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***Halyomorpha halys* invasion front jumps 1,500 kilometres to reach the Canary Islands; a framework for rapid response, identification of urgent questions, and assessment of potential impacts**

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1 ***Halyomorpha halys* invasion front jumps 1,500 kilometres to reach the**  
2 **Canary Islands; a framework for rapid response, identification of urgent**  
3 **questions, and assessment of potential impacts**

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11 trapping

12

13 **Abstract**

14 Early detection and rapid response are cornerstones of effective invasive species  
15 management. However, these strategies can be challenging to implement when the arrival  
16 of a non-native species has not been predicted, as may be the case when a species is  
17 discovered large distances from known populations. Brown marmorated stink bugs  
18 *Halyomorpha halys* are rapidly spreading across much of the world, causing substantial  
19 economic losses to agriculture as well as nuisance when entering houses to overwinter  
20 communally. Multiple individuals were recently confirmed in the Canary archipelago,  
21 marking a dramatic range expansion into subtropical Atlantic areas and the northwest  
22 African region. The potential establishment of this major pest species in this region raises  
23 important questions on its ecological adaptation to new climatic and biotic conditions as  
24 well as the unknown impacts and control effectiveness in novel host crop plants such as  
25 banana. Previous attempts to control *H. halys* elsewhere have been typically only partially  
26 successful but we suggest eradication in the Canary Islands might still be possible and a  
27 sensible goal. Within a conceptual framework we review potential management options and  
28 encourage local authorities and stakeholders to implement specific surveillance, control and  
29 biosecurity measures aiming to swiftly eradicate this species. Not doing so risks a potentially  
30 severe invasion by this species across the region and significant damage to the local  
31 agricultural and ornamental plant economy. Our framework provides a basis for rapid

32 response and management in other scenarios where an unexpected non-native species is  
33 detected.

34

## 35 **Introduction**

36

37 It is well-established that a comprehensive approach to invasive species management  
38 involves early detection and rapid response (Reaser et al., 2020). Development of such a  
39 strategy requires effective predictions of future invaders so that appropriate, species-  
40 specific measures for surveillance and management can be implemented. For invasive  
41 species that are spreading along known pathways or by recognised vectors, horizon-  
42 scanning can be an effective tool in identifying and prioritising future invaders (e.g. Roy et  
43 al., 2014). However, when non-native species are transported large distances from historical  
44 locations the stakeholders in recipient geographies may be ill-prepared for species that have  
45 been overlooked as potential threats. Here we report the unexpected discovery of a high  
46 risk invader 1,500 km away from its previously known distribution, and set-out a framework  
47 for rapid response and management that could be applied under similar scenarios for other  
48 discoveries of unpredicted non-native species.

49 The brown marmorated stink bug *Halyomorpha halys* (Stål, 1855) is a highly polyphagous,  
50 mobile, and damaging invasive hemipteran of global concern that has been spreading in the  
51 past two decades from its native east and south Asia to much of the USA, parts of South  
52 America (Chile) and large parts of continental Europe, from Russia and Georgia in the east to  
53 Portugal in the west, as reviewed by Kriticos et al. (2017) and Leskey & Neilsen (2018). In  
54 southern Europe, including Spain where it has been recorded in the Catalunya region since  
55 2016 (Dioli et al. 2018), the species is bivoltine and occupies farmland and Mediterranean  
56 forested habitats. In such mild climate areas rapid increases in abundance and sometimes  
57 extensive crop damage have been observed within less than five years post-establishment  
58 (Bariselli et al. 2016; Costi et al. 2017; Pajač Živković et al. 2021).

59 *H. halys* has an extensive host range exceeding 300 plant species and can severely impact a  
60 large variety of fruit crops, row crops, vegetables and ornamental plants (Bergmann et al.  
61 2016; Kriticos et al. 2017; Leskey & Neilsen 2018). For example, in Italy the species has  
62 locally caused up to 50% or even 80% fruit crop losses in some sites (Costi et al. 2017;  
63 Maistrello et al. 2017). Damage is caused via feeding injuries and injecting digestive

64 enzymes that produce necrotic plant tissue, wilting and deformities, which is further  
65 aggravated by the feeding behaviour of adults that move frequently between individual  
66 plants and different crop types (Bariselli et al. 2016). It can also transmit plant pathogens,  
67 such as witches' broom disease *Paulownia tomentosa* (Kriticos et al. 2017). In addition, its  
68 tendency to form extensive overwintering aggregations in buildings and the pungent,  
69 unpleasant smell produced when disturbed, makes it a significant urban nuisance pest (Lee  
70 et al. 2013). It has shown multi-pesticide resistance and, once established, farmers have had  
71 to substantially increase insecticide applications, for example four-fold increase in the  
72 aftermath of *H. halys* outbreaks in USA apple orchards (Leskey et al. 2012), causing  
73 supplementary chemical pollution, higher production costs and damaging integrated pest  
74 management, especially as most insecticides available to growers for *H. halys* control are  
75 broad-spectrum (Kuhar and Kamminga, 2017).

