



Distribution modeling of *Psellogrammus kennedyi* (Eigenmann, 1903) and new records in the Lower Paranapanema River, Brazil

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Abstract

Understanding the potential distribution of non-native species can be an important tool in preventing biological invasions. We recorded for the first time *Psellogrammus kennedyi*, a small non-native characiform, in the Lower Paranapanema River, Brazil. According to environmental variables and prediction modeling, the species presents high potential distribution in the Upper Paraná river basin. The model used herein is an efficient tool to determine where non-native species may be able to establish. This approach can be used as a preventive measure, once the control and eradication measures are often ineffective and uneconomical.

Keywords

Characidae, dispersal, distribution, freshwater, invasion, Neotropical region, South America, Upper Paraná river basin

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Introduction

The Lower Paranapanema River, a major tributary of the Upper Paraná River system, shares many fish species which naturally occurs in the neighboring Paraguay–Lower Paraná River system (Jarduli et al. 2020). Several non-native species were introduced in the Upper Paraná basin due to the removal of the natural biogeographic barrier at Sete Quedas Falls due to the construction of the Itaipu Dam (Júlio Júnior et al. 2009; Ota et al. 2018). The introductions of these species have substantially modified the Paranapanema ichthyofauna, coupled

with altered lentic conditions caused by impoundment, which suggests that the rivers' native fishes are increasingly threatened (Garcia et al. 2018a).

Recently, Jarduli et al. (2020) reviewed 90 independent studies. They compiled 225 fish species present in the Paranapanema river basin, and *Psellogrammus kennedyi* (Eigenmann, 1903) had not been previously recorded. *Psellogrammus* Eigenmann, 1908 is a monotypic genus, including only the lentic *Psellogrammus kennedyi*, which can grow to 5.9 cm long (Lima et al.

2003) and inhabits floodplains (Alves and Pompeu 2001). It was described from Campo Grande, Asuncion, Paraguay, in the Paraguay river basin (Reis et al. 2003), and its natural distribution is reportedly in the Lower Paraná–Paraguay system (Ota et al. 2018) and São Francisco river basin (Tondato et al. 2013; Fricke et al. 2021). However, the natural occurrence of *P. kennedyi* is reported to the stretch downstream of Sete Quedas Falls in the Lower Paraná River (Reis et al. 2020). The occurrence of *P. kennedyi* in the Upper Paraná river basin is a result of the elimination of this natural barrier after flooding by the Itaipu Dam; thus, this species is classified as non-native for this region (Graça and Pavanelli 2007; Júlio Júnior et al. 2009; Ota et al. 2018).

The large number of non-native fishes present in the Lower Paranapanema river basin (Garcia et al. 2018a; Jarduli et al. 2020) shows that this stretch of the river presents several optimal environmental conditions and resources for fish species in the system (Blackburn et al. 2011; Garcia et al. 2018b, 2019). Biological invasion is one of the main drivers of biodiversity loss and ecosystem damage. Invasive species are difficult to eradicate, and prevention is considered the best approach (Adelino et al. 2017). In this way, the knowledge of environmental conditions and species' distribution can be used as predictive tools for determining the establishment and invasive potential (Peterson 2003; Stohlgren et al. 2010).

Through occurrence records and environmental predictors of the native range of *P. kennedyi*, we used the species distribution models (SDM) (Franklin and Miller 2010) to generate a predictive model for the potential distribution of *P. kennedyi* in the Paraná river basin (Peterson 2003; Pereira et al. 2020). Models predicted high suitability for *P. kennedyi* in the Upper Paraná river basin. New records corroborate the prediction of the potential distribution in the Lower Paranapanema river basin, indicating that the basin presents the environmental conditions for this species to become well established. In addition, we discuss how the construction of dams and the biological aspects of *P. kennedyi* may have facilitated this invasive process.

Methods

Sampling was performed between April 2019 and June 2020 during a fish fauna monitoring project in the Paranapanema river basin (Project #11218/2018). Specimens were collected in the Rosana Reservoir at site 1 (22°32'06"S, 052°01'55"W) and in the Taquaruçu Reservoir at site 2 (22°39'12"S, 051°37'36"W), site 3 (22°38'48"S, 051°26'45"W), and site 4 (22°40'07"S, 051°24'10"W), both in the Lower Paranapanema River, which is the portion of the basin that begins after Salto Grande Falls (currently flooded by the Salto Grande Reservoir; Fig. 1). The samples were collected using standard ichthyological gear comprising sieve and seine nets of 2 mm mesh size. Fish were euthanized by overexposing to 1 g/ml clove oil, fixed with 10% formalin, and transferred to

70% alcohol. The Animal Ethics Committee authorized the field sampling (Comissão de Ética no uso de Animais, CEUA no. 24310.2017.78; collection license no. 16578). All specimens are deposited at the Museu de Zoologia da Universidade Estadual de Londrina (MZUEL). We confirmed identification based on the original description (Eigenmann and Kennedy 1903) and the description by Ota et al. (2018); we also consulted expert assistance. Measurements and counts were performed on 125 specimens following Fink and Weitzman (1974), using digital calipers, point-to-point, on the left side of the specimens whenever possible, and with a precision of 0.1 mm (SL = standard length).

