 (1)	1 (1)	1 (1)	1 (1)	0/0	0/0	0/0	0/0	0/0
Nemacheilidae	1 (1)	1 (1)	0 (0)	1 (1)	1 (1)	0/0	0/0	0/1	0/0	0/0
Ophiidiidae	0 (0)	0 (0)	0 (0)	0 (0)	0 (1)	0/0	0/0	0/0	0/0	0/0
Percidae	3 (26)	3 (19)	2 (13)	4 (5)	6 (21)	0/3	0/3	0/4	0/2	2/2
Peristediidae	0 (0)	0 (0)	0 (1)	0 (0)	0 (0)	0/0	0/0	0/0	0/0	0/0
Petromyzontidae	0 (1)	0 (0)	1 (1)	0 (1)	0 (1)	0/1	0/1	1/1	0/1	0/1
Pholidae	0 (0)	0 (0)	0 (0)	0 (0)	0 (1)	0/0	0/0	0/0	0/0	0/0
Salmonidae	5 (10)	5 (6)	3 (9)	4 (5)	3 (12)	1/0	1/0	0/1	0/0	1/2
Siluridae	1 (1)	1 (1)	1 (1)	1 (1)	1 (1)	0/0	0/0	0/0	0/0	0/0
Stichaeidae	0 (0)	0 (0)	0 (0)	0 (0)	1 (1)	0/0	0/0	0/0	0/0	1/0
Tincidae	1 (6)	1 (8)	1 (7)	1 (2)	1 (6)	0/0	0/0	0/0	0/0	0/0
Umbridae	0 (1)	0 (1)	1 (3)	1 (1)	0 (2)	0/1	0/1	1/1	1/1	0/1
Xenocyprididae	1 (3)	1 (4)	3 (5)	1 (1)	3 (7)	0/0	0/0	2/0	0/0	2/0
Zoarcidae	0 (0)	0 (0)	0 (0)	0 (0)	0 (1)	0/0	0/0	0/0	0/0	0/0

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828 **Table 3:** Relative read abundances (%) for all detected fish and lamprey species of all five primer
 829 pairs, including all species present in the mock community (i.e., true positive species, A) and all
 830 non-target species (i.e., false-positive species, B). For each species the number of positive
 831 detections (occurrences) and the standard deviation (STDEV) were calculated.

A)