76 Under controlled conditions average *H. halys* egg mortality was 100% at 12.5 °C but  
77 dropped to 1.3% at 30 °C (Haye et al. 2014) and optimal development temperatures for *H.*  
78 *halys* have been estimated at 25-30°C, with maximum feeding at 16-17°C, meaning large  
79 parts of the Southern Hemisphere, Central America, northwest and sub-Saharan Africa are  
80 considered suitable for the species (Kriticos et al. 2017; Leskey et al. 2018). However,  
81 despite its high invasive potential in many subtropical and tropical regions, where it could  
82 exhibit continuous activity throughout the year and have even shorter generation time the  
83 species has so far remained largely absent in such areas (Leskey et al. 2018). In the warm  
84 climate of its native southern China the species was reportedly able to produce up to 4-6  
85 annual generations in by Hoffmann (1931) but more generally it is known to produce 1-2  
86 generations per year (Leskey & Neilsen, 2018).

87 In December 2021 multiple individuals of *H. halys* were detected in an urban setting on the  
88 island of Tenerife (Figure 1). This marks a southern expansion of around 1,500 km, well  
89 beyond the range of self-sustained flight from the closest known existing populations in the  
90 Iberian Peninsula, and into subtropical areas. The strong Atlantic influence in the Canary  
91 Islands' climate results in an "eternal spring", nearly constant 18-24°C temperature at sea  
92 level but with wide differences in rainfall between, and within, islands due to complex island  
93 morphologies. Large parts of the eastern islands of Lanzarote and Fuerteventura might be  
94 too hot and dry for *H. halys* development, but the areas used for agriculture across the  
95 islands are likely suitable. It is unknown how the lack of a defined cold season may influence

96 diapause in *H. halys* as the species was shown to breed late into autumn in southern Europe  
97 but photoperiod, which is assumed to be a critical diapause cue, might delay development  
98 and female reproductive maturity at such low latitudes (Leskey et al. 2018).

99 These *H. halys* observations in the Canary Islands are very recent and to date individuals  
100 have been found only in a very localised site, indicating the possibility of early-stage  
101 invasion. The aim of this paper is to discuss this range expansion, the potential risk to  
102 Tenerife and the Canary Islands, identify urgent research questions, discuss possible  
103 management opportunities and consider how this example may provide a framework for  
104 rapid responses to discoveries of other invasive species a long distance from previous  
105 records.