Occurrence records of *P. kennedyi* were obtained from online databases of species in zoological collections by searching the Centro de Referência em Informação Ambiental (272 records) (SpeciesLink; CRIA 2021) and the Global Biodiversity Information Facility (126 records) (GBIF 2021). Only georeferenced records containing voucher specimens were considered valid for modeling purposes. For modeling the environmental niche in the native geographic range of *P. kennedyi* (Jiménez-Valverde et al. 2011), we used only native species occurrence data (261 records) (i.e., Paraguay basin) (Ota et al. 2018; Fricke et al. 2021) (Figs. 1, 3).

Bioclimatic variables are derived from the monthly temperature and rainfall values to generate more biologically meaningful variables. Temperature is responsible for altering the metabolism (e.g., enzymatic activity) of living organisms and precipitation for the seasonal variations of droughts and floods, synchronizing biological events of fish, such as migration, spawning, home range, and growth (Lopes et al. 2017; Ruaro et al. 2019). We used six bioclimatic variables related to the environmental tolerance of *P. kennedyi* to temperature and precipitation. Their respective codes were: BIO1 = annual mean temperature, BIO5 = maximum temperature of warmest month, BIO6 = minimum temperature of coldest month, BIO12 = annual precipitation, BIO13 = precipitation of wettest month, and BIO14 = precipitation of driest month. Bioclimatic variables were obtained from the WorldClim (<http://www.worldclim.org/>) v. 2.0 (Fick and Hijmans 2017). We used a principal component analysis (PCA) to remove redundancy among environmental variables to produce uncorrelated variables (Peterson et al. 2011). We chose to use the MaxEnt (maximum entropy; Phillips et al. 2006) model algorithm, as this method has a high capacity for predictive accuracy of modeling species–environment relationships using presence-only data (Franklin and Miller 2010).

We obtained an initial set of 100 models for the species, selected at random 75% of the occurrence localities at each run for training, and left the remaining 25% for testing models. The area under the ROC curve (AUC) was calculated to validate the quality of the final models generated (Phillips et al. 2006), as well as a true skill statistic (TSS) analysis which compares the number of correct forecasts, minus those attributable to random

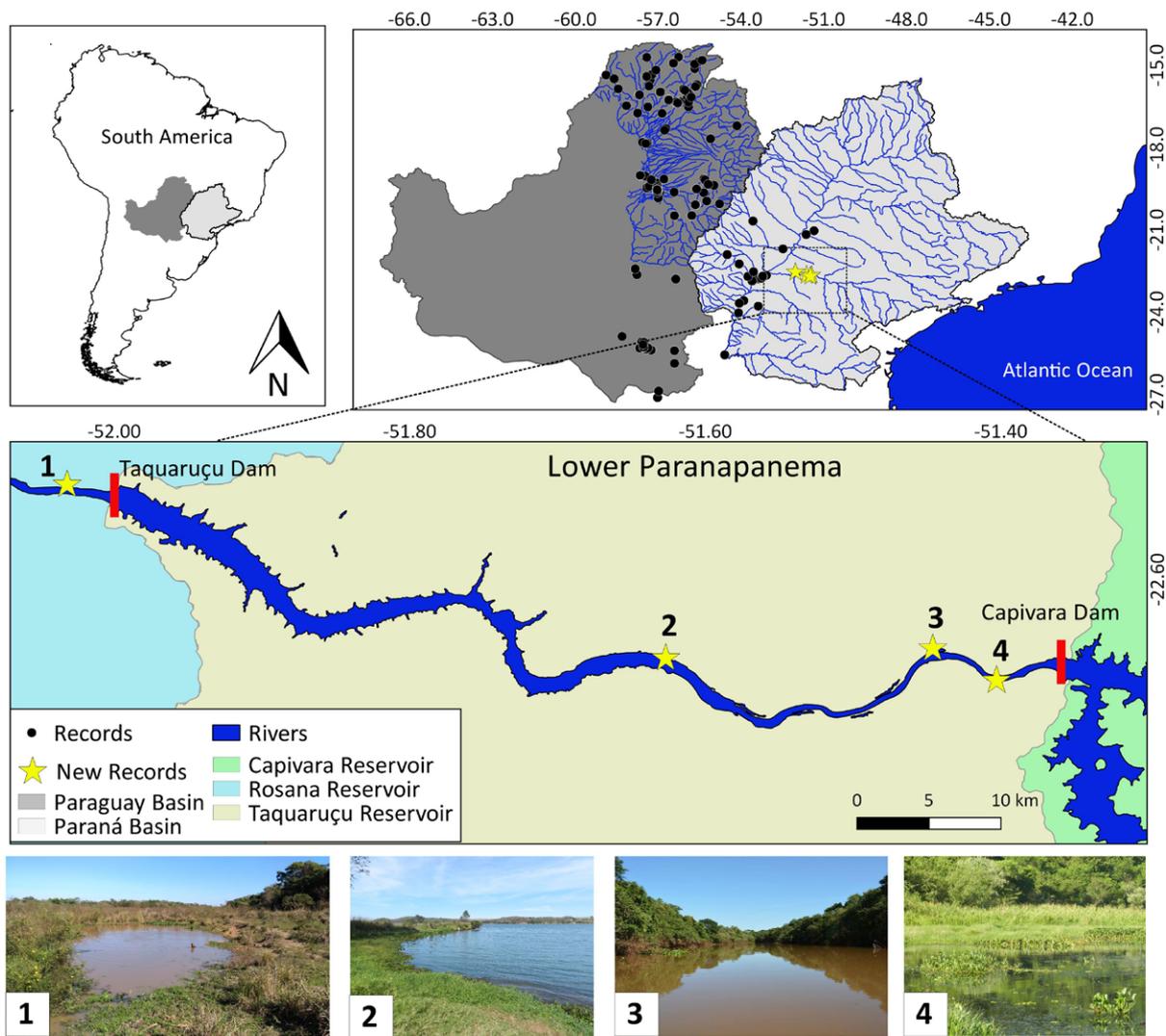


Figure 1. Distribution of *Psellogrammus kennedyi*. Yellow stars represent new records from the Paranapanema river basin. Black circles represent records from the Paraná and Paraguay river basins.