True positive species	tele02	MIFish-U	12SV5	SeaDNA-mid	LH16S	Occurrences	STDEV
<i>Anguilla anguilla</i>	0.103	0.095	0.119	0.002	0.147	5	0.1
<i>Silurus glanis</i>	0.129	0.02	0.164	0.15	0.064	5	0.1
<i>Barbus barbus</i>	0.306	0.341	0.544	0.775	0.635	5	0.2
<i>Thymallus thymallus</i>	0.466	0.507	0.731	0.033	1.208	5	0.4
<i>Tinca tinca</i>	3.682	4.099	4.393	4.852	4.523	5	0.4
<i>Gymnocephalus cernua</i>	2.243	2.241	2.37	0.357	1.108	5	0.9
<i>Ctenopharyngodon idella</i>	1.15	0.153	1.39	2.576	0.667	5	0.9
<i>Lota lota</i>	0.264	0.253	0.329	3.746	0.191	5	1.6
<i>Gobio gobio</i>	0.179	0.192	0.24	3.816	0.141	5	1.6
<i>Rhodeus sericeus/amarus</i>	1.46	1.412	1.544	5.838	0.73	5	2.1
<i>Carassius carassius</i>	2.366	2.539	3.167	7.149	2.121	5	2.1
<i>Rutilus rutilus</i>	2.502	2.558	0.353	1.746	6.3	5	2.2
<i>Esox lucius</i>	6.232	5.7	5.131	0.049	0.432	5	3.0
<i>Perca fluviatilis</i>	1.485	1.321	12.65	0.18	1.174	5	5.2
<i>Proterorhinus semilunaris</i>	0.267	0.007	0.007	7.917	11.719	5	5.5
<i>Hucho hucho</i>	2.067	2.288	2.494	17.357	3.042	5	6.7
<i>Pungitius pungitius</i>	0.491	0.58	0.677	17.491	0.839	5	7.5
<i>Phoxinus phoxinus</i>	0.003	0.002	0.003	0	0.005	4	0.0
<i>Barbatula barbatula</i>	0.003	0.004	0	0.006	0.002	4	0.0
<i>Oncorhynchus mykiss</i>	0.149	0.192	0.294	0.54	0	4	0.2
<i>Cyprinus carpio</i>	0.372	0.363	0.506	0.019	0	4	0.2
<i>Gasterosteus aculeatus</i>	0.197	0.218	0	0.732	0.25	4	0.3
<i>Squalius cephalus</i>	1.072	1.09	0.09	0.007	0	4	0.6
<i>Pseudorasbora parva</i>	0.214	0.212	0	1.972	0.206	4	0.9
<i>Chondrostoma nasus</i>	2.333	2.508	0.047	0.408	0	4	1.3
<i>Blicca bjoerkna</i>	5.043	5.176	0.161	1.269	0	4	2.6
<i>Abramis brama</i>	19.839	19.776	25.78	17.748	0	4	3.5
<i>Alburnus alburnus</i>	6.532	6.837	10.588	1.189	0	4	3.9
<i>Sander lucioperca</i>	8.895	8.623	0	0.054	3.856	4	4.2
<i>Romanogobio albipinnatus</i>	0.003	0.005	0.006	0	0	3	0.0
<i>Neogobius melanostomus</i>	0.003	0	0.004	0.026	0	3	0.0
<i>Salmo trutta</i>	0.01	0.01	0	0.2	0	3	0.1
<i>Misgurnus fossilis</i>	0.039	0.038	0	0.821	0	3	0.5
<i>Cottus rhenanus</i>	1.065	1.029	0	0.005	0	3	0.6
<i>Leucaspis delineatus</i>	0.002	0.001	0	0	0	2	0.0
<i>Zingel streber</i>	0	0	0	0.158	0.333	2	0.1
<i>Leuciscus idus</i>	0	0.131	0	0	0	1	
<i>Leuciscus leuciscus</i>	0.053	0	0	0	0	1	
<i>Scardinius erythrophthalmus</i>	0.07	0	0	0	0	1	
<i>Cottus gobio</i>	0	0	0	0	0	0	
<i>Gymnocephalus schraetser</i>	0	0	0	0	0	0	
<i>Lampetra fluviatilis</i>	0	0	0	0	0	0	
<i>Rutilus pigus</i>	0	0	0	0	0	0	
<i>Umbra krameri</i>	0	0	0	0	0	0	
<i>Zingel zingel</i>	0	0	0	0	0	0	