## 106 **Methods**

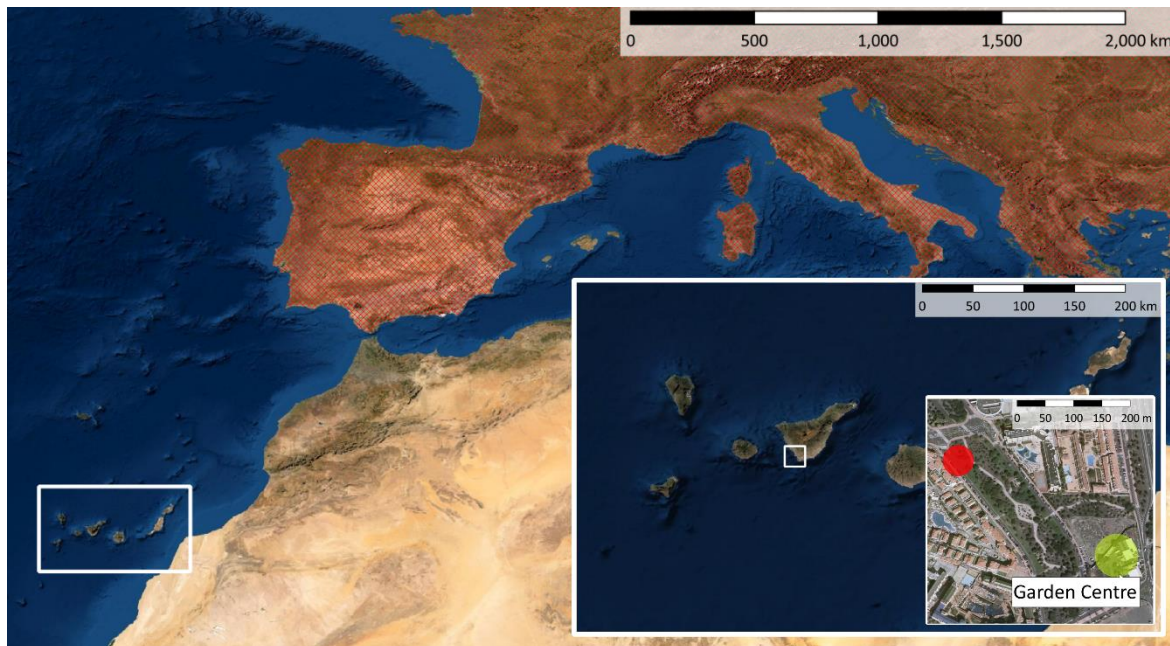
107 A live and active adult brown marmorated stink bug was observed on 17<sup>th</sup> of December  
108 2021 on the external wall of a building in the coastal town of Costa Adeje in south Tenerife  
109 (Figure 1) and flew away while taking pictures to confirm species identification. Once the  
110 invasive data range was checked from citizen science records and the scientific literature,  
111 the images and location were made public that day on INaturalist to alert interested  
112 scientists. Following expert confirmations that this is probably the first record of this species  
113 in the Canary Islands, repeated transect surveys were started the following day and the local  
114 government health plant agency was alerted.

115 An area of approximately 400 m in length around the site of the first observation was  
116 surveyed twice daily, at 8-9am and at 8-10pm until 24<sup>th</sup> of December. Additionally, visual  
117 sampling in the nearby area using 1-2 min observations of 50 randomly chosen trees and  
118 shrubs was started by inspecting leaves, flowers, fruits and woody parts between 0.5 and  
119 2 m above the ground, as well as searches of neighbouring built-up areas (e.g. garden and  
120 building walls with similar surfaces and orientations within a 1 km radius). An additional 40  
121 ornamental street-lining trees, mostly palms and flamboyant trees *Delonix regia*, were  
122 visually checked following the same protocol in three other nearby urban sites (Adeje, La  
123 Caleta and Los Cristianos), at distances of 2-6 km from the original site.

## 124 **Results**

125 During repeated surveys an additional five individuals were located by the 24<sup>th</sup> of December  
126 2021. All individuals were found at the same site, on eastern facing external building walls  
127 and open-sided corridors bordering a public park with palm trees and mature shrubs. One

128 individual was observed landing on the external wall on the third floor of a building and  
129 then captured. Most individuals (4) were recorded on the same day (December 19<sup>th</sup>); three  
130 were recently dead, apparently stepped on by people on the external paved surface.  
131 All six recorded *H. halys* individuals were adults, with three females and at least one male  
132 provisionally identified based on external morphology of the ventral area. Two live and two  
133 freshly dead individuals were collected as record specimens and to allow genetic analysis to  
134 establish their probable origin. These specimens were passed on to the Plant Health agency.  
135 Local gardening and cleaning personnel working at the site of the first observations  
136 indicated they had seen stink bugs several times recently in the same place but could not  
137 say if this had been going on for more than 2-3 weeks. However, another Pentatomidae  
138 species, which is similar in shape and size but not in colour (an adult of the Canary Islands  
139 endemic *Acrosternum rubescens*) was observed live at the same location during surveys and  
140 staff could not specify which of these species they had seen. Despite a generally low  
141 abundance of insects in the area, several live moths (*Palpita vitrealis* and *Herpetogramma*  
142 *licarsisalis*) and the syrphid *Eristalinus taeniops* were also found during visual surveys on the  
143 same building walls, suggesting an influx of flying insects at the site, perhaps due to  
144 prevalent wind direction and the position, facing the park.  
145 The vegetation and building wall searches within the 1km radius and the randomised 40  
146 trees checked in other urban areas did not produce any further *H. halys*. The park at the  
147 original site was briefly searched for *H. halys* adults or nymphs on December 21<sup>st</sup> and  
148 focusing on trees and palm trees, but none were found.