guessing, to that of a hypothetical set of perfect forecasts (Allouche et al. 2006). Models with AUC = 0.7 and TSS = 0.4 values were selected to construct the final model (Buisson et al. 2010). Maxent jackknife test of variable importance was used to evaluate the relative contribution of each predictor variable to construct the models (Yost et al. 2008). All analyses were conducted using the R software v. 3.6.3 (R Core Team 2020; R Studio Team 2020) dismo package (Hijmans et al. 2021).

Results

Psellogrammus kennedyi (Eigenmann 1903)

Figures 1–3; Tables 1, 2

New records. BRAZIL – **Paraná** • Paranapanema river basin, Taquaruçu Reservoir, Lupionópolis municipality; 22°39'12"S, 051°37'36"W; 12.IV.2019; Laboratório de Ecologia de Peixes e Invasões Biológicas LEPiB leg.; 1, sex undet., 33.9 mm SL, MZUEL 20869 • Porecatu municipality; 22°40'07"S, 051°24'10"W; 06.VI.2019; LEPiB leg.; 2, sex undet., 34.7 and 41.8 mm SL, MZUEL

20870 – **São Paulo** • Paranapanema river basin, Taquaruçu Reservoir, Nandiba municipality, near Anhumas River; 22°38'48"S, 051°26'45"W; 06.VI.2019; LEPiB leg.; 2, sex undet., 29.7 and 39.6 mm SL, MZUEL 20872 • Rosana Reservoir, Teodoro Sampaio municipality; 22°32'6"S, 052°01'55"W; 18.VI.2020; LEPiB leg.; marginal lagoon; 120, sex undet., 28.3–43.8 mm SL, MZUEL 20871.

Identification. Body deep and compressed; greatest depth contained 2.1–2.5× and caudal peduncle depth 9.7–10.8× in standard length; head length 3.4–4.0×, predorsal distance 1.9–2.1×; snout length 3.5–4.0×, horizontal orbital diameter 2.3–2.8×, and least interorbital width 2.6–3.2× in head length. Mouth terminal; the inner row of premaxilla with 5 teeth, outer row with 3–5 teeth, dentary row with 8–11 teeth, and maxilla with 1 tooth. Lateral line irregular (incomplete or rarely complete); longitudinal series with 40–45 scales. Dorsal fin with 10 rays, pectoral fin with 12 or 13 rays, pelvic fin with 8 rays, anal fin with 39–46 rays, and caudal fin with 19 rays. Ground color whitish; with a diffuse, black humeral spot; with a dark-brown, rounded blotch on distal portion



Figure 2. *Psellogrammus kennedyi*. MZUEL 20872, 31.5 mm SL, Lower Paranapanema River, Upper Paraná basin, Paraná, Brazil. Scale bar = 1 cm.

of caudal peduncle and caudal-fin base. Fins hyaline (Graça and Pavanelli 2007; Ota et al. 2018). Other measurements are presented in Table 1.

Psellogrammus kennedyi is associated with marginal lagoons and macrophyte banks in the littoral zone of the reservoir in our new records. The model obtained for the present distribution of *P. kennedyi* in the Paraná basin (Fig. 2) performed well, with an average AUC value of 0.74 (\pm 0.02) and TSS 0.834 (\pm 0.064), the algorithm converged after 280 iterations. The model shows greater suitability for *P. kennedyi* in the Upper Paraná river basin areas, like the lower stretches of Paranaíba, Grande, Tietê, and Paranapanema rivers (Fig. 2). The minimum temperature of coldest month (BIO6) presented the greatest contribution to the model construction, followed by the precipitation of driest month (BIO14) and annual mean temperature (BIO1); tree variables combined sum up to a total of 95.6% of the contribution for the model construction (Table 2). The Jackknife analysis also shows the variable BIO1 as the one that presents a greater gain to the training of the model.

Discussion

Ecological niche modeling is an approach widely used to predict species distribution (Guisan and Zimmermann 2000; Franklin and Miller 2010). It has been an important tool in several research lines, such as biodiversity mapping, conservation planning, niche evolution, climate change impacts on species, and biological invasions (Peterson et al. 2008; Kriticos and Leriche 2010; Ruaro et al. 2019). Based on environmental predictors through niche modeling (Guisan and Thuiller 2005; Araújo and Guisan 2006; Pereira et al. 2020), we used data on the natural occurrence of *P. kennedyi* to understand the potential distribution and occurrence of the species outside its

native range. New records of *P. kennedyi* in the Upper Paraná River in recent years are evidence of an expansion of the non-native distribution of this species (Fiori et al. 2016; Peláez et al. 2017; da Costa-Silva et al. 2018; Vicentin et al. 2019).