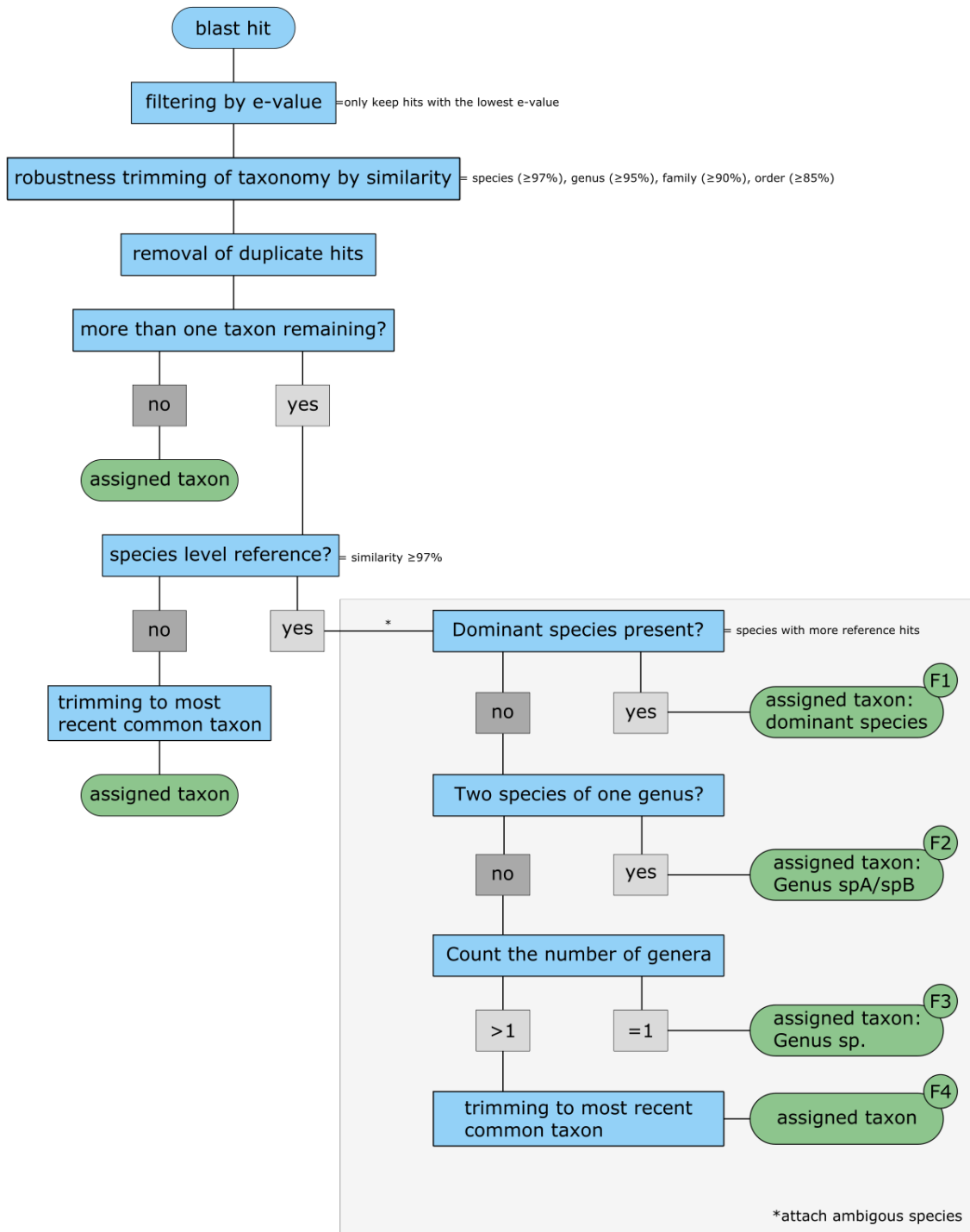
B)

False positive species	tele02	MIFish-U	12SV5	SeaDNA-mid	LH16S	Occurrences	STDEV
<i>Leuciscus aspius</i>	0.512	0.556	0.091	0.016	0	4	0.3
<i>Pungitius platygaster</i>	0	0	0.008	0	0.002	2	0.0
<i>Umbra pygmaea</i>	0	0	1.002	0.005	0	2	0.7
<i>Acrossocheilus monticola</i>	0	0	0.013	0	0	1	
<i>Alburnoides freyhofii</i>	0	0	0	0	37.686	1	
<i>Alburnus tarichi</i>	0	0	0	0	0.146	1	
<i>Aphocypris molrechti</i>	0	0	0	0	0.35	1	
<i>Ballerus sapa</i>	0	0	0.922	0	0	1	
<i>Brachymystax lenok</i>	0	0.005	0	0	0	1	
<i>Carassius auratus</i>	0	0	0.031	0	0	1	
<i>Chondrostoma prespense</i>	0	0	2.083	0	0	1	
<i>Chrosomus erythrogaster</i>	0	0	0.028	0	0	1	
<i>Cirrhinus microlepis</i>	0	0	0.017	0	0	1	
<i>Cottus perflretum</i>	0	0	0	0	0.964	1	
<i>Dionda episcopa</i>	0	0	0.118	0	0	1	
<i>Gymnocypis dobula</i>	0	0	0	0	0.05	1	
<i>Labiobarbus leptocheilus</i>	0	0	0	0	0.02	1	
<i>Lampetra planeri</i>	0	0	0.004	0	0	1	
<i>Margariscus margarita</i>	0	0	0	0	0.24	1	
<i>Microphysogobio elongatus</i>	0	0	0.007	0	0	1	
<i>Micropterus dolomieu</i>	0	0	0	0	0.284	1	
<i>Mylopharyngodon piceus</i>	0	0	0.057	0	0	1	
<i>Naso brachycentron</i>	0	0	0	0	0.687	1	
<i>Notemigonus crysoleucas</i>	0	0	0.159	0	0	1	
<i>Parahucho perryi</i>	0.007	0	0	0	0	1	
<i>Percocypris tchangii</i>	0	0	0.046	0	0	1	
<i>Pogonichthys macrolepidotus</i>	0	0	0	0	0.226	1	
<i>Pseudorasbora interrupta</i>	0	0	0.393	0	0	1	
<i>Rutilus virgo</i>	0	0	0	0.522	0	1	
<i>Sander canadensis</i>	0	0	0	0	0.01	1	
<i>Sander vitreus</i>	0	0	0	0	0.059	1	
<i>Squalidus argentatus</i>	0	0	0	0	0.034	1	
<i>Squalidus gracilis</i>	0	0	0.091	0	0	1	
<i>Squaliobarbus curriculus</i>	0	0	0.071	0	0	1	
<i>Stichaeus punctatus</i>	0	0	0	0	14.238	1	
<i>Thymallus arcticus</i>	0	0	0	0	0.157	1	
<i>Xenocypris argentea</i>	0	0	0	0	0.047	1	