149  
 150 Figure 1. Overall area highlighting the position of the Canary Islands, the location of the  
 151 records in SW Tenerife and a detailed map of the *H. halys* observations (red dot) plus the  
 152 proximity to a garden centre (green dot). Shaded areas in continental Europe show occupied  
 153 range by *Halyomorpha halys* at national scale. Shading of some Mediterranean islands (e.g.  
 154 Corsica and Sardinia) indicates known occupied range of the species.

155  
 156

157 **Discussion**

158 The identified *H. halys* individuals might have congregated from multiple sites or from  
 159 considerable distances. However, a mixed open-air and enclosed garden centre located at  
 160 the edge of the park (Figure 1) and selling a variety of cut flowers, potted garden plants and  
 161 garden décor could be a potential point of accidental introduction for the species via  
 162 undetected hitchhiking on live plant imports or garden objects. The garden centre was not  
 163 surveyed but its potential relevance was raised with the Plant Health agency.

164 We suggest these highly clustered observations should be the urgent focus of rapid  
 165 response management and we offer a potential framework for consideration.

166

167 **Significance of *H. halys* in the Canary Islands and potential for impact and spread**

168 Tenerife is the largest (2,034 km<sup>2</sup>) and highest (3,718 m asl) of the Canary Islands, located  
 169 near the centre of the archipelago, which would make *H. halys* expansion across the region

170 feasible given the combination of the relatively small distances between the seven Canary  
 171 Islands (15-100 km) and that it can fly up to 100km (Wiman et al. 2015). Human-assisted  
 172 colonisations between the islands should be expected if the species becomes locally  
 173 established due to the frequent daily movements of goods, vehicles and people between  
 174 the islands. While the distance to the African mainland is around 100km from the island of  
 175 Fuerteventura, there are also discussions to reinstate the currently suspended direct  
 176 ferryboat to Tarfaya in Morocco.

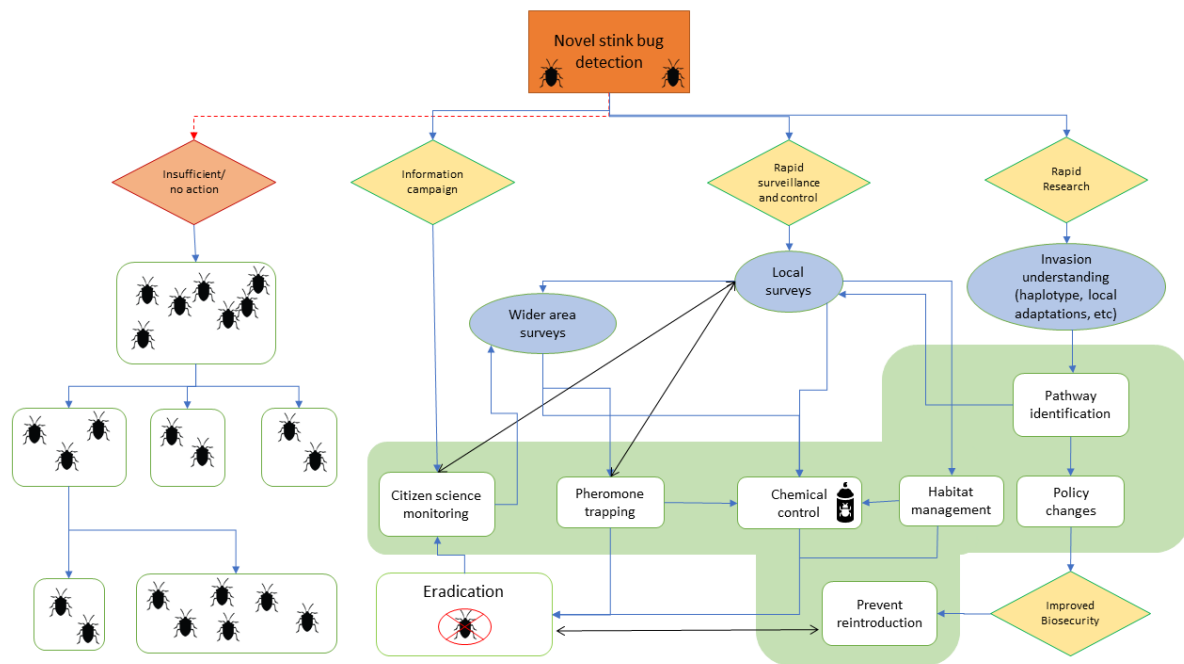
177 Tourism is the largest economic activity in the Canary Islands but agriculture plays an  
 178 important economic, social and cultural role, with an estimated value of 674 million euros in  
 179 2019, especially from bananas (9.095 ha and 420.144 tones in 2020), tomatoes, potatoes,  
 180 avocado and fruits but there are also 6,000 ha of vineyards and over 500 ha of ornamental  
 181 plants and flower nurseries (Estadística Agraria y Pesquera de Canarias, 2019; Gobierno de  
 182 España, 2020), all of which could be impacted by *H. halys*. In addition to cumulative impacts  
 183 of feeding on grapes at different stages, *H. halys* can also affect wine quality by releasing its  
 184 chemical defence compounds if collected with grapes (Mohecar et al. 2017).