The suggested suitability model predicted that the Upper Paraná river basin has good to high potential for the increased distribution and establishment of *P.*

Table 1. Morphometric data of *Psellogrammus kennedyi* specimens collected in the Lower Paranapanema river basin, Brazil ($N = 125$). SD = standard deviation of averages.

Measurements	Range	Mean	SD
Standard length (mm)	28.3–43.8	36.1	—
Percents of standard length			
Body depth	11.2–19.3	15.3	2.9
Body width	2.8–5.0	3.9	0.8
Head length	7.0–10.9	9.0	1.4
Head depth	5.8–9.0	7.4	1.1
Predorsal length	14.6–21.1	17.9	2.3
Prepelvic length	10.9–17.8	14.4	2.4
Preanal length	16.3–23.6	20.0	2.6
Caudal peduncle depth	2.3–3.9	3.1	0.6
Dorsal-fin base length	3.7–5.6	4.7	0.7
Anal-fin base length	12.3–19.2	15.8	2.4
Pectoral-fin length	6.2–10.4	8.3	1.5
Pelvic-fin length	4.1–6.1	5.1	0.7
Dorsal-fin length	7.9–13.7	10.8	2.1
Anal-fin length	5.6–7.9	6.8	0.8
Caudal peduncle length	2.6–5.5	4.1	1.0
Dorsal-fin to adipose-fin distance	5.9–11.8	8.9	2.1
Eye to dorsal-fin origin	11.2–16.6	13.9	1.9
Dorsal origin to caudal origin	14.4–24.5	19.5	3.6
Percentages of head length			
Interorbital width	2.6–3.9	3.3	0.5
Snout length	2.9–3.9	3.4	0.4
Orbital diameter	2.7–3.7	3.2	0.4
Upper jaw length	2.0–2.8	2.4	0.3

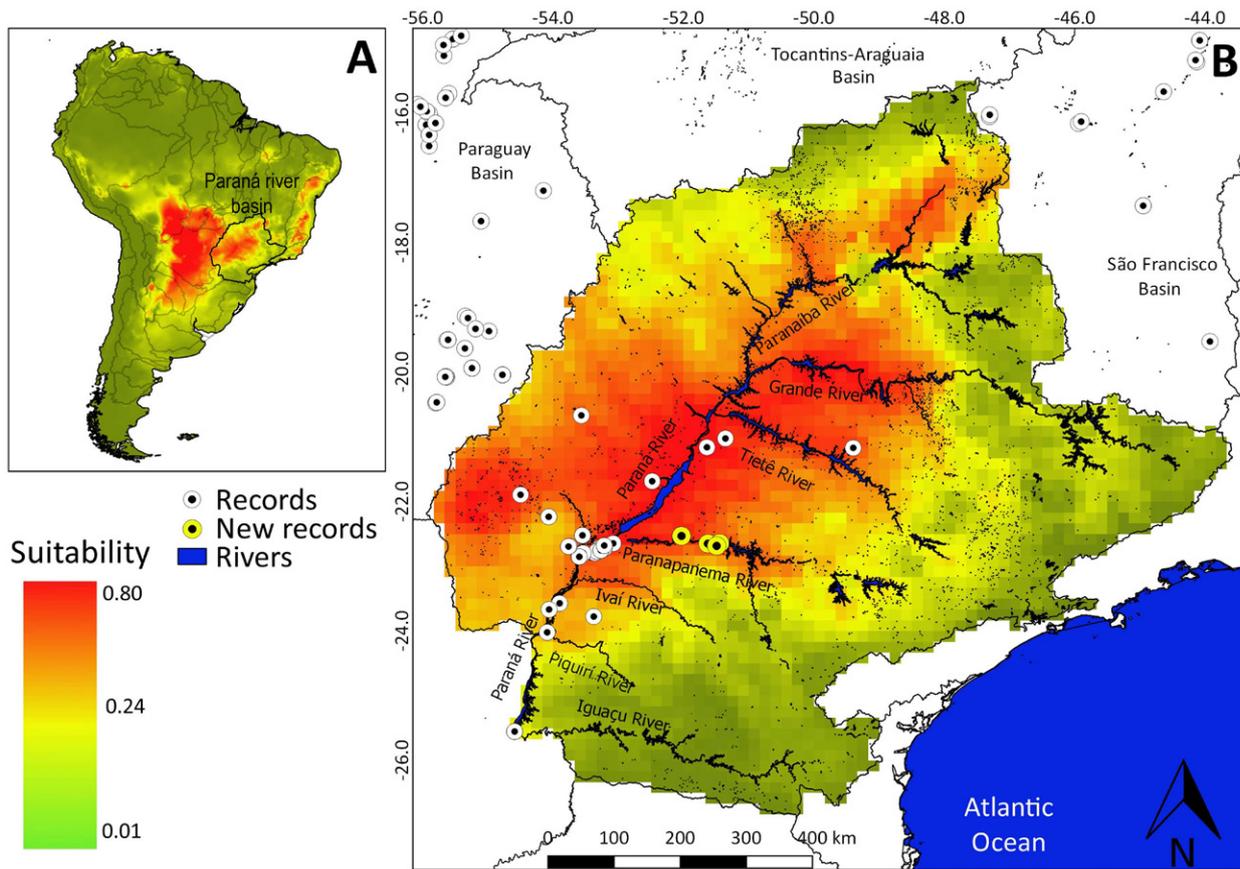


Figure 3. Suitability map for *Psellogrammus kennedyi*. **A.** South America. **B.** Paraná river basin.