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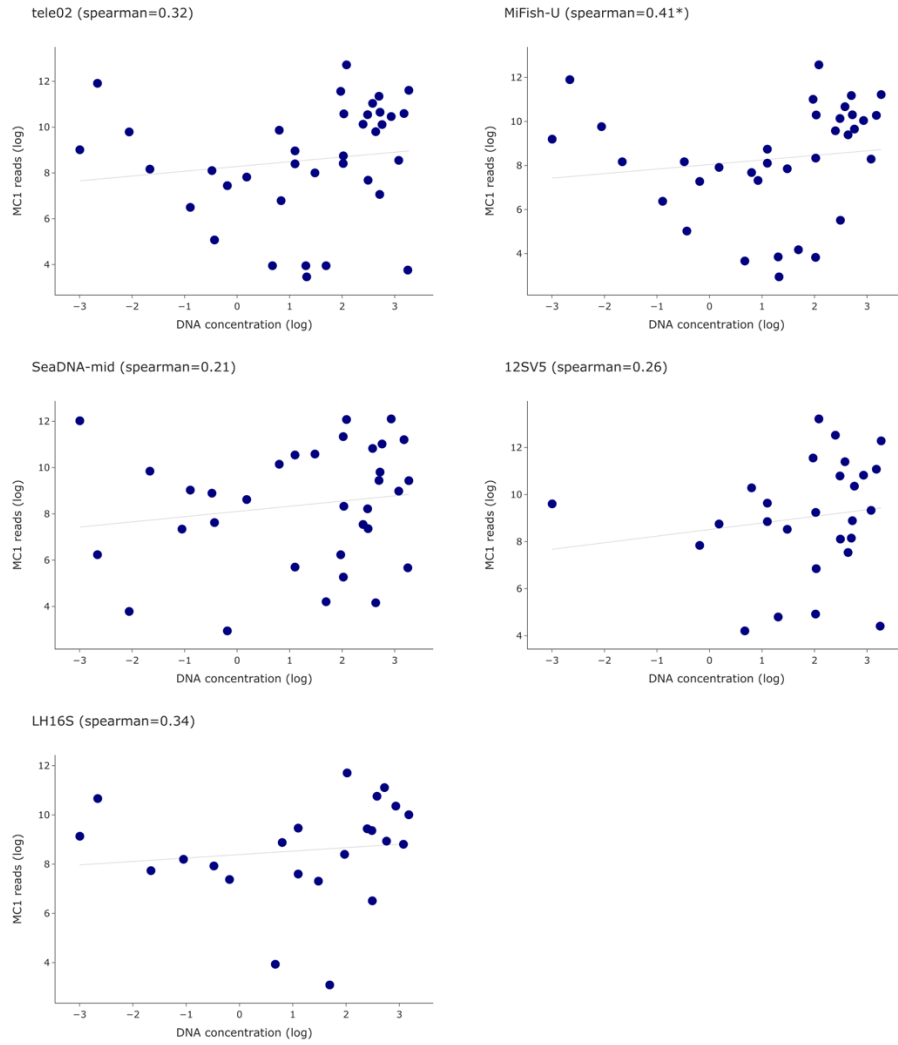
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Supplementary figures and tables