185 Invasive species management is often difficult after entry and establishment. Many  
 186 management strategies focus on temporary and localised control, and models to assess  
 187 optimal strategies decisions to eradicate or to control are now available (Baker & Bode  
 188 2021). Yet, the current invasion in Tenerife might offer the potential for eradication, given  
 189 that all individual *H. halys* were found in a single urban location in the vicinity of a potential  
 190 importation point and not detected elsewhere. Urgent and targeted control measures  
 191 would allow a realistic chance for successful eradication.

### 192 **Early detection and rapid response**

193 *H. halys* was apparently eradicated in Australia (Horwood et al. 2019) and this demonstrates  
 194 that a combination of early detection and rapid response can prove effective for this high-  
 195 risk species. We present a conceptual framework (Figure 2) and briefly review the evidence  
 196 for possible rapid response measures for *H. halys* invasion in Tenerife. Early detection refers  
 197 to the process of “surveying for, reporting, and verifying the presence of a non-native  
 198 species before it becomes established or spreads so widely that eradication is no longer  
 199 feasible” and rapid response as measures aiming to eradicate the non-native founding  
 200 population from a specific location (Reaser et al. 2020). Details for how each of the key  
 201 response options in Figure 2 can be implemented in Tenerife are given below.





202

203 Figure 2 Flowchart of rapid response measures and scenarios to address the early detection  
 204 of multiple brown marmorated stink bugs in an isolated new region.

205

206

207 **1) Rapid surveillance** Professional local surveys using visual checks and trapping (see below)  
 208 should progressively increase the area of surveys from a 1km radius. Optimal allocation of  
 209 survey effort can be calculated as a function of probability of occurrence (higher in the  
 210 proximity of the initial detection and decreasing thereafter), the benefits of detection  
 211 (maximum in the early stage when eradication is still a realistic goal) and local detectability of  
 212 the invasive species (depending on life stage, variability in seasonal activity, habitat type and  
 213 detection method used) (Hauser & Raut, 2017). Efforts should be concentrated in the areas  
 214 with the highest likelihood of stink bug arrival and establishment as well as microhabitat  
 215 considered optimal and suitable (e.g. specific or general plant and tree vegetation, specific  
 216 artificial structures such as those involved in the original detections). Advances in  
 217 technology including artificial intelligence, “web crawlers”, computer vision and eDNA (e.g.  
 218 via fruit rinsing and water testing) could significantly improve the surveillance of populations  
 219 (Martinez et al 2020; Valentin et al 2018). Combined surveys and a public information  
 220 campaign (see below) could help to determine invasion status, whether the individuals  
 221 originated from a recent accidental introduction, whether breeding has occurred, how close

222 their distribution is to a means of arrival (e.g. a garden centre) or if the invasion has been  
223 undetected for some time and populations are now established.

224 **2) Information campaign** Professional surveys could be supplemented with a multi-media  
225 public information campaign asking people at the original locality as well as across Tenerife  
226 and other Canary Islands to check for, and report sightings of, this species (Pawson et al,  
227 2020). A smart phone reporting app could be considered but any type of verifiable reporting  
228 should be encouraged and facilitated (e.g. social media, email form, phone message  
229 platform). Free existing platforms, such as INaturalist, could be used and promoted with  
230 sightings monitored and verified by the authorities. *H. halys* is an excellent candidate for  
231 citizen science monitoring being relatively recognisable, slow moving and easy to  
232 photograph: the first European observation originated from hobby photographs in Zurich,  
233 Switzerland in 2004 (Haye et al. 2015). Adult *H. halys* can disperse into crops in time for  
234 reproduction, but trees and shrubs may be important intermediate hosts before crop  
235 invasion (Rice et al. 2014), thus allowing the collection of chance citizen science records at  
236 sites where they might be otherwise difficult to detect.