Table 2. Estimates of average contribution and permutation importance of the environmental variables used in the MaxEnt modeling algorithm for *Psellogrammus kennedyi*.

Variable	Code	Percentage contribution	Permutation importance
Min. temperature of coldest month	BI06	62.2	35.5
Precipitation of driest month	BI014	23.0	13.6
Annual mean temperature	BI01	10.4	45.9
Precipitation of wettest month	BI013	2.3	0.8
Max. temperature of warmest month	BI05	1.2	1.5
Annual precipitation	BI012	0.9	2.6

Threshold kappa: 0.58524682
Sensitivity: 0.58944197

kennedyi (> 0.50) (i.e., between 15°S and 24°S in Brazil). Our model identified that some environmental variables, such as variations in temperature and the annual mean precipitation, explain the current range of *P. kennedyi*. We found that the minimum temperature of the coldest month, precipitation of the driest month, and annual mean temperature are important climatic contributors in predicting suitable areas for *P. kennedyi*. This species is favored in regions where climate is less variable, like in the Paraguay basin (native region), and also in nearby areas at similar latitudes such as the Upper Paraná river basin (non-native region), where the average annual temperatures vary between 22.5 °C and 26.5 °C and the average annual precipitation in the basin is 1,396 mm, ranging from 800 to 1,600 mm (Gonçalves et al. 2011).

The presence of *P. kennedyi* in the Upper Paraná region is due to the Canal da Piracema, a fish transposition system that connects the region downstream Itaipu Dam to the region upstream (Júlio Júnior et al. 2009; Ota et al. 2018). After the flooding of the reservoir behind the Itaipu Dam, the loss of the Sete Quedas Falls allowed for hydrologic connectivity between the Lower and Upper Paraná River. Many fishes of the lower region of the Paraná River were able to colonize the upper stretches (Vitule et al. 2012), including the Paranapanema River (Garcia et al. 2018a). The damming of stretches of the Lower Paranapanema River for hydroelectric power seems to have favored the establishment of *P. kennedyi* in reservoirs, as reported by Garcia et al. (2018b) for other non-native fish species in the Paranapanema River. Artificial reservoirs are globally known to be hotspots for invaders (Johnson et al. 2008). The colonization of reservoirs by small species is well-documented in the literature, and species that habitually associate with macrophytes in the littoral zone are favored (Casatti et al. 2003; Pelicice and Agostinho 2009; Agostinho et al. 2016). Fiori et al. (2016) found that *P. kennedyi* has physiological adaptations that allow it to obtain the maximum energy from food, even if of low nutritional value, or that this species can compensate for low-nutrition foods by increasing consumption. Thus, *P. kennedyi* has great food plasticity allowing diet changes according to the environment, thus favoring its dispersion and invasion.

Species with generalist reproductive and trophic

niche strategies are more likely to have invasion success (Agostinho et al. 2007, 2016; Garcia et al. 2018a, 2018b, 2019). Galvão et al. (2016) observed that *P. kennedyi* might be sexually mature throughout the year, with both females and males with three stages of gonadal maturation. Thus, *P. kennedyi* has asynchronous development of oocytes and mature males throughout the year. This allows males to always reproduce, making the release of sperm possible in the most favorable conditions. This reproductive strategy can assist in the colonization of new areas, where different niches in temporal and spatial scales are occupied with different size classes (de Carvalho et al. 2009; Agostinho et al. 2016; Araújo et al. 2019). In addition, *P. kennedyi* has a laterally depressed body, and its body is taller than several small characiforms which have a fusiform body. This allows *P. kennedyi* to establish in lentic environments such as marginal zones of reservoirs and marginal lagoons (Breda et al. 2005).

Biological invasions are a serious threat to the Upper Paraná and Paranapanema river basins (Langeani et al. 2007; Júlio Júnior et al. 2009; Pelicice and Agostinho 2009; Garcia et al. 2018a), and understanding the biology, ecology, and potential distribution of non-native species can serve as a basis for preventive measures (Simberloff 2003; Broennimann and Guisan 2008). Our results indicate that distribution modeling should be used to understand potential dispersion, establishment, and invasion of non-native species. This knowledge should be used preventively, since control and eradication measures are often ineffective and with high costs.