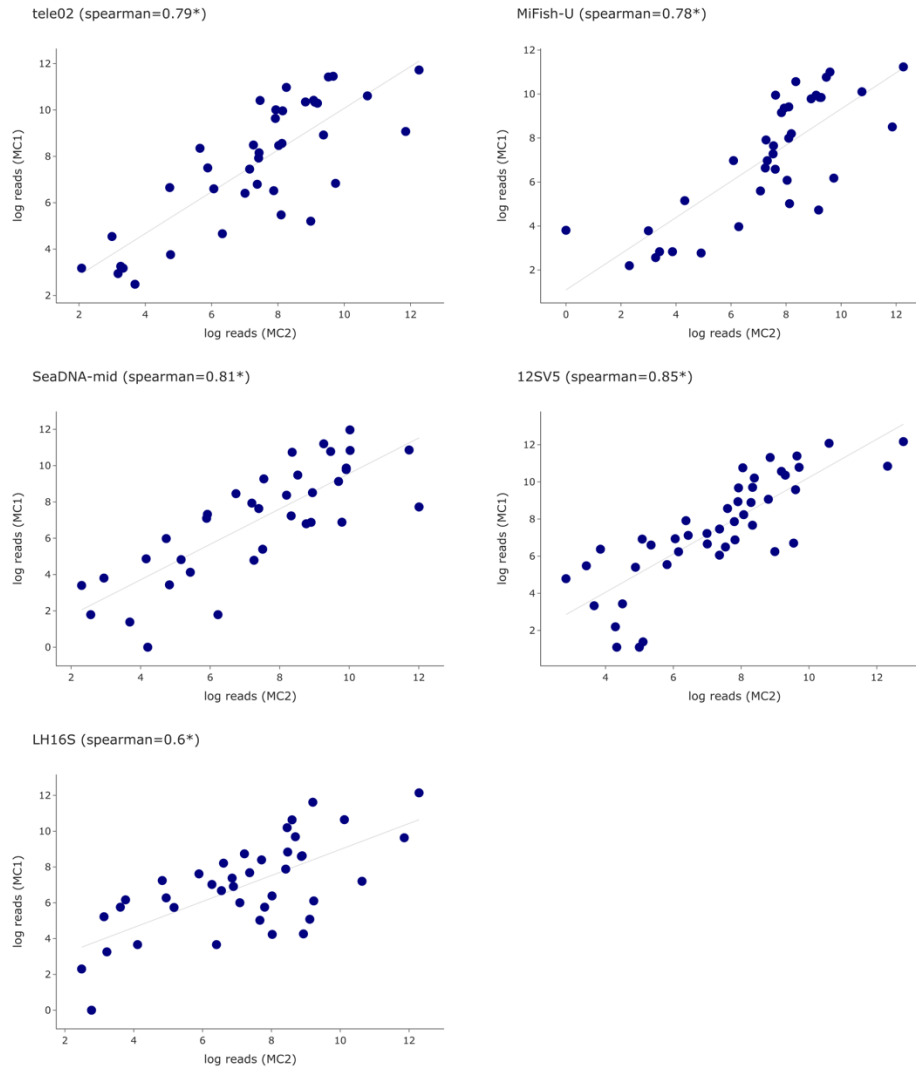
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Supplementary figure 1: Decision tree for taxonomic assignment implemented in APSCALE v1.2.0.



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Supplementary figure 2: Pairwise comparison of the log-transformed reads of the non-normalized mock community (MC1) compared to the DNA concentration of each species.



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847 **Supplementary figure 3:** Pairwise comparison of the log-transformed reads of the non-normalized
848 mock community (MC1) compared to log-transformed reads of the normalized mock community
849 (MC2) of each species.

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851 **Supplementary table 1:** Sampled specimens and their respective species assignment collected
852 for the fish mock community, extraction date, collection site, and concentration after DNA
853 extraction.

854
855 **Supplementary table 2:** List of all species reported from Germany, their occurrence status, and
856 their presence in the mock community (data from fishbase.org).

857
858 **Supplementary table 3:** List of all ambiguous assignments.

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860 **Supplementary table 4:** List of over splitting rates per primer pair for each detected species.

861

862 **Supplementary material**

863 **Supplementary material 1:** Protocol for the adapted NucleoMag Tissue Kit.

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865 **Supplementary material 2:** Unmodified TaXon tables of each primer pair.

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867 **Supplementary material 3:** Processed TaXon tables of each primer pair (subtracted negative
868 controls and filtered for fish and lamprey taxa OTUs).
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870 **Supplementary material 4:** Processed and manually curated TaXon tables of each primer pair.
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