237 **3) Rapid research** could a) confirm the haplotype structure of individuals collected and b)  
238 investigate potential for local adaptation and spread (e.g. predict preference for particular  
239 areas or plant species, seasonal changes in distribution according to the local climatic  
240 conditions and the presence of habitat niches such as irrigated crop areas). Research could  
241 explore distribution modelling scenarios for this species in the Canary Islands to facilitate  
242 surveillance efforts and model potential spread in the absence of effective control. In  
243 addition, eradication can be costly to implement and is only a viable option if the likelihood  
244 of reintroduction is low. Otherwise, a different strategy may look at keeping populations low  
245 or focus on containment as this can still offer ecological and economic benefits, but this  
246 requires rapid research at an early stage. Identification and management of pathways might  
247 reduce the likelihood of reinvasion, in which case eradication can become more compelling.  
248 Haplotype analysis can provide important information and, at a global level, has indicated  
249 multiple introductions from China between eastern and western USA as well as southern  
250 Europe while the populations in Emilia Romagna in Italy resulted from a bridgehead invasion  
251 effect, introduced from eastern USA rather than directly from China (Valentin et al. 2017;  
252 Leskey & Neilsen 2018). This could assist pathway identification and, combined with import

253 data, could allow targeting potential introduction points for surveillance effort and chemical  
 254 control and, in the longer term, revising of policies and biosecurity protocols (Figure 2).

255 **4) Trapping, chemical control and habitat management** The feasibility of successful  
 256 eradication depends upon characteristics of the invading species and the context in which it  
 257 is detected, with a narrow window of opportunity for eradication (Simberloff et al. 2003).

258 Capture and removal of any *H. halys* present is essential for eradication and pheromone  
 259 trapping of the original site as well as other potential sites can offer crucial support.

260 Pheromone luring and trapping has been intensively researched in the USA for this species  
 261 and substantial progress was achieved once Khrimian et al. (2014) identified the male-  
 262 produced aggregation pheromone following isolation of male-specific volatiles. Single males  
 263 were found to produce more pheromone than groups and both adults and nymphs are  
 264 attracted to pheromone. A variety of trap designs and pheromone lures have recently been  
 265 tested for *H. halys* including various designs, colours and sizes of tree canopy and ground  
 266 deployed traps, with ground-deployed black pyramid traps becoming the standard due to  
 267 their higher capture rates (Leskey et al. 2021). Increasing trap loading with 2-stage  
 268 aggregation pheromones increased overall captures, with variability of optimisation possible  
 269 when considering both specific pheromone attraction and production costs (Leskey et al.  
 270 2021). However, *H. halys* are sometimes attracted to, but remain outside and in the vicinity  
 271 of, the traps baited with pheromonal stimuli, indicating “trap spillage” and a need for  
 272 careful and frequent searches. In Australia, rapid response to identification of *H. halys*  
 273 within and on the perimeter of importation warehouses used insecticide fumigations of the  
 274 interior (bifenthrin surface spray and pyrethrum fog), sweep-net sampling and visual  
 275 observation of potential host plants within a 1km radius area as well as 4-5 month ongoing  
 276 trapping using commercial pyramid traps, panel traps and sticky traps with MDT/murgantiol  
 277 pheromone lures for this species, resulting in apparent eradication (Horwood et al. 2019).  
 278 Similarly, trapping and control efforts for *H. halys* in Tenerife would require integrated  
 279 surveillance to confirm effectiveness and successful eradication.

280 Habitat management is often employed to manage invasive species and has been tested for  
 281 *H. halys* including via the use of attractant crops (e.g. sunflower) or management focus on  
 282 field margins (Leskey et al. 2018) but results have been mixed given that the species is so  
 283 mobile and polyphagous. However, in the drier climate and sparsely vegetated areas of  
 284 Canary Islands it might offer new opportunities that should be explored.