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References

- Adelino JRP, Anjos L dos, Lima MR (2017) Invasive potential of the pied crow (*Corvus albus*) in eastern Brazil: best to eradicate before it spreads. *Perspectives in Ecology and Conservation* 15 (3): 227–233. <https://doi.org/10.1016/j.pecon.2017.07.001>
- Agostinho AA, Gomes LC, Pelicice FM (2007) Ecologia e manejo de recursos pesqueiros em reservatórios de Brasil. EDUEM, Maringá, Brazil, 502 pp.
- Agostinho AA, Gomes LC, Santos NCL, Ortega JCG, Pelicice FM (2016) Fish assemblages in Neotropical reservoirs: colonization patterns, impacts and management. *Fisheries Research* 173 (1): 26–36. <https://doi.org/10.1016/j.fishres.2015.04.006>
- Allouche O, Tsoar A, Kadmon R (2006) Assessing the accuracy of species distribution models: prevalence, kappa and the true skill statistic (TSS). *Journal of Applied Ecology* 43 (6): 1223–1232. <https://doi.org/10.1111/j.1365-2664.2006.01214.x>
- Alves CBM, Pompeu PS (2001) Peixes do Rio das Velhas: passado e presente. SEGRAC, Belo Horizonte, Brazil, 194 pp.
- Araújo FG, Nascimento AA, Gomes ID, Sales A, de Oliveira BAC (2019) Gonadal development and reproductive period of the characin *Astyanax* aff. *bimaculatus* (Characiformes: Characidae) in a tropical reservoir in southeastern Brazil. *Zoologia* 36: 1–14. <https://doi.org/10.3897/zoologia.36.e30610>
- Araújo MB, Guisan A (2006) Five (or so) challenges for species distribution modelling. *Journal of Biogeography* 33 (10): 1677–1688. <https://doi.org/10.1111/j.1365-2699.2006.01584.x>
- Blackburn TM, Pyšek P, Bacher S, Carlton JT, Duncan RP, Jarošík V, Wilson JR, Richardson DM (2011) A proposed unified framework for biological invasions. *Trends in Ecology & Evolution* 26 (7): 333–339. <https://doi.org/10.1016/j.tree.2011.03.023>
- Breda L, Oliveira EF, Goulart E (2005) Ecomorfologia de locomoção de peixes com enfoque para espécies neotropicais. *Acta Scientiarum. Biological Sciences* 27 (4): 371–381. <https://doi.org/10.4025/actascibiolsci.v27i4.1271>
- Broennimann O, Guisan A (2008) Predicting current and future biological invasions: both native and invaded ranges matter. *Biology Letters* 4 (5): 585–589. <https://doi.org/10.1098/rsbl.2008.0254>
- Buisson L, Thuiller W, Casajus N, Lek S, Grenouillet G (2010) Uncertainty in ensemble forecasting of species distribution. *Global Change Biology* 16 (4): 1145–1157. <https://doi.org/10.1111/j.1365-2486.2009.02000.x>
- de Carvalho PA, Paschoalini AL, Santos GB, Rizzo E, Bazzoli N (2009) Reproductive biology of *Astyanax fasciatus* (Pisces: Characiformes) in a reservoir in southeastern Brazil. *Journal of Applied Ichthyology* 25 (3): 306–313. <https://doi.org/10.1111/j.1439-0426.2009.01238.x>
- Casatti L, Mendes HF, Ferreira KM (2003) Aquatic macrophytes as feeding site for small fishes in the Rosana Reservoir, Paranapanema River, southeastern Brazil. *Brazilian Journal of Biology* 63 (2): 213–222. <https://doi.org/10.1590/S1519-69842003000200006>
- da Costa-Silva GJ, Yuldi Ashikaga F, Kioko Shimabukuro Dias C, Garcia Pereira LH, Foresti F, Oliveira C (2018) DNA barcoding techniques used to identify the shared ichthyofauna between the Pantanal floodplain and Upper Parana River. *Mitochondrial DNA Part A* 29 (7): 1063–1072. <https://doi.org/10.1080/24701394.2017.1404046>
- CRIA (Centro de Referência em Informação Ambiental) (2005) SpeciesLink network. CRIA, Campinas, Brazil. <http://www.splink.org.br/>. Accessed on: 2021-03-27.
- Eigenmann CH., Kennedy CH (1903) On a collection of fishes from Paraguay, with a synopsis of the American genera of cichlids. *Proceedings of the Academy of Natural Sciences of Philadelphia* 55: 497–537.
- Fick SE, Hijmans RJ (2017) WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology* 37 (12): 4302–4315. <https://doi.org/10.1002/joc.5086>
- Fink WL, Weitzman SH (1974) The so-called cheirodontin fishes of Central America with descriptions of two new species (Pisces: Characidae). *Smithsonian Contributions to Zoology*: 1–46. <https://doi.org/10.5479/si.00810282.172>
- Fiori LF, Alves GHZ, Hahn NS, Benedito E (2016) Influence of feeding plasticity on the fitness of small Neotropical characids. *Iheringia, Série Zoologia* 106. <https://doi.org/10.1590/1678-4766e2016006>
- Franklin J, Miller JA (2010) Mapping species distributions: spatial inference and prediction. Cambridge University, Cambridge, UK, 320 pp.
- Fricke R, Eschmeyer WN, Van der Laan R (2021) Eschmeyer's catalog of fishes: genera, species, references. California Academy of Sciences, San Francisco, USA. <http://researcharchive.calacademy.org/research/ichthyology/catalog/fishcatmain.asp>. Accessed on:

- 2021-03-21.
- Galvão GA, Silva ALB, Cardoso AS, Santos HS, Pereira PAN, Ribeiro LB (2016) Histomorfometria gonadal comparada de *Astyanax lacustris* (Lütken, 1875) e *Psellogrammus kennedyi* (Eigenmann, 1903) (Characiformes, Characidae) em um reservatório no semiárido brasileiro. *Boletim do Instituto de Pesca* 42 (4): 734–748. <https://doi.org/10.20950/1678-2305.2016v42n4p734>
- Garcia DAZ, Britton JR, Vidotto-Magnoni AP, Orsi ML (2018a) Introductions of non-native fishes into a heavily modified river: rates, patterns and management issues in the Paranapanema River (Upper Paraná ecoregion, Brazil). *Biological Invasions* 20 (5): 1229–1241. <https://doi.org/10.1007/s10530-017-1623-x>
- Garcia DAZ, Vidotto-Magnoni AP, Orsi ML (2018b) Diet and feeding ecology of non-native fishes in lentic and lotic freshwater habitats. *Aquatic Invasions* 13 (4): 565–573. <https://doi.org/10.3391/ai.2018.13.4.13>
- Garcia DAZ, Vidotto-Magnoni AP, Orsi ML (2019) Características reprodutivas de peixes invasores no rio Paranapanema, bacia do alto rio Paraná, sul do Brasil. *Neotropical Biology and Conservation* 14 (4): 511–528. <https://doi.org/10.3897/neotropical.14.e49079>
- GBIF (Global Biodiversity Information Facility) (2021) Natural History Museum of Denmark, Copenhagen, Denmark. <https://www.gbif.org/>. Accessed on: 2021-03-27. <https://doi.org/10.15468/dl.fs.ewc3>
- Gonçalves H, Mercante M, Santos E (2011) Hydrological cycle. *Brazilian Journal of Biology* 71 (1): 241–253. <https://doi.org/10.1590/S1519-69842011000200003>
- Graça WJ Da, Pavanelli CS (2007) Peixes da planície de inundação do alto rio Paraná e áreas adjacentes. EDUEM, Maringá, Brazil, 241 pp.
- Guisan A, Zimmermann NE (2000) Predictive habitat distribution models in ecology. *Ecological Modelling* 135 (2–3): 147–186. [https://doi.org/10.1016/S0304-3800\(00\)00354-9](https://doi.org/10.1016/S0304-3800(00)00354-9)
- Guisan A, Thuiller W (2005) Predicting species distribution: offering more than simple habitat models. *Ecology Letters* 8 (9): 993–1009. <https://doi.org/10.1111/j.1461-0248.2005.00792.x>
- Hijmans RJ, Phillips S, Leathwick J, Elith J (2021) dismo: species distribution modeling. R package version 1.3-3. <https://cran.r-project.org/web/packages/dismo/index.html>. Accessed on: 2021-03-27.
- Jiménez-Valverde A, Peterson AT, Soberón J, Overton JM, Aragón P, Lobo JM (2011) Use of niche models in invasive species risk assessments. *Biological Invasions* 13 (12): 2785–2797. <https://doi.org/10.1007/s10530-011-9963-4>
- Jarduli LR, Garcia DAZ, Vidotto-Magnoni AP, Casimiro ACR, Vianna NC, Almeida FS, Jerpe FC, Orsi ML (2020) Fish fauna from the Paranapanema river basin, Brazil. *Biota Neotropica* 20 (1): 1–19. <https://doi.org/10.1590/1676-0611-bn-2018-0707>
- Johnson PT, Olden JD, Vander Zanden MJ (2008) Dam invaders: impoundments facilitate biological invasions into freshwaters. *Frontiers in Ecology and the Environment* 6 (7): 357–363. <https://doi.org/10.1890/070156>
- Júlio Júnior HF, Tós CD, Agostinho AA, Pavanelli CS (2009) A massive invasion of fish species after eliminating a natural barrier in the upper rio Paraná basin. *Neotropical Ichthyology* 7 (4): 709–718. <https://doi.org/10.1590/S1679-62252009000400021>
- Kriticos DJ, Leriche A (2010) The effects of climate data precision on fitting and projecting species niche models. *Ecography* 33 (1): 115–127. <https://doi.org/10.1111/j.1600-0587.2009.06042.x>
- Langeani F, Castro RMC, Oyakawa OT, Shibatta OA, Pavanelli CS, Casatti L (2007) Diversidade da ictiofauna do Alto Rio Paraná: composição atual e perspectivas futuras. *Biota Neotropica* 7 (3): 181–197. <https://doi.org/10.1590/S1676-06032007000300020>
- Lima FCT, Malabarba LR, Buckup PA, Silva JFP, Vari RP, Benine R, Oyakawa OT, Pavanelli CS, Menezes NA, Lucena CAS, Malabarba MCSL, Lucena ZMS, Reis RE, Langeani F, Cassati L, Bertaco VA, Moreira C, Lucinda PHF (2003) Genera incertae sedis in Characidae. In: Reis RE, Kullander SO Jr. (Eds.) Check list of the freshwater fishes of South and Central America. EDIPUCRS, Porto Alegre, Brazil, 106–169.
- Lopes TM, Bailly D, Almeida BA, Santos NCL, Gimenez BCG, Landgraf GO, Sales PCL, Lima-Ribeiro MS, Cassemiro FAS, Rangel TF, Diniz-Filho JAF, Agostinho AA, Gomes LC (2017) Two sides of a coin: effects of climate change on the native and non-native distribution of *Colossoma macropomum* in South America. *PLOS ONE* 12(6): e0179684. <https://doi.org/10.1371/journal.pone.0179684>
- Ota RR, Deprá GC, Graça WJ, Pavanelli CS (2018) Peixes da planície de inundação do alto rio Paraná e áreas adjacentes: revised, annotated and updated. *Neotropical Ichthyology* 16 (2): e170094. <https://doi.org/10.1590/1982-0224-20170094>
- Peláez OE, Azevedo FM, Pavanelli CS (2017) Environmental heterogeneity explains species turnover but not nestedness in fish assemblages of a Neotropical basin. *Acta Limnologica Brasiliensia* 29: e117. <https://doi.org/10.1590/s2179-975x8616>
- Pelicice FM, Agostinho AA (2009) Fish fauna destruction after the introduction of a non-native predator (*Cichla kelberi*) in a Neotropical reservoir. *Biological Invasions* 11 (8): 1789–1801. <https://doi.org/10.1007/s10530-008-9358-3>
- Pereira AD, Pires Adelino JR, Garcia DAZ, Rodrigues Casimiro AC, Vizintim Marques AC, Vidotto-Magnoni AP, Bazilio S, Orsi ML (2020) Modeling the geographic distribution of *Myocastor coypus* (Mammalia, Rodentia) in Brazil: establishing priority areas for monitoring and an alert about the risk of invasion. *Studies on Neotropical Fauna and Environment* 55 (2): 139–148. <https://doi.org/10.1080/01650521.2019.1707419>
- Peterson AT (2003) Predicting the geography of species' invasions via ecological niche modeling. *The Quarterly Review of Biology* 78 (4): 419–433. <https://doi.org/10.1086/378926>
- Peterson AT, Papeş M, Soberón J (2008) Rethinking receiver operating characteristic analysis applications in ecological niche modeling. *Ecological Modelling* 213 (1): 63–72. <https://doi.org/10.1016/j.ecolmodel.2007.11.008>
- Peterson AT, Soberón J, Pearson RG, Anderson RP, Martínez-Meyer E, Nakamura M, Araújo MB (2011) Ecological niches and geographic distributions. Monographs (MPB-49). Princeton University Press, Princeton, USA, 328 pp.
- Phillips SJ, Anderson RP, Schapire RE (2006) Maximum entropy modeling of species geographic distributions. *Ecological Modelling* 190 (3–4): 231–259. <https://doi.org/10.1016/j.ecolmodel.2005.03.026>
- R Core Team (2020) R: a language and environment for statistical computing (version 3.6.3). R Foundation for Statistical Computing, Vienna, Austria. <https://www.R-project.org/>. Accessed on: 2021-03-29.
- R Studio Team. 2020. RStudio: integrated development environment for R. RStudio, Inc., Boston, MA. <https://rstudio.com>. Accessed on: 2021-03-29.
- Reis RE, Kullander SO, Ferraris Jr CJ (Eds) (2003) Check List of the freshwater fishes of South and Central America. Edipucrs, Porto Alegre, Brazil, 742 pp.
- Reis RB dos, Frota A, Deprá G de C, Ota RR, Graça WJ da (2020) Freshwater fishes from Paraná state, Brazil: an annotated list, with comments on biogeographic patterns, threats, and future perspectives. *Zootaxa* 4868 (4): 451–494. <https://doi.org/10.11646/zootaxa.4868.4.1>
- Ruaro R, Conceição EO, Silva JC, Cafoto EG, Angulo-Valencia MA, Mantovano T, Pineda A, de Paula ACM, Zanco BF, Capparro EM, Moresco GA, de Oliveira IJ, Antikeira JL, Ernandes-Silva J, da Silva JVF, Adelino JRP, dos Santos JA, Ganassin MJM, Iquematsu MS, Landgraf GO, Lemes P, Cassemiro FAS, Batista-Silva VF, Diniz-Filho JAF, Rangel TF, Agostinho AA, Bailly D (2019) Climate change will decrease the range of a keystone fish species in La Plata river basin, South America. *Hydrobiologia* 836 (1): 1–19. <https://doi.org/10.1007/s10750-019-3904-0>
- Simberloff D (2003) Eradication: preventing invasions at the outset. *Weed Science* 51 (2): 247–253.