285 **5) Policy changes and biosecurity** Understanding the invasion pathway using surveys and  
286 genetic tracing could offer important evidence for reviewing policies related to checks of  
287 imports and ultimately better targeted biosecurity. Island areas can maximise the  
288 advantages of their isolation to prevent invasions and the Canary Islands already have  
289 separate biosecurity importation checks for agricultural products to and from mainland  
290 Europe. However, *H. halys* has been imported in “low risk” goods, such as 38 live adults  
291 detected in an Australia in November 2017 from a shipping container carrying electrical  
292 components from Italy (Horwood et al. 2019).

293 In addition, biocontrol policies might be considered as part of longer-term biosecurity. For  
294 example, although *H. halys* is yet to establish in New Zealand, authorities have pre-  
295 emptively developed a biological control program for this species, which includes risk  
296 assessment of the egg parasitoid *Trissolcus japonicus* and approval of its conditional release  
297 if *H. halys* is detected (Charles et al. 2019). In the USA, both endemic *Trissolcus* species as  
298 well as *T. japonicus* were found to parasitise *H. halys* eggs in wooded habitats, although *T.*  
299 *japonicus*, which was accidentally introduced in the USA, was more successful (Leskey and  
300 Nielsen, 2018).

301

302 **Urgent questions and factors to consider-** December and January are the coolest and  
303 wettest months in Tenerife potentially suggesting that the observed brown marmorated  
304 stink bug individuals might have been attracted to urban areas and buildings in search of  
305 overwintering areas yet that seems unlikely as daily local temperatures in late December  
306 2021 in South Tenerife were 23-25°C. In Italy, experiments indicated that overwintered  
307 female *H. halys* were all breeding and had high productivity (285 eggs/female  $\pm$  22.8 SE);  
308 only 13 of 15 summer generation females bred with lower mean lifetime productivity (214.7  
309 eggs/female  $\pm$  30.6 SE) (Costi et al. 2017). How the ecology and productivity of this species  
310 might change and adapt to the constant, subtropical dry (southern Tenerife slopes or  
311 eastern group islands) or wet and humid (northern Tenerife slopes or western islands)  
312 climate of the Canary Islands remains unknown. Understanding and answering this question  
313 is however directly relevant for modelling species demographic outputs and an estimated  
314 speed of population expansion in the absence of eradication and could ultimately improve  
315 control measures by targeting specific areas, times of the year or life stages.

316 There are also numerous questions about preferred host ornamental plants or interactions  
317 with endemic Canary plant and insect species, but these are important given the very  
318 diverse local communities compared to Europe, USA or Asia.

319 Given the disproportionate importance of banana crops for local agricultural production  
320 there are critical and urgent questions for the local agricultural economy on the potential  
321 impact, crops losses and methods for control in banana farms as to our knowledge there are  
322 currently no studies detailing *H. halys* impacts on this crop. However, the species is fed  
323 banana fruits in captive conditions (Wong et al. 2021), suggesting high potential for  
324 targeting this crop plant as a host in the Canary Islands. Quantifying the impact of *H. halys*  
325 on this major global crop would be of broad relevance for other regions in the predicted  
326 expansion range. We urge researchers to rapidly evaluate such impacts and test a variety of  
327 trapping and lethal control methods in this crop.

328 Finally, there are unknown aspects relating to optimal trapping density for both *H. halys*  
329 surveillance and trap and kill protocols but these should not prevent the application of rapid  
330 response measures as these can be managed adaptively in the field.

### 331 **Broader implications**

332 The presence of *H. halys* in Tenerife represents a dramatic expansion in its range. The  
333 potential ecological and economic impacts of the species in the Canary Islands is difficult to  
334 extrapolate from experiences elsewhere because it will be relatively novel in comparison to  
335 native insects, and species and crops in the new locality are less likely to have evolved traits  
336 to cope with this new invader (Simberloff, 2009). Ricciardi et al. (2017) identified the  
337 importance of intercontinental trade as a growing pathway for the long-distance transport  
338 of stowaways. The relatively high transport volumes of goods to nations with overseas  
339 territories, such as from Spain to the Canary Islands, combined with the extension of free  
340 trade agreements (Genovesi, 2011) suggests that such trade routes may be especially  
341 important introduction pathways that deserve tighter biosecurity regulation. By proposing a  
342 structured action plan in response to the discovery of *H. halys* in Tenerife we provide a more  
343 general framework for managing new and relatively isolated invasion events.

344

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### 352 **Availability of data and material**

353 All the data and materials are available in the manuscript. All authors have contributed and  
354 approved the manuscript. We declare no conflict of interest and no animal testing or  
355 experiments were used in our study.

356

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