- Stohlgren TJ, Ma P, Kumar S, Rocca M, Morisette JT, Jarnevich CS, Benson N (2010) Ensemble habitat mapping of invasive plant species. *Risk Analysis* 30 (2): 224–235. <https://doi.org/10.1111/j.1539-6924.2009.01343.x>
- Tondato KK, Fantin-cruz I, Pedrollo OC, Suárez YR (2013) Spatial distribution of fish assemblages along environmental gradients in the temporary ponds of northern Pantanal, Brazil. *Journal of Limnology* 72 (1): 95–102. <https://doi.org/10.4081/jlimnol.2013.e8>
- Vicentín W, Ferreira FS, Suárez YR (2019) Ichthyofauna of lotic environments in the Ivinhema river basin, Upper Paraná River, Mato Grosso do Sul state, Brazil. *Biota Neotropica* 19 (4): e20190735. <https://doi.org/10.1590/1676-0611-bn-2019-0735>
- Vitule JRS, Skóra F, Abilhoa V (2012) Homogenization of freshwater fish faunas after the elimination of a natural barrier by a dam in Neotropics. *Diversity and Distributions* 18 (2): 111–120. <https://doi.org/10.1111/j.1472-4642.2011.00821.x>
- Yost AC, Petersen SL, Gregg M, Miller R (2008) Predictive modeling and mapping sage grouse (*Centrocercus urophasianus*) nesting habitat using maximum entropy and a long-term dataset from Southern Oregon. *Ecological Informatics* 3 (6): 375–386. <https://doi.org/10.1016/j.ecoinf.2008.08.004>